

1 Submission: Original article to **Ecology**

2 **Title: Propagules are not all equal: traits of vegetative fragments and disturbance**
3 **regulate invasion success.**

4 Running title: Propagules, disturbance and establishment.

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22 **ABSTRACT**

23 Invasion success is regulated by multiple factors. While the roles of disturbance and
24 propagule pressure in regulating the establishment of non-native species are widely
25 acknowledged, that of propagule morphology (a proxy for quality) is poorly known. By
26 means of a multi-factorial field experiment, we tested how the number (5 versus 10) and
27 quality (intact, without fronds or without rhizoids) of fragments of the clonal invasive
28 seaweed, *Caulerpa cylindracea*, influenced its ability to establish in patches of the native
29 seagrass, *Posidonia oceanica*, exposed to different intensities of disturbance (0, 50 or 100 %
30 reduction in canopy cover). We hypothesized that the ability of fragments to establish would
31 be greater for intact fragments (high quality) and reduced more by frond removal (low
32 quality) than rhizoid removal (intermediate quality). At low propagule pressure or quality,
33 fragment establishment was predicted to increase with increasing disturbance, whereas, at
34 high propagule pressure or quality, it was predicted to be high regardless of disturbance
35 intensity. Disturbance intensity, fragment number and quality had independent effects on *C.*
36 *cylindracea* establishment success. Disturbance always facilitated fragment establishment.
37 However, fragments retaining fronds, either intact or deprived of rhizoids, had higher
38 establishment success than fragments deprived of fronds. Increasing propagule number had
39 weak effects on the cover of *C. cylindracea*. Our results demonstrate that propagule traits
40 enabling the acquisition of resources made available by disturbance can be more important
41 than propagule number in determining the establishment and spread of clonal non-native
42 plants. More generally, our study suggests that propagule quality is a key, yet underexplored,
43 determinant of invasion success.

44 *Keywords:* biological invasion; biotic resistance; disturbance; propagule pressure; propagule
45 quality; clonal seaweeds; *Caulerpa cylindracea*

46 1. INTRODUCTION

47 Understanding the mechanisms that underlie biological invasions, one of the major causes
48 of biodiversity loss worldwide (Mack et al. 2000, Dextrase and Mandrak 2006), is key to
49 predicting habitats vulnerable to future invasions. Many factors, such as features of the
50 recipient native community, disturbance regimes, the biology and ecology of the invasive
51 species and its associated propagule pressure, can regulate the success of invasive species
52 (D'Antonio and Vitousek 1992, Davis et al. 2000, Stachowicz et al. 2002, Lockwood et al.
53 2005). However, manipulative experimental studies have generally focused on only one of
54 these factors, despite compelling evidence indicating that invasion success is likely dependent
55 on interactions among them (Leung and Mandrak 2007, Britton-Simmons and Abbott 2008,
56 Clark and Johnston 2009, Eschtruth and Battles 2009).

57 Disturbance to native communities is often a key determinant of invasion success (Davis
58 et al. 2000). For example, disturbance to native canopies can release non-native plants from
59 biotic resistance by providing access to limited resources such as light and space (Elton 1958,
60 Hobbs and Huenneke 1992, Stachowicz et al. 2002, Corbin and D'Antonio 2004, Clark and
61 Johnston 2005, Bulleri et al. 2010, Byun et al. 2017). However, the effects of increased
62 resource availability (Davis et al. 2000) can interact with invader propagule pressure to
63 determine invasion success (D'Antonio et al. 2001, Thomsen et al. 2006, Britton-Simmons
64 and Abbott 2008, Byun et al. 2015). For instance, disturbance to native fouling communities
65 fostered the establishment of the invasive bryozoan *Watersipora suborquata* only when it
66 occurred in association with high propagule pressure (Clark and Johnston 2009). Likewise, in
67 a Californian coastal grassland, decreased biotic resistance due to higher springtime water
68 availability amplified the positive effects of increased propagule pressure on the
69 establishment of the European perennial grass *Holcus lanatus* (Thomsen et al. 2006). On the
70 other hand, resistance to invasion can be overwhelmed by high invader propagule pressure,

71 irrespective of disturbance levels (Hollebone and Hay 2007, 2008). For example, high
72 propagule pressure of the Japanese stiltgrass, *Microstegium vimineum*, overwhelmed the
73 resistance to invasion of a deciduous forest ecosystem (Warren et al. 2012). Thus, disturbance
74 might be essential for invader establishment when propagule pressure is low, but have weaker
75 effects when propagule pressure is high.

76 An overlooked factor influencing the success of invasive species is propagule quality
77 (here referred to as the physical condition, such as the morphology or biomass of a
78 propagule). For example, the settlement, metamorphosis and performance of the larvae of
79 some invasive marine invertebrates increase with their body size (Marshall et al. 2003,
80 Marshall and Keough 2003, Marshall et al. 2006). Likewise, the number and size of
81 internodes on rhizome fragments of invasive clonal plants with vegetative reproduction
82 significantly affect their settlement and expansion dynamics (Quinn and Holt 2008, Estrada et
83 al. 2016). Production of these propagules by biotic or abiotic disturbance results in fragments
84 of differing size/biomass or containing different morphological components (e.g.
85 combinations of roots/shoots/leaves) that are essential for acquiring different resources
86 (leaves or fronds for photosynthesis; roots for nutrient uptake). Thus, the presence/absence of
87 different morphological components - which may generate fragments of different quality -
88 could be as important as propagule number in regulating individual fragment success (Estrada
89 et al. 2016). Disturbance is, in fact, unlikely to foster the establishment of plant fragments
90 lacking those functional structures necessary for an efficient uptake of freed resources.
91 Despite asexually reproducing plants being some of the most invasive species globally
92 (Kaiser 2000, Allendorf and Lundquist 2003, Williams and Smith 2007), we know very little
93 about how variations in the morphological characteristics of vegetative fragments interact
94 with disturbance regimes to determine invasion success.

95 The clonal seaweed, *Caulerpa cylindracea* Sonder, (previously *Caulerpa racemosa* var.
96 *cylindracea*), is among the most widespread non-native species in the Mediterranean Sea
97 (Renoncourt and Meinesz 2002, Piazzzi and Balata 2009). *C. cylindracea* colonizes a variety
98 of habitats, including dead rhizomes of the native seagrass *Posidonia oceanica* (Linnaeus)
99 Delile (Bulleri et al. 2011). Although intact seagrass meadows appear resistant to *C.*
100 *cylindracea*, canopy removal can promote the establishment of *C. cylindracea* at their
101 margins, suggesting a key role of disturbance in facilitating its establishment (Ceccherelli et
102 al. 2014). *C. cylindracea* can reproduce sexually but mostly spreads through drifting asexual
103 fragments generated by abiotic (i.e. wave surge and currents) or biotic disturbance (i.e.
104 herbivore grazing) (Ceccherelli et al. 2002, Klein and Verlaque 2008, Bulleri et al. 2009).
105 Fragments vary not only in number but also in morphology since they can be formed by
106 prostrate stolons carrying both fronds and rhizoids or lacking either one or both structures
107 (authors' personal observation). In *Caulerpales*, rhizoids provide firm attachment to the
108 substratum and allow nutrient uptake, while fronds are deputed to light harvesting (Komatsu
109 et al. 1997, Chisholm and Moulin 2003). Therefore, the presence/absence of either rhizoids or
110 fronds likely determines the response of *C. cylindracea* to disturbances increasing the
111 availability of different resources. Importantly, however, enhanced uptake of nutrients may
112 not foster establishment if photosynthetic efficiency is impaired by the absence or reduced
113 density of fronds. Moreover, rhizoids, representing a smaller proportion of fragment biomass
114 (Capiomont et al. 2005), can be re-generated more rapidly compared to fronds and their loss
115 has less impact on fragment establishment than frond loss (Bulleri et al. unpublished data).
116 Thus, fragments lacking rhizoids may generally be of higher quality than fragments lacking
117 fronds.

