

1 Loss of genetic diversity and increased embryonic mortality in non-native lizard populations

2 Sozos N. Michaelides^{1*}, Geoffrey M. While ^{1,2}, Natalia Zajac¹, Fabien Aubret³, Brittny Calsbeek⁴, Roberto
3 Sacchi⁵, Marco A.L. Zuffi⁶, Tobias Uller^{1,7*}

4

5 ¹Edward Grey Institute, Department of Zoology, University of Oxford, South Parks Rd, OX1 3PS, Oxford, UK.

6 ²School of Biological Sciences, University of Tasmania, PO Box 55, Hobart, Tas. 7001, Australia.

7 ³Station d'Ecologie Expérimentale du CNRS à Moulis, 09200 Moulis, France.

8 ⁴Department of Biological Sciences, Dartmouth College, Hanover, NH 03755, USA.

9 ⁵Dipartimento Sci Terra & Ambiente, Lab Ecoetol, Università di Pavia, I-27100 Pavia, Italy.

10 ⁶Museo di Storia Naturale, Università di Pisa, Via Roma, 79 56011 Calci, Pisa, Italy.

11 ⁷Department of Biology, Lund University, Sölvegatan 37, SE 223 62 Lund, Sweden.

12

13 *Corresponding authors

14 msozos@gmail.com

15 tobias.uller@biol.lu.se

16

17 Keywords: colonization, genetic diversity, inbreeding, hatching failure, lizard

18

19 Running head: Genetic diversity and inbreeding in alien lizards

20

21

22

23

24

25

26

27

28

29

30

31

32 **Abstract**

33 Many populations are small and isolated with limited genetic variation and high risk of mating with
34 close relatives. Inbreeding depression is suspected to contribute to extinction of wild populations, but
35 the historical and demographic factors that contribute to reduced population viability are often difficult
36 to tease apart. Replicated introduction events in non-native species can offer insights into this problem
37 because they allow us to study how genetic variation and inbreeding depression are affected by
38 demographic events (e.g., bottlenecks), genetic admixture and the extent and duration of isolation.
39 Using detailed knowledge about the introduction history of 21 non-native populations of the wall
40 lizard *Podarcis muralis* in England we show greater loss of genetic diversity (estimated from
41 microsatellite loci) in older populations and in populations from native regions of high diversity. Loss
42 of genetic diversity was accompanied by higher embryonic mortality in non-native populations,
43 suggesting that introduced populations are sufficiently inbred to jeopardize long-term viability.
44 However, there was no statistical correlation between population-level genetic diversity and average
45 embryonic mortality. Similarly, at the individual level, there was no correlation between female
46 heterozygosity and clutch size, infertility, or hatching success, or between embryo heterozygosity and
47 mortality. We discuss these results in the context of human-mediated introductions and how the
48 history of introductions can play a fundamental role in influencing individual and population fitness in
49 non-native species.

50

51

52

53

54

55

56

57

58

59

60 **Introduction**

61 During the process of colonization, populations may experience dramatic changes in genetic diversity
62 due to founder and bottleneck events (Sakai *et al.* 2001; Dlugosch & Parker 2008). Such reduction in
63 genetic diversity can affect establishment success, population growth and adaptive potential (Nei *et al.*
64 1975; Lee 2002; Dlugosch *et al.* 2015). For example, a small population size increases the probability
65 of inbreeding, which increases homozygosity and could lead to the expression of deleterious recessive
66 mutations that reduce individual fitness (i.e., inbreeding depression) and population viability (Keller &
67 Waller 2002; Charlesworth & Willis 2009). Establishing predictors of genetic diversity and its
68 relationship to estimates of individual and population viability is therefore fundamental to our
69 understanding of what promotes (or hinders) biological invasions and natural range expansion (Lee
70 2002; Keller & Taylor 2008; Excoffier *et al.* 2009; Bock *et al.* 2015; Dlugosch *et al.* 2015), insights
71 that can ultimately assist in conservation management (Frankham *et al.* 2014).

72 Despite the importance of understanding the links between the demographic and ecological
73 processes that reduce genetic diversity and lead to inbreeding depression, establishing these links
74 empirically has proven surprisingly difficult. This is largely because the historical record is often poor
75 and replication of colonization events limited, making it difficult to test for predictors of loss of
76 genetic variation (Estoup & Guillemaud 2010; Uller & Leimu 2011). Generating good evidence for
77 loss of fitness can also be problematic since inbred individuals may die at an early stage in
78 development, making inbreeding depression cryptic or mistakenly classified as parental infertility
79 (Hemmings *et al.* 2012). Indeed, some of the best examples that inbreeding depression (e.g. increased
80 hatching failure) is associated with the severity of bottlenecks (Briskie & Mackintosh 2004; Heber &
81 Briskie 2010) come from hole nesting passerines where early mortality or infertility can be determined
82 with some accuracy (Bensch *et al.* 1994; Kempnaers *et al.* 1996; Spottiswoode & Moller 2004). Also,
83 selection against inbred juveniles might reduce the evidence of inbreeding depression in adults (Keller
84 & Waller 2002). Nevertheless, estimating inbreeding in natural populations is not trivial and data
85 linking introduction history, loss of genetic diversity and inbreeding depression are therefore scarce in

86 other vertebrates. As a result, the extent to which loss of genetic variation and inbreeding depression
87 negatively impact persistence of wild populations remains debatable (Bouzat 2010).