118 Here, we experimentally evaluated how the quality and number of *C. cylindracea*
119 fragments influenced its ability to establish in patches of *P. oceanica* exposed to disturbances

120 of varying intensity (i.e. amount of canopy removal). We predicted that: (1) the ability of
121 fragments to establish would be greater for intact fragments (high quality) and that it would
122 be reduced more by frond removal (low quality) than rhizoid removal (intermediate quality);
123 (2) at either low propagule density or low fragment quality, fragment establishment would
124 increase with increasing disturbance; (3) disturbance would have weaker effects on
125 establishment success at either high propagule density or high fragment quality; (4) propagule
126 quality would have stronger effects on fragment success compared to propagule number at
127 any disturbance level.

128 **2. MATERIALS AND METHODS**

129 *Study system*

130 *Caulerpa cylindracea* is a clonal green alga considered one of the 100 most invasive
131 species in the Mediterranean (Streftaris and Zenetos 2006). It was first recorded in Libya in
132 1990 and has now spread throughout the Mediterranean (see Klein and Verlaque 2008 for
133 review). *C. cylindracea* occurs on a variety of habitats, from rocky shores to soft-sediments,
134 and across a broad depth range, from the intertidal to 70 m (Klein and Verlaque 2008, Bulleri
135 et al. 2011).

136 *Posidonia oceanica* is one of the most important habitat-forming seagrass species in
137 the Mediterranean, occurring across a broad depth range on sandy bottoms (Bethoux and
138 Copinmontegut 1986). *P. oceanica* has leaves up to 70 cm long and forms large, dense beds
139 that support high biodiversity and important fisheries (Ott 1980, Marbà et al. 1996, Guidetti
140 2000).

141 This study was conducted about 10 km south of Livorno (Antignano, 43°29'N,
142 10°19'E; NW Mediterranean), in a dense subtidal *P. oceanica* seagrass meadow (1 ha x 1.5

143 ha, mean \pm SE, 316.4 ± 11.5 shoots per m^2 ; Uyà unpublished data), occurring at 4-8 m water
144 depth and surrounded by a matrix of boulder and sandy substrata. *C. cylindracea* is abundant
145 at this site, often occurring along the margins of the seagrass meadow.

146 **Effects of disturbance, propagule quality and pressure on *C. cylindracea* establishment.**

147 We experimentally tested the effects of disturbance intensity (3 levels; canopy intact,
148 50 % canopy reduction and 100 % canopy reduction), propagule pressure (2 levels; 5 and 10
149 fragments) and propagule quality (3 levels; intact fragments carrying both rhizoids and
150 fronds, fragments lacking either rhizoids or fronds) on the establishment of *C. cylindracea*.
151 Disturbance intensity treatments were created just before the peak in abundance of *C.*
152 *cylindracea*, in mid-July (summer) (Ruitton et al. 2005). We randomly established seventy-
153 two 0.5 x 0.5 m plots at the margin of the *P. oceanica* meadow, at a depth of ~6 m. Plots had
154 an initial seagrass cover of 100 % and were scattered along a 170 m stretch of the meadow.
155 Twenty four plots were then randomly assigned to either 0, 50 or 100% *P. oceanica* cover
156 reduction (Fig. 1). In order to allow the seagrass to recover, reductions in canopy cover were
157 achieved by cutting the leaves whilst leaving the rhizomes intact. Seagrass canopy cover in
158 plots assigned to different disturbance intensity levels were maintained every three weeks
159 throughout the duration of the experiment (10 weeks), by cutting re-grown leaves without
160 disturbing *C. cylindracea* fragments.

161 Four plots for each level of disturbance were then randomly allocated to each of the
162 six combinations of propagule pressure and quality (Fig. 1). Fragments of *C. cylindracea* (10
163 cm stolon length) were collected two days after the implementation of disturbance treatments
164 from a nearby area (~ 100 m away from the experiment location) characterized by dead
165 seagrass rhizomes. Either fronds or rhizoids were removed from fragments in the field, soon
166 after their collection, using scissors. Either five or ten fragments were fixed with U-shaped

167 metal staples within the central 20 x 20 cm area of each plot, to avoid edge effects ($n = 540$
168 fragments in total). Plots were thoroughly searched for the presence of *C. cylindracea* before
169 the experimental transplantation and, when present, it was removed. Both fronds and rhizoids
170 represent a small proportion of the total biomass of the invasive seaweed (up to 12x less
171 biomass than the stolon) (Capiomont et al. 2005), thus we controlled for initial propagule
172 length, and not biomass, as differences between propagule morphologies in biomass were
173 considered negligible. The attachment of all fragments was checked two days after they were
174 transplanted to ensure all fragments had remained in place.

175 *Sampling and data analysis*

176 The percentage cover of *C. cylindracea* in each plot was visually estimated after three
177 (August), six (September) and ten weeks (October) from the start of the experiment, using a
178 20 x 20 cm plastic frame subdivided into 25 sub-quadrats. A score from 0 (absence) to 4
179 (completely covered) was given to each sub-quadrat and the percentage cover was obtained
180 by summing over the entire set of sub-quadrats (Dethier et al. 1993). At the end of the
181 experiment, *C. cylindracea* fragments were retrieved from the central 20 x 20 cm area of each
182 plot and brought to the lab for analysis. The total number of fronds, mean length of three
183 randomly selected fronds, total stolon length and the number of rhizoids per quadrat were
184 measured. Finally, total fragment biomass per quadrat was estimated as dry weight ($\text{g} \cdot \text{m}^2$)
185 after drying the material at 60 °C for 48 hours.

186 The effects of canopy disturbance, propagule number and quality on the percentage
187 cover of *C. cylindracea* were analyzed using a linear mixed model. Seagrass disturbance
188 intensity, propagule number and quality were considered as fixed effects. Time of sampling
189 was considered as a random effect to take into account temporal auto-correlation in the data
190 generated by repeatedly sampling the same quadrats. The baseline for the linear mixed model

191 was set *a priori* as the, supposedly, most invisable scenario: the combination of 100 %
192 disturbance inoculated with 10 intact fragments. Chi-Square likelihood ratio tests were used
193 to detect significant differences amongst treatments. The analysis was performed in R
194 (version 3.3.2) using the lmer function within the lme4 package and the anova function with
195 Chisq test (Bates et al. 2015). When significant main effects were detected, multiple *post-hoc*
196 TukeyHSD tests were used to determine differences among the levels of disturbance intensity
197 and propagule quality using the glht function within the multcomp package (Hothorn et al.
198 2013). Assumptions of linearity and variance homogeneity were checked by plotting the
199 standardized residuals against fitted values (Zuur et al. 2009).

200 The effects of experimental conditions on the density and length of fronds, total stolon
201 length, density of rhizoids and biomass of *C. cylindracea*, sampled at the end of the
202 experiment, were analyzed by means of three-way ANOVA. The model included canopy
203 disturbance, propagule number and quality as fixed, crossed factors. Data were square root
204 transformed when Cochran's test indicated significant heterogeneity of variances
205 (Underwood 1997). TukeyHSD tests were used for *post-hoc* comparison of the means. One
206 plot assigned to the 100 % seagrass cover reduction and transplanted with 10 intact fragments
207 of *C. cylindracea* was lost at the start of the experiment. In order to maintain a balanced
208 design, the missing value was replaced with the mean of the remaining replicates for this
209 treatment and residual degrees of freedom were adjusted accordingly (Underwood 1997). All
210 ANOVA tests were performed in R (version 3.3.2) using the lm function within the GAD
211 package (Sandrini-Neto and Camargo 2012).

212 **3. RESULTS**

213 Disturbance, propagule pressure and quality had significant effects on the cover of *C.*
214 *cylindracea* in experimental plots (Fig 2, Table 1A). Across the study, the cover of the

215 seaweed increased significantly with increasing disturbance intensity. In particular, the
216 percent cover of the seaweed was, on average, 30 and 10 times higher in 100 % and 50 %
217 canopy removals compared to control treatments, respectively, on the last sampling date (Fig.
218 2, Table 1B).

219 The cover of *C. cylindracea* was significantly higher in plots inoculated with intact
220 and no-rhizoid fragments than no-frond fragments (Fig. 2, Table 1) and increased with
221 increasing propagule pressure (Fig. 2, Table 1A). When transplanted at high density (i.e. 10
222 fragments/plot), some intact and no-rhizoid fragments were also able to persist under intact
223 canopies throughout the experiment. The interaction between disturbance and either
224 propagule pressure or quality was non-significant, but only marginally so ($P \leq 0.06$, Table
225 1A).

226 At the end of the experiment, the density of fronds and rhizoids, total stolon length
227 and biomass increased with increasing intensity of disturbance (Fig. 3-4, Table 2A and
228 Appendix S1: Fig. S1-S3). Moreover, total stolon length increased significantly with the
229 number of fragments inoculated (Table 2A, Appendix S1: Fig S3). There was a significant
230 effect of the interaction Disturbance x Propagule quality on the mean length of fronds. The
231 *post-hoc* test indicated that fronds grew longer from intact and no-rhizoid fragments than no-
232 frond fragments in 50 % canopy reduction plots, while no differences occurred for the other
233 disturbance levels (Table 2B and Appendix S1: Fig S1).

234 The biomass of *C. cylindracea* was up to 50x and 80x higher in the 50 % and 100 %
235 canopy cover reduction treatments, respectively, when compared to intact *P. oceanica*. In
236 addition, in 100 % canopy reduction plots inoculated with 5 fragments, final seaweed
237 biomass was about 2.4x higher for intact than no-fronds (Fig. 4, Table 2A). Biomass was also

238 significantly higher in plots inoculated with intact and no-rhizoid fragments than plots
239 inoculated with no-frond fragments (Fig. 4, Table 2B).

240 **4. DISCUSSION**

241 While there is increasing acknowledgement that invasion success is regulated by
242 multiple factors, experimental tests including combinations of factors are still relatively
243 uncommon (but see Thomsen et al. 2006, Clark and Johnston 2009, Estrada et al. 2016, Byun
244 et al. 2017). Here, we show that disturbance to native canopies, propagule quality and number
245 influence invasion success. As predicted, disturbance had the strongest effect on invasion
246 success, with seaweed cover, biomass, total stolon length and frond density per plot
247 increasing with growing intensity of disturbance to seagrass canopies. In addition, propagule
248 quality had a stronger influence on fragment biomass and traits compared to propagule
249 number. The effects of the individual factors were independent and did not interact to
250 determine fragment establishment.

251 Increasing reductions in the cover of the seagrass *Posidonia oceanica* enhanced the
252 establishment of *C. cylindracea*. This supports previous studies showing positive effects of
253 disturbance to native assemblages on invasion success via increased resource availability
254 (Burke and Grime 1996, Davis et al. 2000, Bulleri et al. 2010, Ceccherelli et al. 2014).
255 Although we did not investigate how disturbance of *P. oceanica* canopy promoted the
256 establishment of *C. cylindracea*, the main mechanism was likely an increase in access to light
257 (Marin-Guirao et al. 2015). Light levels can be reduced by 60-89 % in intact seagrass beds
258 compared to outside, potentially limiting the growth of *C. cylindracea* (Marin-Guirao et al.
259 2015). In fact, in our study, a 50 % canopy reduction was sufficient to increase the cover,
260 biomass and frond density of *C. cylindracea* to values observed in the 100 % canopy removal
261 treatment. Thus, invasion of native marine macrophytes may happen at lower levels of

262 disturbance than often considered, highlighting the limitations of marine studies that have
263 typically considered removal of the entire native canopy vs. non-removal (Valentine and
264 Johnson 2003, Britton-Simmons and Abbott 2008, Bulleri et al. 2010). Understanding critical
265 levels, or thresholds, at which disturbance facilitates invasion is clearly an important avenue
266 for future research. This is a pressing issue for *C. cylindracea* given the current declining
267 trends of *P. oceanica* seagrass meadows in the Mediterranean Sea (Marbà et al. 2014).

268 General theory and empirical research provide compelling evidence of the role of
269 species' traits in regulating establishment success (Stearns 1992, Marshall et al. 2006).
270 Although rarely studied in an invasion context, fragment quality had strong consequences for
271 cover and total biomass measured at the end of the experiment. Importantly, fragments
272 without fronds consistently performed worst when compared to fragments with fronds (both
273 intact and without rhizoids). This was somewhat surprising given that the rhizoids of other
274 invasive *Caulerpa* spp. perform vital functions such as fixing N₂, which in turns promotes
275 organic matter turnover and nutrient acquisition (Chisholm and Moulin 2003). The strong
276 effect of the presence/absence of fronds suggests that light was, again, the likely limiting
277 resource for *C. cylindracea*.

278 The stolon and fronds of *C. cylindracea* contain photosynthetic pigments such as
279 chlorophyll *a*, siphonoxanthin and siphonein (Raniello et al. 2004). *C. cylindracea* fragments
280 initially deprived of fronds had lower total stolon length and biomass, but, by the end of the
281 experiment, they were able to re-grow fronds to densities matching those found in other types
282 of fragments. Since this seaweed is coenocytic, the removal of fronds could result in a rapid
283 reallocation of energy to increase their re-growth at expense of the lateral expansion, as found
284 for several other *Caulerpa* species (Collado-Vides and Robledo 1999). The lack of
285 differences in rhizoid density between fragments that had their rhizoids removed and intact
286 ones at the end of the experiment suggests that there is also a reallocation of energy to rhizoid

287 re-growth. Because of the small loss of biomass associated with rhizoid removal (Capiomont
288 et al. 2005), new rhizoids may be quickly generated without major consequences on final
289 biomass and total stolon length.