88 The common wall lizard, *Podarcis muralis* provides an opportunity to study how introduction
89 history shapes genetic diversity and how well estimates of genetic diversity correlate with signs of
90 inbreeding depression. Native to southern and western Europe, the species has been repeatedly
91 introduced to England, Germany and North America (Allan *et al.* 2006; Burke & Deichsel 2008;
92 Schulte *et al.* 2012; Michaelides *et al.* 2013). In England, more than 30 extant populations were the
93 result of escapees and deliberate release of captive animals and/or their offspring (Uller and While,
94 unpublished; Lever 1977; Michaelides *et al.* 2013; Michaelides *et al.* 2015). A comprehensive analysis
95 of the colonization history of 23 non-native populations in England revealed nine independent
96 introduction events from two native geographic regions (France and Italy), with evidence of multiple
97 introductions, secondary introductions (i.e., the source was an already established population in
98 England) and admixture (presence of mtDNA haplotypes of more than one lineage; Michaelides *et al.*
99 2013, Michaelides *et al.* 2015). Using 1546 native and non-native animals we test whether genetic
100 diversity (measured using microsatellite markers) of non-native populations was shaped by their
101 geographic and genetic origin, and introduction history (primary vs. secondary and single vs. multiple
102 introductions, admixture, year of introduction and propagule size). Furthermore, for 11 native and 13
103 non-native populations we also collected data on female fecundity, infertility and embryonic mortality
104 to test if loss of genetic diversity and individual heterozygosity was associated with loss of fitness.

105

106 **Material and methods**

107 *Sampling and molecular laboratory work*

108 We used 1318 genotypes from Michaelides *et al.* (2015) and sampled 11 additional populations
109 (228 individuals) from native locations in Italy and France (Figure 1, see also tables S1 and S2 in
110 supplementary information). We extracted genomic DNA from tail tissue preserved in ethanol (70-
111 90%) with DNeasy 96 plate kit (Qiagen, Valencia, CA) following the manufacturer's instructions
112 (with overnight lysis) and genotyped all individuals at 16 microsatellite loci (Richard *et al.* 2012;

113 Heathcote *et al.* 2014). The selected microsatellite set included markers that were developed using
114 individuals from the two focal lineages and geographic regions (France and Italy). This ensured
115 reliable and accurate estimation of genetic diversity (Queiros *et al.* 2015). Multiplexed polymerase
116 chain reactions (PCRs) were carried out in a total volume of 11µl reaction mix containing 1µl of
117 genomic DNA, 5µl of Qiagen MasterMix, 0.2µl of each primer (forward and reverse, [from 10mM](#)
118 [working stock](#)) and 3.8µl (for multiplex 1,2,3 and 5) or 3.6µl (for multiplex 4) of PCR grade dH₂O.
119 PCR conditions were as follows: 15min of initialization step at 95°C, 26 cycles of 30sec at 94°C, 90sec
120 at 57°C (for multiplexes 1 - 3) or 55°C (for multiplexes 4, 5) and 1min at 72°C and a final extension
121 step of 20min at 60°C. The 5'-end of each forward primer was labeled with a fluorescent dye either 6-
122 FAM, HEX or NED. PCR products were run with an internal ladder (red ROX-500), on an ABI 3130
123 genetic analyser (Applied Biosystems Inc.). We scored alleles in GENEIOUS 6.1.7 and any ambiguous
124 peaks (peaks with low relative fluorescence unit) were repeated (PCR and genotyping) to confirm
125 genotype.

126

127 *Microsatellite analyses*

128 We used MICROCHECKER V.2.2.3 (Van Oosterhout *et al.* 2004) to check for null-alleles, large
129 allele dropouts and scoring errors and FSTAT (Goudet 1995, 2001) to calculate deviations from Hardy-
130 Weinberg equilibrium (at the 0.05 nominal level for multiple tests using sequential Bonferroni
131 corrections). We excluded three loci due to very limited amplification in some populations (i.e.,
132 lineage specific loci). Therefore for all subsequent analyses we used 13 microsatellite loci. We
133 calculated observed (H_O) and unbiased expected heterozygosity (H_E) using GENALEX v.6.0 (Peakall &
134 Smouse 2012), allelic richness (A_R , corrected for sample size) using FSTAT (Goudet 1995, 2001) and
135 genetic differentiation among populations (F_{ST}) and linearized F_{ST} ($F_{ST}/(1-F_{ST})$) in ARLEQUIN 3.5.1.3
136 (Excoffier & Lischer 2010).

137

138

139

140 *Genetic diversity in the native and non-native range*

141 To determine how gene flow (or in the case of non-native populations, their introduction history)
142 and genetic drift have influenced population genetic structure within the native and non-native ranges
143 we analyzed the correlation between geographical distance and genetic differentiation (linearized F_{ST})
144 using Mantel tests with 9999 permutations using the *ade4* package in R v.3.1.2 (R Development Core
145 Team 2015). We assessed the structure of genetic variation in the two ranges by hierarchical analysis
146 of molecular variance (AMOVA, Excoffier *et al.* 1992) in ARLEQUIN 3.5.1.3 (Excoffier & Lischer
147 2010). We used two-way ANOVA to assess the effects of geographic range (native vs. non-native)
148 and genetic origin (Italian vs. French) on genetic diversity (H_E and A_R). To improve normality of data,
149 we arcsine-square root transformed H_E and square transformed A_R . We further used Tukey's posthoc
150 tests in R v.3.1.2 (R Development Core Team 2015) to identify significant pairwise comparisons
151 between groups (native Italian, native French, non-native Italian, and non-native French).

152

153 *Predictors of genetic diversity in the non-native range*

154 We used a GLM with Gaussian distribution on transformed data to test if genetic origin (Italian
155 vs. French) and introduction history explained variation in genetic diversity in non-native populations.
156 We included the mode of introduction (primary vs. secondary), number of years since introduction (or
157 first observed) and admixture (presence of mtDNA haplotypes of more than one lineage; yes vs. no) as
158 our variables describing introduction history (Michaelides *et al.* 2015). We also tested for the effects
159 of propagule size (founder size) on genetic diversity of the subset of non-native populations for which
160 this was documented or established with high certainty from interviews with, or written accounts by,
161 those involved in the introductions (supplementary table S1; see also Michaelides *et al.* 2013;
162 Michaelides *et al.* 2015).

163

164 *Fecundity, infertility, and embryonic mortality*

165 We caught 413 gravid females from 11 native and 13 non-native populations during the field
166 seasons 2010-2014 (supplementary table S3). Females were housed in individual cages (590 x 390 x

167 415 mm) at the facilities in Oxford following our standard protocol (see While et al. 2015). We
168 collected the first clutch of the season (from a mating while still in the wild) to generate data on
169 fecundity (C_S , clutch size), infertility (I_N , proportion of infertile eggs) and hatching failure (H_F , the
170 proportion of fertile eggs within a clutch where the embryo died before full term). Infertile eggs can
171 easily be identified on the basis of the lack of egg shell (Olsson & Shine 1997). All other eggs had
172 normal calcified egg shells. Eggs that failed to hatch or that did not show heart beat (using a heart rate
173 monitor; Buddy, Avitronics, England) were dissected to confirm the presence of a dead embryo. We
174 did not attempt to score the exact developmental stage, but mortality typically happened before or
175 soon after oviposition (based on the embryonic staging table in Dufaure & Hubert 1961).

176 We assessed the effects of geographic range (native vs. non-native) and genetic origin (Italian vs.
177 French) on fecundity using a linear mixed model with range, origin and their interaction as a fixed
178 effect, and population as a random effect. Infertility and hatching failure were analyzed using
179 generalized linear mixed models (GLMMs) with the same predictors, adding female identity as a
180 random effect, and a binomial error distribution with logit link function. The statistical analysis was
181 carried out using the *nlme* and *lme4* packages (Bates et al. 2014; Pinheiro et al. 2015) in R v.3.1.2 (R
182 Development Core Team 2015) and significant pairwise comparison between groups (native Italian,
183 native French, non-native Italian, non-native French) was assessed using Tukey posthoc tests. In
184 addition, for non-native populations we used a GLM with Gaussian distribution on transformed data
185 (arcsine square root) to test if population average infertility and hatching failure in populations can be
186 explained by their introduction history. We included genetic origin (Italian vs. French), the mode of
187 introduction (primary vs. secondary), number of years since introduction (or first observed) and
188 admixture (presence of haplotypes of more than one lineage; yes vs. no).

189

190 *Heterozygosity – fitness correlations (HFCs)*

191 Because loss of genetic diversity is associated with inbreeding which in turn reduces reproductive
192 fitness, a correlation is expected between heterozygosity and fitness-related traits (Reed & Frankham
193 2003). We assessed the relationship between expected heterozygosity and average clutch size (C_S),

194 infertility (I_N) and hatching failure (H_F) among non-native populations. Populations with fewer than 10
195 females with complete data on C_S , I_N and H_F were excluded from this analysis to minimize biased
196 estimates of averages.