290 Increasing fragment number, although having weaker effects than disturbance and
291 fragment quality on seaweed establishment, increased the final percent cover of *C.*
292 *cylindracea*. However, a doubling from 5 to 10 fragments/quadrat did not result in a doubling
293 of percent cover or biomass of fragments in plots. This might be a consequence of
294 intraspecific competition among fragments at the highest fragment number treatment.
295 Competition for light among fragments is unlikely to explain this pattern, as *C cylindracea*
296 can be found up to 70 m deep at high densities (Klein and Verlaque 2008). At higher
297 densities, fragments may be competing for other resources such as nutrient supply. Intra-
298 specific competition as a consequence of nutrient limitation among early growth stages has
299 been documented in invasive terrestrial plants (Blank 2010) and marine macroalgae (Steen
300 2003, Steen and Scrosati 2004). We suggest that intraspecific competition among invasive
301 propagules may have important, but currently underestimated, consequences for the
302 establishment and demography of non-native plants.

303 Previous studies have clearly demonstrated that different factors (e.g. disturbance and
304 propagule pressure) interact to determine invasion success (D'Antonio et al. 2001, Thomsen
305 et al. 2006, Britton-Simmons and Abbott 2008). In our experiment, the effects of canopy
306 disturbance, fragment quality and number were largely independent. Our prediction that the
307 success of low quality fragments (fragments without fronds) would increase with increasing
308 disturbance levels was supported, as no-frond fragments could establish only in total canopy
309 removal plots. In contrast, the prediction that high quality fragments (i.e. intact) would be
310 successful across all disturbance regimes, independently of their number was not supported;
311 in fact, high and intermediate quality fragments persisted under intact seagrass canopy only

312 when transplanted in large numbers. This suggests subtle interactions among disturbance,
313 propagule quality and number may occur, but were not detected in our analyses because of
314 the low values of cover and biomass achieved by *C. cylindracea* under full canopies. *P*-
315 values close to significance for interactions between disturbance and either propagule quality
316 or propagule pressure on seaweed cover (see Table 1) support the proposition that the cover
317 of *C. cylindracea*, at the peak of the growing season (when our experiment was conducted),
318 may be dependent upon the characteristics of both propagules and the recipient habitat.

319 In summary, our results suggest a stronger role of propagule quality compared to
320 propagule number in the establishment of *C. cylindracea*. For fast-growing, asexually
321 reproducing invasive species, propagule traits enhancing the ability to acquire the specific
322 resources freed by disturbance - in our case light - appear key for their successful
323 establishment and spread. Under these circumstances, the traits of propagules conferring
324 greater establishment ability likely vary among disturbances that free different resources (i.e.
325 beneficial traits may be disturbance-specific). Our results may also reconcile contrasting
326 results of the effects of disturbance on invasion success (Moles et al. 2012): disturbance may
327 facilitate non-native establishment only when propagules possess the traits necessary to
328 acquire freed resources. By contrast, weak effects of resource release or input can be
329 expected when propagules are unable to exploit them, independently of their abundance.
330 Under these circumstances, assessing the matching between the quality or type of resource
331 made available through disturbance and the characteristics of non-native propagules may
332 greatly enhance our ability to forecast invasion success in environments increasingly exposed
333 to human perturbations. More broadly, a functional approach, based on resource-exploitation
334 traits of propagules (i.e. fragments, larvae, spores or adults), might allow identifying non-
335 native species or morpho-types within species more likely to benefit from a specific
336 disturbance regime.

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347

348 **REFERENCES**

- 349 Allendorf, F. W., and L. L. Lundquist. 2003. Introduction: Population biology, evolution, and
350 control of invasive species. *Conservation Biology* 17:24-30.
- 351 Bates, D., M. Maecheler, B. Bolker, and S. Walker. 2015. lme4: Fitting linear mixed-model
352 effects using lme4. *Journal of Statistical Software* 67:1-48.
- 353 Bethoux, J. P., and G. Copinmontegut. 1986. Biological fixation of atmospheric nitrogen in
354 the Mediterranean Sea. *Limnology and Oceanography* 31:1353-1358.
- 355 Blank, R. R. 2010. Intraspecific and interspecific pair-wise seedling competition between
356 exotic annual grasses and native perennials: plant-soil relationships. *Plant and Soil*
357 326:331-343.
- 358 Britton-Simmons, K. H., and K. C. Abbott. 2008. Short- and long-term effects of disturbance
359 and propagule pressure on a biological invasion. *Journal of Ecology* 96:68-77.
- 360 Bulleri, F., L. Tamburello, and L. Benedetti-Cecchi. 2009. Loss of consumers alters the
361 effects of resident assemblages on the local spread of an introduced macroalga. *Oikos*
362 118:269-279.
- 363 Bulleri, F., D. Balata, I. Bertocci, L. Tamburello, and L. Benedetti-Cecchi. 2010. The
364 seaweed *Caulerpa racemosa* on Mediterranean rocky reefs: from passenger to driver
365 of ecological change. *Ecology* 91:2205-2212.
- 366 Bulleri, F., T. Alestra, G. Ceccherelli, L. Tamburello, S. Pinna, N. Sechi, and L. Benedetti-
367 Cecchi (2011) Determinants of *Caulerpa racemosa* distribution in the north-western
368 Mediterranean. *Mar Ecol Prog Ser* 431:55–67
- 369 Burke, M. J. W., and J. P. Grime. 1996. An experimental study of plant community
370 invasibility. *Ecology* 77:776-790.
- 371 Byun, C., S. de Blois, and J. Brisson. 2015. Interactions between abiotic constraint, propagule
372 pressure, and biotic resistance regulate plant invasion. *Oecologia* 178:285-296.

373 Byun, C., S. de Blois, and J. Brisson. 2017. Management of invasive plants through
374 ecological resistance. *Biological Invasions*.

375 Capiomont, A., E. Breugnot, M. den Haan, and A. Meinesz. 2005. Phenology of a deep-water
376 population of *Caulerpa racemosa* var. *cylindracea* in the northwestern Mediterranean
377 Sea. *Botanica Marina* 48:80-83.

378 Ceccherelli, G., L. Piazzzi, and D. Balata. 2002. Spread of introduced *Caulerpa* species in
379 macroalgal habitats. *Journal of Experimental Marine Biology and Ecology* 280:1-11.

380 Ceccherelli, G., S. Pinna, V. Cusceddu, and F. Bulleri. 2014. The role of disturbance in
381 promoting the spread of the invasive seaweed *Caulerpa racemosa* in seagrass
382 meadows. *Biological Invasions* 16:2737-2745.

383 Chisholm, J. R. M., and P. Moulin. 2003. Stimulation of nitrogen fixation in refractory
384 organic sediments by *Caulerpa taxifolia* (Chlorophyta). *Limnology and*
385 *Oceanography* 48:787-794.

386 Clark, G. F., and E. L. Johnston. 2005. Manipulating larval supply in the field: a controlled
387 study of marine invasibility. *Marine Ecology Progress Series* 298:9-19.

388 Clark, G. F., and E. L. Johnston. 2009. Propagule pressure and disturbance interact to
389 overcome biotic resistance of marine invertebrate communities. *Oikos* 118:1679-
390 1686.