197 At the individual level, heterozygosity-fitness correlations (HFC) are statistical associations
198 between individual multilocus heterozygosity and fitness traits. HFCs are expected to arise when there
199 is within population variation in inbreeding, heterozygosity and non-genetic component of trait
200 variance (Szulkin *et al.* 2010). Because spurious HFCs can arise when individuals are sampled from
201 different localities or geographic origins (e.g HFCs can be an artefact of between population variation,
202 Slate *et al.* 2004) and since some non-native populations have shown to share demographic history and
203 genetic composition (Michaelides *et al.* 2015) we used STRUCTURE (Pritchard *et al.* 2000) to assign
204 individuals (females) into demes (K), representing clusters of populations that share close genetic
205 relationships (e.g., because one was established through introduction of individuals from another;
206 Michaelides *et al.* 2015). We ran simulations with a burn-in of 10^5 iterations and a run length of 10^6
207 iterations from $K = 1$ to $K=11$ (for native females) or $K=13$ (for non-native females). Runs for each K
208 were replicated five times and the best K was determined according to the method described by
209 Evanno *et al.* (2005) in the online software Structure Harvester (Earl & vonHoldt 2011). Multiple runs
210 were combined in CLUMPP (Jakobsson & Rosenberg 2007) and each female was assigned into a deme
211 when the proportion of membership (q) for a deme was ≥ 0.9 . Structure results identified high posterior
212 probability at $K=2$ for native females (DemeNativeItalian and DemeNativeFrench) and $K=4$ for non-
213 native females (four demes with females belonging to populations of either Italian-only or French-only
214 populations; DemeIntroITA-A (BS, DL, PO, WS), DemeIntroITA-B (WW, SH), DemeIntroITA-C
215 (VT, VB, SW) and DemeIntroFRA (BU, CW, EP, WE). There was one deme that included females of
216 mixed ancestry ($0.1 < q < 0.9$); DemeIntroMix (BS, DL, SH, SW, VB, WE, WS); see Table S3 for list of
217 populations and their abbreviations). Therefore, for subsequent analyses we partitioned our data
218 accordingly to determine whether the presence and/or magnitude of HFC varied among the different
219 partitions (demes).

220 We estimated individual multilocus heterozygosity by calculating the uncorrected homozygosity
221 index (HO, proportion of homozygous loci) and the corrected homozygosity by locus index (HL,
222 weights the contribution of each locus to the homozygosity index depending on allelic variability) in
223 CERNICALIN (Aparicio *et al.* 2006). We performed these calculations separately in each deme
224 (DemeNativeItalian, DemeNativeFrench, DemeIntroItalianA-C, DemeIntroFrench and
225 DemeIntroMix). Since both indices were highly correlated we only report results for HL (see Results).

226 Identity disequilibrium (ID, a correlation in heterozygosity and/or homozygosity across loci
227 (Weir & Cockerham 1973)) is considered a fundamental cause of HFC (Szulkin *et al.* 2010). We
228 therefore estimated ID and its significance using the parameter g_2 (David *et al.* 2007). HFC emerge
229 from variance in individual inbreeding and should only exist if $g_2 > 0$ (Szulkin *et al.* 2010), therefore we
230 assessed the significance of departure from zero based on 1000 permutations in RMES (David *et al.*
231 2007) for each deme.

232 We analyzed the effects of female heterozygosity (F_{HL}) on clutch size (C_s) and hatching failure
233 (H_f) within each deme, and for each fitness trait separately (we did not perform the corresponding
234 analysis on infertility due to the comparably low incidence of infertile eggs). We used Poisson
235 generalized linear models on C_s and binomial GLMMs on H_f including F_{HL} as fixed effect and female
236 ID as a random effect (to control for overdispersion; Bolker *et al.* 2009). We converted the results of
237 each HFC analysis to r , the equivalent of the Pearson product moment correlation coefficient, which is
238 a common measurement of effect size (Nakagawa & Cuthill 2007). We used the z -values from each
239 model to calculate r which was subsequently transformed into Z_r (Fisher's transformation) as
240 described in Coltman and Slate (2003). Since we used HL (homozygosity by locus) for the HFC
241 estimates, we reverse the sign of the effect to match results from published meta-analyses (e.g.
242 Chapman *et al.* 2009). We then used univariate analyses and calculated the average effect size across
243 fitness traits (all effect sizes treated as independent data) and the average effect sizes for each fitness
244 trait separately.

245 Finally, because non-native populations of Italian origin were found to have lost genetic diversity
246 and have increased hatching failure (see Results) we used a subset of females from non-native

247 populations of Italian ancestry to test whether high offspring homozygosity was associated with
248 embryonic mortality. For this analysis we used 31 females and clutches that had at least one embryo
249 that hatched and one that died early. Embryos (dead and alive) were genotyped at 13 microsatellite
250 loci and the homozygosity indices were also calculated in CERNICALIN (Aparicio *et al.* 2006). We then
251 fitted a GLMM with offspring heterozygosity (O_{HL}), femaleID as a random effect and a binomial error
252 distribution with logit link function. P-values were obtained by LRTs of the full model with O_{HL}
253 against the model without O_{HL} . The statistical analyses were carried out in R v.3.1.2 (R Development
254 Core Team 2015) using the *lme4* package (Bates *et al.* 2014).

255

256 **Results**

257 In the native range, there was a clear spatial genetic structure with the Italian region showing
258 higher levels of genetic diversity (H_E and A_R) compared to the French (post hoc Tukey test $p < 0.05$,
259 Figure 3). Across the whole data set most of the variation was found within populations with only 10-
260 15% of variation between ranges and origins (Table 1). Significant isolation-by-distance patterns were
261 observed within both the native and non-native populations (Mantel tests, $p < 0.05$, Figure 2).

262 Genetic diversity (expressed as H_E and A_R) was substantially lower in the non-native populations
263 of Italian origin compared to their native range, whereas non-native populations of French origin only
264 showed a weak loss of diversity compared to their native range (post-hoc Tukey tests between French
265 native and French non-native being statistically significant only for A_R ; Table 2 and Figure 3A and
266 3B). The number of years since introduction was the only statistically significant predictor of genetic
267 diversity for H_E (this was not significant for A_R ; Table 3), with older populations having lower genetic
268 diversity. In the subset of populations for which we had data on propagule size, we found a
269 significantly positive correlation between the number of founders and genetic diversity for H_E ($R =$
270 0.85 , $p = 0.01$, Figure 4) with borderline statistical significance for A_R ($R = 0.74$, $p = 0.058$, Figure 4).

271 Females from non-native populations had significantly larger clutches than females from native
272 populations ($F_{1,411} = 6.17$, $p = 0.02$, Figure 3D). Infertility was low overall and the incidence of
273 infertility did not differ significantly between ranges and origins (range: $Z_{1,409} = -1.07$, $p = 0.29$;

274 origin: $Z_{1,409} = -0.57$, $p = 0.57$). In contrast, hatching failure was affected by the interaction between
275 range and origin ($Z = -3.88$, $p < 0.001$), with significantly higher hatching failure in non-native
276 populations of Italian origin than in their native counterparts (post-hoc Tukey test $p < 0.05$, Table 4,
277 Figure 3C). Within the non-native range, none of the predictors (region of origin, admixture, mode of
278 introduction and years since introduction) significantly affected population average hatching failure or
279 fertility (Supplementary Table S4).

280 Population average expected heterozygosity (H_E) in non-native populations was not significantly
281 correlated with clutch size or hatching failure, but populations with higher heterozygosity had
282 significantly lower incidence of infertility (Figure S1). At the individual level, HFCs are expected to
283 arise from variance in inbreeding, measured with the g_2 statistic, within the various partitions
284 identified by Structure (at $K = 2$; DemeNativeItalian, DemeNativeFrench and at $K = 4$;
285 DemeIntroItalianA-C, DemeIntroFrench and DemeIntroMix). We found positive values for all demes
286 except one (DemeIntroFrench) but statistically significant values only for the DemeIntroItalian-B ($g_2 =$
287 0.067 , $p = 0.04$, see also Supplementary Table S5). Generalized Linear Mixed Models of HFCs
288 indicated no significant association between female heterozygosity (F_{HL}) and fitness traits (H_F , C_S) in
289 any of the data partitions (Supplementary Table S6). The overall average effect size on all demes
290 combined was low ($\check{Z}r = 0.039$) and the 95% confidence interval included zero (Supplementary Table
291 S6). Finally, within clutches, embryos that died before hatching were no more homozygous than their
292 successfully hatched siblings ($\chi^2 = 0.01$, $p = 0.91$; Supplementary Table S7).

293

294 **Discussion**

295 Marginal populations, such as non-native populations, are often founded by a small number of
296 animals, have restricted gene flow and, as a consequence, may have low genetic diversity and suffer
297 from inbreeding depression. Our analyses of non-native wall lizard populations in England showed
298 loss of genetic diversity and an increase in embryonic mortality compared to native populations.
299 Despite this, we failed to establish individual-level correlations between heterozygosity and various
300 measures of fitness.

301 During and following the colonisation of a new area, populations are expected to lose genetic
302 variation and display increased differentiation amongst populations due to founder effects,
303 bottlenecks, and genetic drift (Nei *et al.* 1975; Dlugosch & Parker 2008). As predicted, we found a
304 consistent loss of genetic diversity in non-native compared to native populations. Interestingly, non-
305 native populations from the native region with higher genetic diversity have lost proportionally more
306 genetic variation. This could imply that bottlenecks may have been more severe for non-native Italian
307 populations, but it may also reflect a sampling effect or perhaps an extinction threshold that eliminates
308 populations with lower diversity, making the diversity in extant non-native populations of French and
309 Italian origin similar in magnitude. The lineages diverged from each other approximately 2-3 MYA
310 (Gassert *et al.* 2013; Michaelides *et al.* 2013) and the higher genetic diversity in Italy compared to
311 France likely reflects historical processes that periodically separated populations in refugia. In
312 particular, there appears to have been multiple refugia within Italy, leading to contemporary zones of
313 secondary contact following range expansion in the region of Italy from which the UK populations
314 originated (Giovannotti *et al.* 2010; Gassert *et al.* 2013; Salvi *et al.* 2013). Consequently, our study
315 emphasizes how the phylogeographic structure in the native range may shape patterns of genetic
316 diversity in the non-native range (Taylor & Keller 2007).