391 Collado-Vides, L., and D. Robledo. 1999. Morphology and photosynthesis of *Caulerpa*
392 (Chlorophyta) in relation to growth form. *Journal of Phycology* 35:325-330.

393 Corbin, J. D., and C. M. D'Antonio. 2004. Competition between native perennial and exotic
394 annual grasses: Implications for an historical invasion. *Ecology* 85:1273-1283.

395 D'Antonio, C. M., and P. M. Vitousek. 1992. Biological invasions by exotic grasses, the grass
396 fire cycle, and global change. *Annual Review of Ecology and Systematics* 23:63-87.

397 D'Antonio, C. M., J. M. Levine, and M. Thomsen. 2001. Ecosystem resistance to invasion
398 and the role of propagule supply: a California perspective. *Journal of Mediterranean*
399 *Ecology* 2:233-245

400 Davis, M. A., J. P. Grime, and K. Thompson. 2000. Fluctuating resources in plant
401 communities: a general theory of invasibility. *Journal of Ecology* 88:528-534.

402 Dethier, M. N., E. S. Graham, S. Cohen, and L. M. Tear. 1993. Visual versus random-point
403 percent cover estimations - objective is not always better. *Marine Ecology Progress*
404 *Series* 96:93-100.

405 Dextrase, A. J., and N. E. Mandrak. 2006. Impacts of Alien Invasive Species on Freshwater
406 Fauna at Risk in Canada. *Biological Invasions* 8:13-24.

407 Elton, C. S. 1958. *The ecology of invasions by animals and plants*. Methuen, London.

408 Eschtruth, A. K., and J. J. Battles. 2009. Assessing the relative importance of disturbance,
409 herbivory, diversity, and propagule pressure in exotic plant invasion. *Ecological*
410 *Monographs* 79:265-280.

411 Estrada, J. A., C. H. Wilson, J. E. NeSmith, and S. L. Flory. 2016. Propagule quality mediates
412 invasive plant establishment. *Biological Invasions* 18:2325-2332.

413 Guidetti, P. 2000. Differences among fish assemblages associated with nearshore *Posidonia*
414 *oceanica* seagrass beds, rocky-algal reefs and unvegetated sand habitats in the
415 Adriatic Sea. *Estuarine Coastal and Shelf Science* 50:515-529.

416 Hobbs, R. J., and L. F. Huenneke. 1992. Disturbance, diversity, and invasion - implications
417 for conservations. *Conservation Biology* 6:324-337.

418 Hollebhone, A. L., and M. E. Hay. 2007. Propagule pressure of an invasive crab overwhelms
419 native biotic resistance. *Marine Ecology Progress Series* 342:191-196.

420 Hollebhone, A. L., and M. E. Hay. 2008. An invasive crab alters interaction webs in a marine
421 community. *Biological Invasions* 10:347-358.

422 Hothorn, T., F. Bretz, P. Westfall, R. M. Heiberger, and A. Schuetzenmeister. 2013. Package
423 “multcomp”. R package version 1.4-6.

424 Kaiser, J. 2000. Ecology - California algae may be feared European species. *Science*
425 289:222-223.

426 Klein, J., and M. Verlaque. 2008. The *Caulerpa racemosa* invasion: A critical review. *Marine*
427 *Pollution Bulletin* 56:205-225.

428 Komatsu, T., A. Meinesz, and D. Buckles. 1997. Temperature and light responses of alga
429 *Caulerpa taxifolia* introduced into the Mediterranean Sea. *Marine Ecology Progress*
430 *Series* 146:145-153.

431 Leung, B., and N. E. Mandrak. 2007. The risk of establishment of aquatic invasive species:
432 joining invasibility and propagule pressure. *Proceedings of the Royal Society B-*
433 *Biological Sciences* 274:2603-2609.

434 Lockwood, J. L., P. Cassey, and T. Blackburn. 2005. The role of propagule pressure in
435 explaining species invasions. *Trends in Ecology & Evolution* 20:223-228.

436 Mack, R. N., D. Simberloff, W. M. Lonsdale, H. Evans, M. Clout, and F. A. Bazzaz. 2000.
437 Biotic invasions: Causes, epidemiology, global consequences, and control. *Ecological*
438 *Applications* 10:689-710.

439 Marbà, N., C. M. Duarte, J. Cebrian, M. E. Gallegos, B. Olesen, and K. SandJensen. 1996.
440 Growth and population dynamics of *Posidonia oceanica* on the Spanish
441 Mediterranean coast: Elucidating seagrass decline. *Marine Ecology Progress Series*
442 137:203-213.

443 Marbà, N., E. Diaz-Almela, and C. M. Duarte. 2014. Mediterranean seagrass (*Posidonia*
444 *oceanica*) loss between 1842 and 2009. *Biological Conservation* 176:183-190.

445 Marin-Guirao, L., J. Bernardeau-Esteller, J. M. Ruiz, and J. M. Sandoval-Gil. 2015.
446 Resistance of *Posidonia oceanica* seagrass meadows to the spread of the introduced

447 green alga *Caulerpa cylindracea*: assessment of the role of light. *Biological Invasions*
448 17:1989-2009.

449 Marshall, D. J., T. F. Bolton, and M. J. Keough. 2003. Offspring size affects the post-
450 metamorphic performance of a colonial marine invertebrate. *Ecology* 84:3131-3137.

451 Marshall, D. J., and M. J. Keough. 2003. Variation in the dispersal potential of non-feeding
452 invertebrate larvae: the desperate larva hypothesis and larval size. *Marine Ecology*
453 *Progress Series* 255:145-153.

454 Marshall, D. J., C. N. Cook, and R. B. Emlet. 2006. Offspring size effects mediate
455 competitive interactions in a colonial marine invertebrate. *Ecology* 87:214-225.

456 Moles, A. T., H. Flores-Moreno, S. P. Bonser, D. I. Warton, A. Helm, L. Warman, D. J.
457 Eldridge, E. Jurado, F. A. Hemmings, P. B. Reich, J. Cavender-Bares, E. W.
458 Seabloom, M. M. Mayfield, D. Sheil, J. C. Djietror, P. L. Peri, L. Enrico, M. R.
459 Cabido, S. A. Setterfield, C. E. R. Lehmann, and F. J. Thomson. 2012. Invasions: the
460 trail behind, the path ahead, and a test of a disturbing idea. *Journal of Ecology*
461 100:116-127.

462 Ott, J. A. 1980. Growth and Production in *Posidonia Oceanica* (L.) Delile*. *Marine Ecology*
463 1:47-64.

464 Piazzzi, L., and G. Ceccherelli. 2002. Effects of competition between two introduced
465 *Caulerpa*. *Marine Ecology Progress Series* 225:189-195.

466 Piazzzi, L., and D. Balata. 2009. Invasion of alien macroalgae in different Mediterranean
467 habitats. *Biological Invasions* 11:193-204.

468 Quinn, L. D., and J. S. Holt. 2008. Ecological correlates of invasion by *Arundo donax* in three
469 southern California riparian habitats. *Biological Invasions* 10:591-601.

470 Raniello, R., M. Lorenti, C. Brunet, and M. C. Buia. 2004. Photosynthetic plasticity of an
471 invasive variety of *Caulerpa racemosa* in a coastal Mediterranean area: light

472 harvesting capacity and seasonal acclimation. Marine Ecology Progress Series
473 271:113-120.