317 Propagule size is the most consistent predictor of genetic diversity in introduced populations
318 (Dlugosch & Parker 2008; Simberloff 2009; Uller & Leimu 2011; Blackburn *et al.* 2015). This was
319 confirmed in our study where, despite that information regarding the number of founders was only
320 available for seven populations, diversity increased significantly with the number of animals released.
321 Older populations also harbored less genetic variation than more recently established populations. This
322 may reflect a prolonged period of isolation and absence of gene flow. It is also possible that natural
323 selection contributes to loss of diversity given the evidence that populations established several
324 decades ago (approximately ten to forty generations) have adapted to the colder climate in the UK
325 (While *et al.* 2015). In contrast there was no evidence for further reduction in diversity in secondary
326 introductions. A loss of genetic variation is expected to be a characteristic of sequential founder events
327 (Clegg *et al.* 2002), but our results are not unique for lizards. Successive colonization of *Hemidactylus*

328 *mabouia* in Florida (US), via human-mediated dispersal, did not result in further loss of genetic
329 diversity (Short & Petren 2011). Secondary introductions from admixed populations may explain this
330 pattern (e.g. Tonione *et al.* 2011) as genetic admixture is common in biological invasions and can
331 increase genetic diversity (Kolbe *et al.* 2004; Genton *et al.* 2005; Kolbe *et al.* 2007; Facon *et al.* 2008)
332 sometimes creating novel combinations of alleles in the new range (Ellstrand & Schierenbeck 2000).
333 However, in our study there was no evidence that multiple introductions and admixture, occurring
334 from genetically (and phenotypically) differentiated lineages in the native range, had higher overall
335 nuclear genetic diversity. We can conclude that non-native wall lizard populations are less genetically
336 diverse on average, but that populations have retained variation through secondary introductions and
337 not gained much variation through admixture, at least with respect to neutral markers.

338 Small population size should result in mating between close relatives, which may cause
339 inbreeding depression (Keller & Waller 2002). Hatching failure is a common outcome of inbreeding
340 depression in captive birds and reptiles (Bensch *et al.* 1994), and has been directly attributed to loss of
341 genetic variation in wild birds (Briskie & Mackintosh 2004; Heber & Briskie 2010; Hemmings *et al.*
342 2012). In our study, non-native populations of Italian origin showed high hatching failure, reaching
343 over 30% in some populations, compared to both their native counterparts (mean ca 7%) and non-
344 native populations of French origin (10%). Because eggs were incubated at constant temperatures in
345 the laboratory and hence environmental conditions were standardized across clutches, these effects are
346 likely to be due to expression of deleterious recessives. A high hatching failure in non-native
347 populations of Italian origin is consistent with the greater reduction in genetic diversity relative to the
348 native range compared to French populations. This may suggest that populations of Italian origin have
349 experienced stronger bottlenecks events (although the low sample size for French populations suggests
350 the difference between lineages needs to be treated with caution). Indeed, the severity of the
351 bottleneck has been shown to significantly influence the degree of hatching failure in birds (Briskie &
352 Mackintosh 2004; Heber & Briskie 2010). It is worth noting that the high levels of early mortality are
353 consistent between sampling years and hence likely to reflect a significant genetic load in non-native
354 populations.

355 An approach to quantify the effects of genetic erosion on fitness is to estimate correlations
356 between molecular variation and fitness (or fitness-related) traits among and within populations
357 (Szulkin *et al.* 2010). Heterozygosity-fitness correlations (HFCs) at the population level reveal
358 “ambient inbreeding” shared by all members of the population which is due to fixation of deleterious
359 alleles (fixation load). In a meta-analysis, Reed and Frankham (2003) showed that 19% of the
360 variation in fitness among populations was a result of significant correlations between molecular
361 variation and population fitness. In our study, only one of the non-native demes of shared ancestry
362 showed statistically significant identity disequilibrium (ID, the correlation in heterozygosity and/or
363 homozygosity across loci; Weir & Cockerham 1973; Szulkin *et al.* 2010). It is therefore perhaps not
364 surprising that, despite a reasonable sample size relative to other published studies (Chapman *et al.*
365 2009), we did not find a statistically significant correlation between population genetic diversity and
366 average clutch size or hatching failure among non-native populations. The average effect sizes across
367 demes also suggested that the true effect size is close to zero. Also within clutches, we failed to detect
368 any differences in heterozygosity between embryos that died early in development and their
369 successfully hatched siblings. However, populations with low genetic diversity had increased
370 incidence of infertility, although the absolute levels of infertility were still low (less than 8% of eggs)
371 compared to the high incidence of embryonic mortality.

372 It is unclear why the effect was stronger for infertility than for embryo mortality, but it could
373 reflect that inbreeding depression primarily affects sperm production or sperm viability in males.
374 Indeed, inbreeding depression is often manifested in low sperm viability in captivity (Asa *et al.* 2007),
375 and has been demonstrated in wild populations of rabbits (Gage *et al.* 2006). Recent evidence for male
376 effects on offspring through epigenetic modifications of sperm (e.g., Lambrot *et al.* 2013; Radford *et*
377 *al.* 2014) also raises the possibility that inbred males may produce sperm with compromised genomic
378 or epigenomic stability, which may contribute to early mortality. In addition, mating only with close
379 relatives could result in infertility if fertilization success is lower for genetically similar males, as has
380 been demonstrated in sand lizards (Olsson *et al.* 1996). Further studies of sperm production, sperm

381 viability and post-copulatory discrimination of males in native and non-native populations are needed
382 to test these hypotheses.

383 How can we reconcile the consistent loss of genetic diversity and increased hatching failure in
384 non-native populations with the lack of a bivariate relationship between individual-level
385 heterozygosity and hatching failure? Although there are many known examples of individual
386 multilocus heterozygosity and fitness correlations (reviewed in Chapman *et al.* 2009) effects are
387 relatively weak and effect sizes generally small. If effects are strongest in males, we may not be able
388 to detect HFC by focusing on females even if there is substantial evidence for inbreeding depression,
389 as suggested by the high incidence ~~of of infertility and~~ embryonic mortality in some non-native
390 populations. It is also possible that some populations with low heterozygosity have undergone purging
391 of deleterious mutations (e.g Pujol *et al.* 2009; Facon *et al.* 2011). This would imply that not all
392 populations or individuals with low heterozygosity should show high incidence of inbreeding
393 depression. However, the efficiency of purging depends on many genetic and demographic factors
394 (Keller & Waller 2002) and the time necessary to lessen inbreeding depression could be highly
395 variable (Chapman *et al.* 2009). Finally, our study was restricted to 13 microsatellite markers.
396 Significant HFCs have been reported with fewer markers (e.g., Chapman *et al.* 2009; Brommer *et al.*
397 2015; Velando *et al.* 2015), but neutral markers used might not be sufficient to capture HFCs
398 adequately (Balloux *et al.* 2004; Miller & Coltman 2014), especially as g_2 values suggested a moderate
399 level of inbreeding at most. Thus, our failure to detect ID and/or HFC's should not be taken as
400 evidence that inbreeding depression is absent (Kardos *et al.* 2014). Using a large number of markers
401 such as single nucleotide polymorphisms (e.g. Miller *et al.* 2014; Huisman *et al.* 2016) and/or analysis
402 of functional genes such as genes of the Major-Histocompatibility Complex (e.g. Agudo *et al.* 2012)
403 may be more appropriate when estimated genome-wide heterozygosity and the effect on fitness. The
404 large number of independent introductions of wall lizards to England would provide a good study
405 system to explore how consistent these measures of genetic variation correlate with introduction
406 history and loss of fitness due to inbreeding.

407 In conclusion, the levels of genetic diversity in non-native populations of *P. muralis* reflect their
408 origin and phylogeographic structuring in the native range, with greater loss of diversity in non-native
409 populations from native regions with high genetic variation. Older populations and populations
410 founded by a low number of individuals had lower genetic diversity. Embryonic mortality was high in
411 non-native populations of Italian origin. Although this is consistent with the greater loss of genetic
412 diversity for Italian-origin populations, we found no evidence that heterozygosity across microsatellite
413 markers is significantly correlated with inbreeding depression at the population or individual levels.