474 Renoncourt, L., and A. Meinesz. 2002. Formation of propagules on an invasive strain of
475 *Caulerpa racemosa* (Chlorophyta) in the Mediterranean Sea. Phycologia 41:533-535.

476 Ruitton, S., M. Verlaque, and C. F. Boudouresque. 2005. Seasonal changes of the introduced
477 *Caulerpa racemosa* var. *cylindracea* (Caulerpales, Chlorophyta) at the northwest limit
478 of its Mediterranean range. Aquatic Botany 82:55-70.

479 Sandrini-Neto, L., and M. G. Camargo. 2012. GAD: Analysis of variance from general
480 principles. R package version 1.1.1.

481 Stachowicz, J. J., H. Fried, R. W. Osman, and R. B. Whitlatch. 2002. Biodiversity, invasion
482 resistance, and marine ecosystem function: Reconciling pattern and process. Ecology
483 83:2575-2590.

484 Stearns, S. C. 1992. The evolution of life histories, Oxford, UK.

485 Steen, H. 2003. Intraspecific competition in *Sargassum muticum* (Phaeophyceae) germlings
486 under various density, nutrient and temperature regimes. Botanica Marina 46:36-43.

487 Steen, H., and R. Scrosati. 2004. Intraspecific competition in *Fucus serratus* and *F.*
488 *evanescens* (Phaeophyceae : Fucales) germlings: effects of settlement density, nutrient
489 concentration, and temperature. Marine Biology 144:61-70.

490 Streftaris, N., and A. Zenetos. 2006. Alien Marine Species in the Mediterranean - the 100
491 'Worst Invasives' and their Impact. Mediterranean Marine Science 7:87-117.

492 Thomsen, M. A., C. M. D'Antonio, K. B. Suttle, and W. P. Sousa. 2006. Ecological
493 resistance, seed density and their interactions determine patterns of invasion in a
494 California coastal grassland. Ecology Letters 9:160-170.

495 Underwood, A. J. 1997. Experiments in ecology: their logical design and interpretation using
496 analysis of variance. Cambridge University Press, Cambridge.

497 Valentine, J. P., and C. R. Johnson. 2003. Establishment of the introduced kelp *Undaria*
498 *pinnatifida* in Tasmania depends on disturbance to native algal assemblages. Journal
499 of Experimental Marine Biology and Ecology 295:63-90.

500 Warren, R. J., V. Bahn, and M. A. Bradford. 2012. The interaction between propagule
501 pressure, habitat suitability and density-dependent reproduction in species invasion.
502 *Oikos* 121:874-881.

503 Williams, S. L., and J. E. Smith. 2007. A global review of the distribution, taxonomy, and
504 impacts of introduced seaweeds. Pages 327-359 *Annual Review of Ecology*
505 *Evolution and Systematics*. Annual Reviews, Palo Alto.

506 Zuur, A. F., E. N. Ieno, N. J. Walker, A. A. Saveliev, and G. M. Smith. 2009. Mixed effects
507 models and extensions in ecology with R. 1 edition. Springer-Verlag New York, New
508 York.

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519 **Table 1.** (A) Chi-Square (X^2) likelihood ratio test and (B) *post-hoc* Tukey HSD tests of the
 520 linear mixed model on the effects of disturbance intensity (D), propagule pressure (P) and
 521 propagule quality (Q) on the percentage cover of *Caulerpa cylindracea*. The baseline for the
 522 model is the supposedly most invisable scenario: 100% disturbance inoculated with 10 intact
 523 fragments.

A) Chi-Square (X^2) likelihood ratio test

Source of Variation	df	MS	F-value
Disturbance intensity (D)	2	10419.5	68.801***
Propagule pressure (P)	1	1583.3	10.455**
Propagule quality (Q)	2	1556.8	10.280***
DxP	2	436.1	2.880 ·
DxQ	4	327.3	2.161 ·
PxQ	2	268.2	1.77
DxPxQ	4	89.9	0.594

Significance codes: · $p \leq 0.06$; * $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$.

B) *post-hoc* Tukey HSD tests

D	100 % > 50 % > Control
Q	Intact = no rhizoids > no fronds

Table 2. (A) Source of Variation and (B) *post-hoc* Tukey HSD tests on the effects of disturbance intensity (D), propagule pressure (P) and propagule quality (Q) on frond density, mean frond length, total stolon length, rhizoid density and biomass of *Caulerpa cylindracea*.

A) Source of Variation

Source of Variation	df	Frond density		Mean frond length		Total stolon length		Rhizoid density		Biomass	
		MS	F	MS	F	MS	F	MS	F	MS	F
D	2	16368.7	27.871***	12.125	47.925***	104601	21.086***	35212.0	16.638***	1.746	39.682***
P	1	1250	2.128	0.001	0.004	23875	4.181*	5425.0	2.563	0.071	1.614
Q	2	1320.9	2.247	1.794	7.091**	13461	2.714	4150.0	1.961	0.191	4.341*
DxP	2	339.5	0.578	0.030	0.119	6323	1,275	1803.0	0.852	0.009	0.205
DxQ	4	384.9	0.655	0.884	3.494*	3956	0.797	1303.0	0.616	0.058	1.312
PxQ	2	174.9	0.300	0.188	0.743	2874	0.579	216.0	0.102	0.046	1.045
DxPxQ	4	148.9	0.254	0.414	1.636	2828	0.570	682.0	0.322	0.044	1
Residual	53	587.3		0.253		4960.6		2116.3		0.044	

Transformation	none	none	none	none	Square root
Cochran's test	p<0.05	p=0.120	p<0.05	p<0.05	p=0.120

Significance codes: *p<0.05; ** p<0.01; *** p<0.001.

B) *post-hoc* Tukey HSD tests

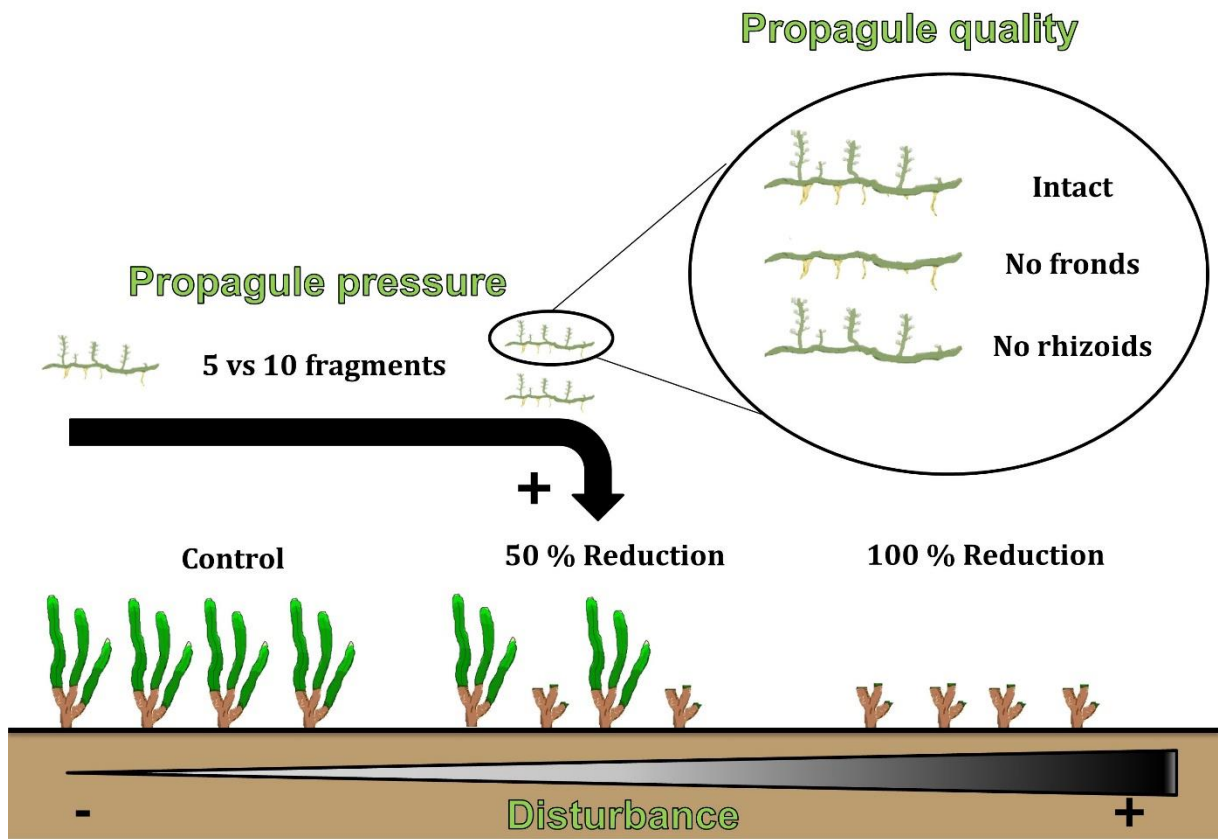
Frond density	D	100 % > 50 % > Control
Mean frond length	DxQ	Control: Intact = no rhizoids = no fronds 50 %: Intact = no rhizoids > no fronds 100 %: Intact = no rhizoids = no fronds
Total stolon length	D	100 % > 50 % > Control
Rhizoid density	D	100 % > 50 % ; 100 % > Control
Biomass	D Q	100 % > 50 % > Control Intact = no rhizoids > no fronds

Figure 1. Schematic diagram of the fully factorial experiment conducted in a *Posidonia oceanica* seagrass meadow investigating how disturbance intensity, propagule pressure and propagule quality regulate the establishment of *Caulerpa cylindracea* fragments. Four plots were assigned to each combination of disturbance intensity (3 levels), propagule pressure (2 levels) and propagule quality (3 levels), resulting in a total of 72 manipulated plots.

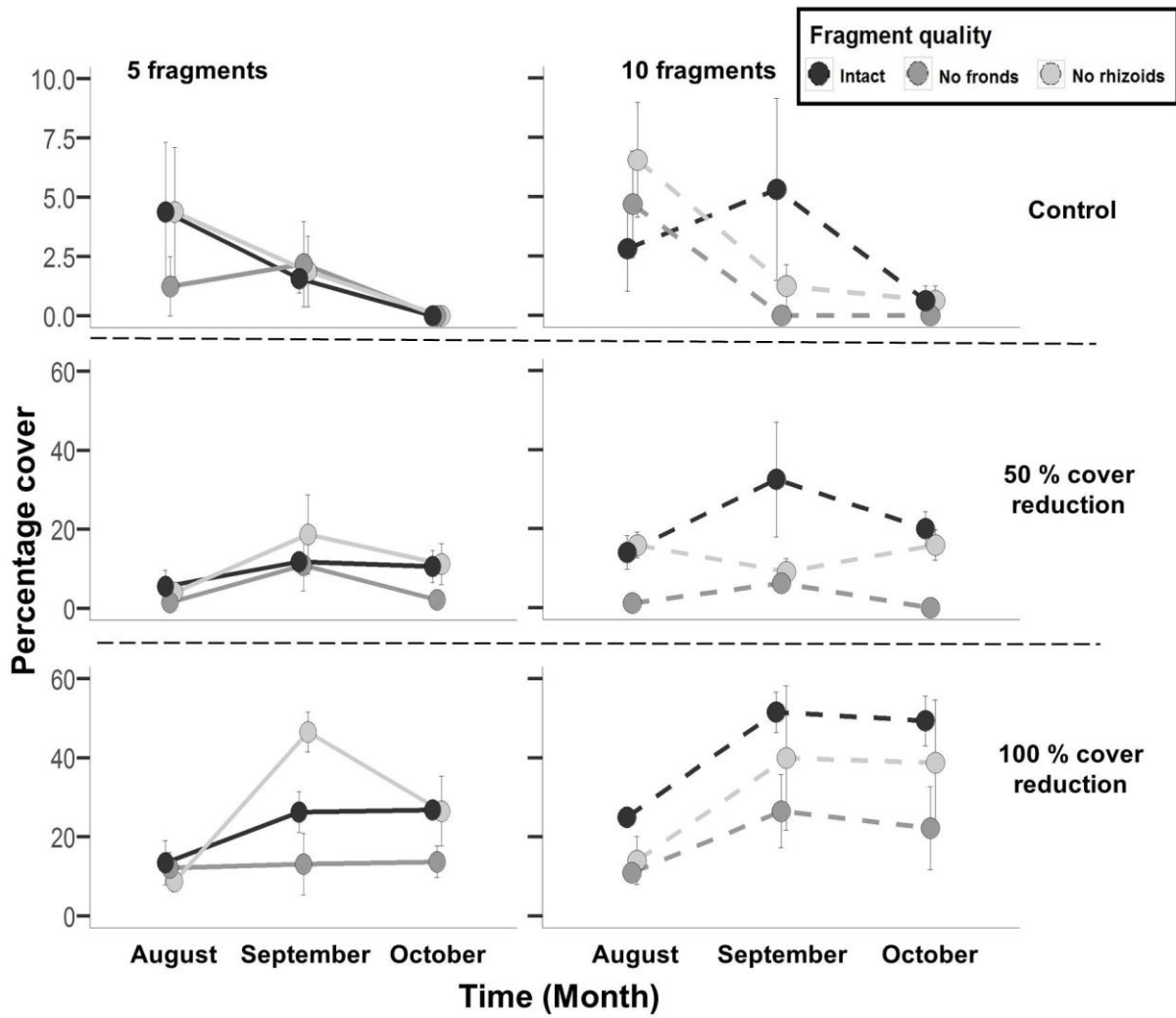
Figure 2. Temporal trend of the percentage cover of *Caulerpa cylindracea* under different combinations of disturbance intensity, propagule pressure and propagule quality. Data are means \pm Standard Error. Continuous lines represent 5 fragment treatments whilst dashed lines indicate 10 fragment treatments. Differences in y-axes scale should be considered when comparing disturbance intensity levels.

Figure 3. Density of fronds of *Caulerpa cylindracea* (number/m²) under different combinations of disturbance intensity, propagule pressure and propagule quality. Data are means \pm Standard Error. Differences in y-axes scale should be considered when comparing disturbance intensity levels.

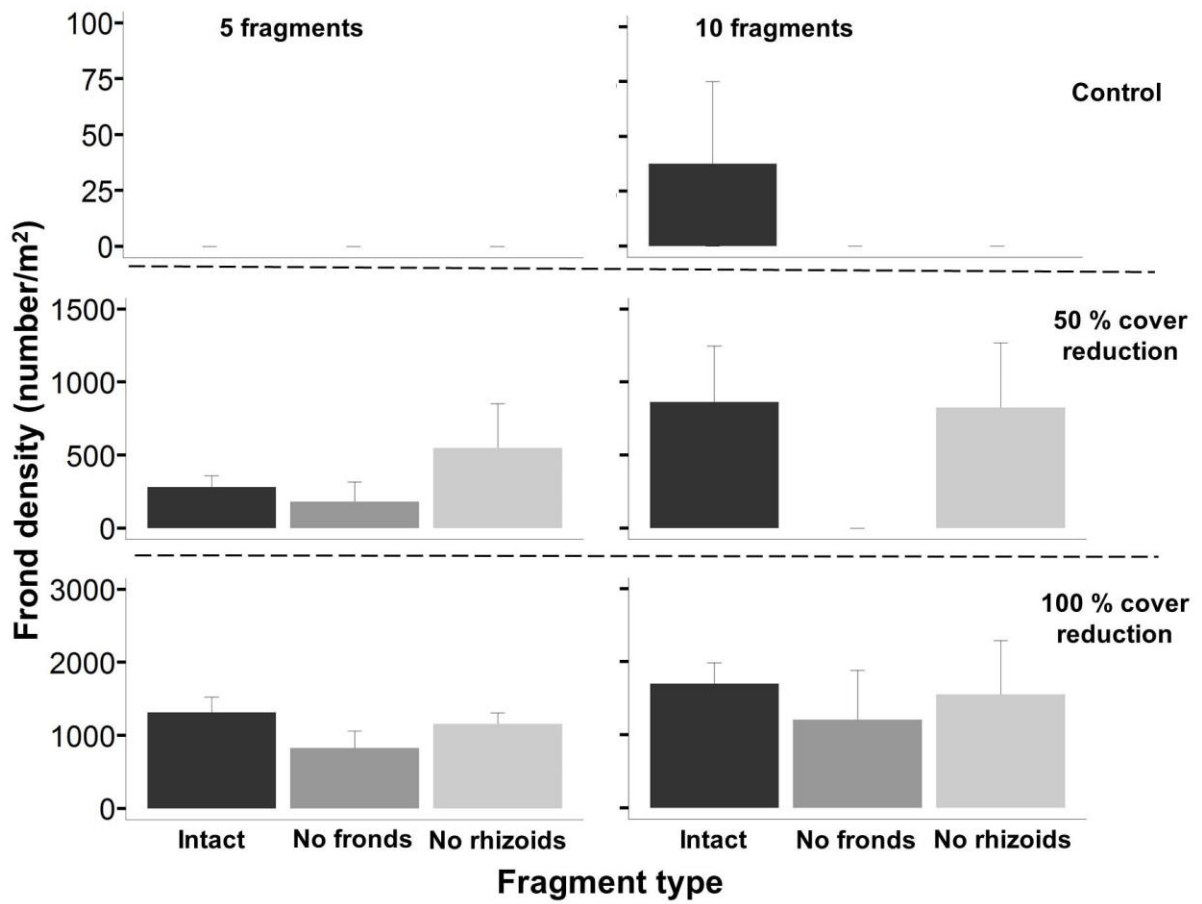
Figure 4. Biomass of *Caulerpa cylindracea* (g/m²) under different combinations of disturbance intensity, propagule pressure and propagule quality. Data are means \pm Standard Error. Differences in y-axes scale should be considered when comparing disturbance intensity levels.



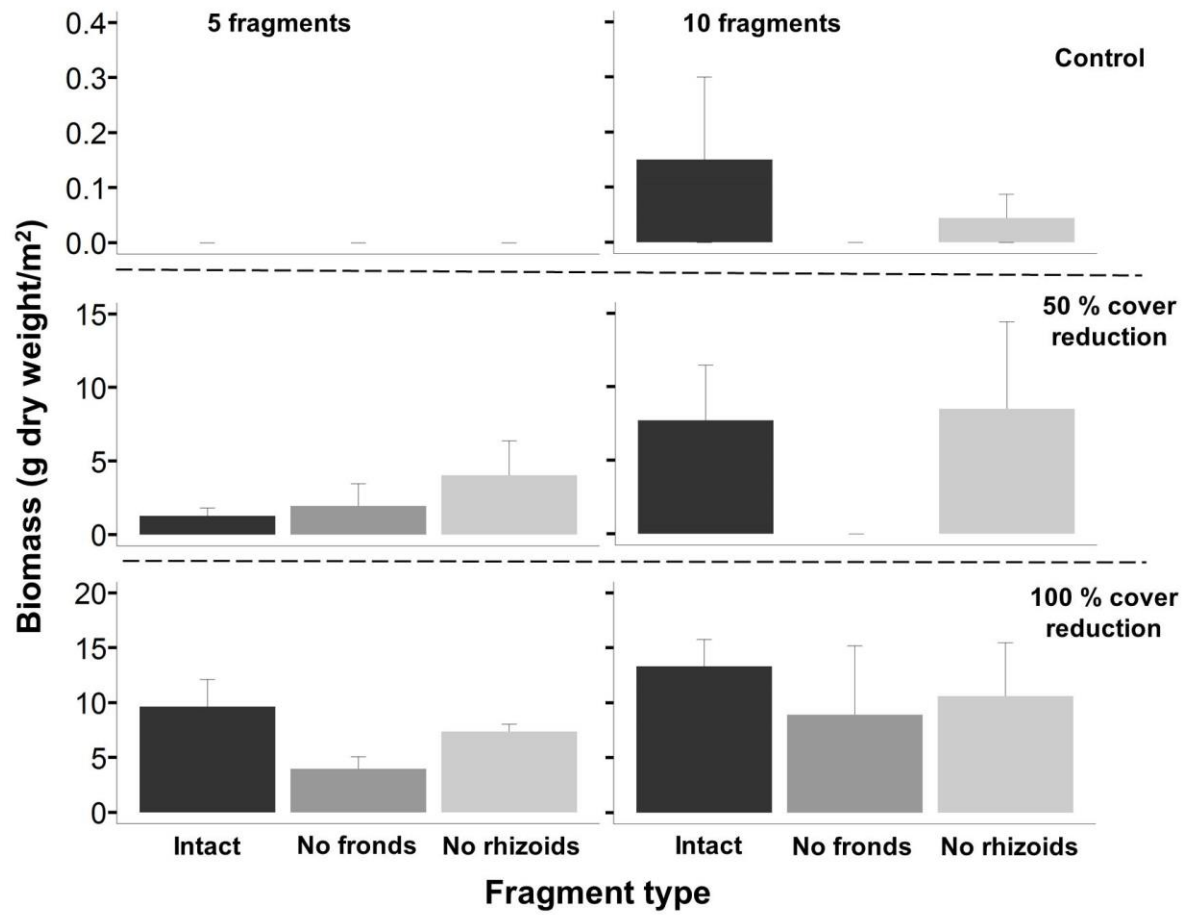
Uyà et al. Figure 1



Uyà et al. Figure 2



Uyà et al. Figure 3



Uyà et al. Figure 4