414

415 **Acknowledgements**

416 We are grateful to Steve Langham, Charles Snell, Martin Noble, Fred and Pat Howarth, Shona McDonough, the
417 Lever family, Tony Pashley, Mark Anderson, Ian Boyd, Nick Squirrel, Tim Bernhard, Tanya French, and
418 Anthony Mitchell for outstanding help with locating UK populations and providing access to private gardens and
419 lands. Lindall Kidd and Hannah MacGregor assisted with field work and Mary Magorrian with labwork. We also
420 thank three anonymous reviewers for their helpful comments and suggestions that improved our manuscript. We
421 are grateful to the Royal Society of London, the British Ecological Society, and the National Geographic Society
422 for project funding (all to TU). SNM was supported by Biotechnology and Biological Sciences Research Council
423 (BBSRC) scholarship and an A.G Leventis Foundation Grant. GMW was supported by an FP7 Marie Curie post-
424 doctoral fellowship. TU was supported by the Royal Society of London and the Knut and Alice Wallenberg
425 Foundations. The research was approved by the UK Home Office Ethical License PPL30/56. All work and
426 procedures during fieldwork were carried out under annual licenses and permits from Natural England
427 (20091978; 20102163; 20112817), Direction Régionale de l'Environnement, de l'Aménagement et du Logement
428 (No 2010/DDEA/SEPR/175, No 2010-11, No 11/2012, No 2010-DDEA-SE-105, No 29/2012, No
429 11/DDTM/657-SERN-NB, No SE-2010-24), Ministero dell'Ambiente e della Tutela del Territorio del Mare –
430 DG Protezione della Natura e del Mare (prot. PNM-2012-2738, prot. 0011511/PNM, prot. PNM-2012-3878,
431 ISRA prot. 14392, 2764/PNM,) and Societas Herpetologica Italica (prot. ISPRA 9139 T/-A31).

432

433

434

435

436 **References**

- 437 Agudo R, Carrete M, Alcaide M, *et al.* (2012) Genetic diversity at neutral and adaptive loci determines
438 individual fitness in a long-lived territorial bird. *Proc Biol Sci* **279**, 3241-3249.
- 439 Allan GM, Prelypchan CJ, Gregory PT (2006) Population profile of an introduced species, the common wall
440 lizard (*Podarcis muralis*), on Vancouver Island, Canada. *Canadian Journal of Zoology-Revue*
441 *Canadienne De Zoologie* **84**, 51-57.
- 442 Aparicio JM, Ortego J, Cordero PJ (2006) What should we weigh to estimate heterozygosity, alleles or loci?
443 *Molecular Ecology* **15**, 4659-4665.
- 444 Asa C, Miller P, Agnew M, *et al.* (2007) Relationship of inbreeding with sperm quality and reproductive success
445 in Mexican gray wolves. *Animal Conservation* **10**, 326-331.
- 446 Balloux F, Amos W, Coulson T (2004) Does heterozygosity estimate inbreeding in real populations? *Molecular*
447 *Ecology* **13**, 3021-3031.
- 448 Bates D, Maechler M, Bolker B, Walker S (2014) Fitting Linear Mixed-effects models using lme4. In: *ArXiv e-*
449 *prints*.
- 450 Bensch S, Hasselquist D, Torbjorn von S (1994) Genetic Similarity between Parents Predicts Hatching Failure:
451 Nonincestuous Inbreeding in the Great Reed Warbler? *Evolution* **48**, 317-326.
- 452 Blackburn TM, Lockwood JL, Cassey P (2015) The influence of numbers on invasion success. *Molecular*
453 *Ecology* **24**, 1942-1953.
- 454 Bock DG, Caseys C, Cousens RD, *et al.* (2015) What we still don't know about invasion genetics. *Molecular*
455 *Ecology* **24**, 2277-2297.
- 456 Bolker BM, Brooks ME, Clark CJ, *et al.* (2009) Generalized linear mixed models: a practical guide for ecology
457 and evolution. *Trends Ecol Evol* **24**, 127-135.
- 458 Bouzat JL (2010) Conservation genetics of population bottlenecks: the role of chance, selection, and history.
459 *Conservation Genetics* **11**, 463-478.
- 460 Briskie JV, Mackintosh M (2004) Hatching failure increases with severity of population bottlenecks in birds.
461 *Proceedings of the National Academy of Sciences of the United States of America* **101**, 558-561.
- 462 Brommer JE, Kekkonen J, Wikstrom M (2015) Using heterozygosity-fitness correlations to study inbreeding
463 depression in an isolated population of white-tailed deer founded by few individuals. *Ecol Evol* **5**, 357-
464 367.
- 465 Burke R, Deichsel G (2008) Lacertid lizard introductions into North America: history and future. In: *Urban*
466 *herpetology* (eds. J.C. Mitchell, Brown REJ, Bartholomew B), pp. 347-353. Society for the Study of
467 Amphibians and Reptiles, Salt Lake City, UT.
- 468 Chapman JR, Nakagawa S, Coltman DW, Slate J, Sheldon BC (2009) A quantitative review of heterozygosity-
469 fitness correlations in animal populations. *Molecular Ecology* **18**, 2746-2765.
- 470 Charlesworth D, Willis JH (2009) The genetics of inbreeding depression. *Nat Rev Genet* **10**, 783-796.
- 471 Clegg SM, Degnan SM, Kikkawa J, *et al.* (2002) Genetic consequences of sequential founder events by an
472 island-colonizing bird. *Proc Natl Acad Sci U S A* **99**, 8127-8132.
- 473 Coltman D, Slate J (2003) Microsatellite measures of inbreeding: A meta-analysis. *Evolution* **57**, 971-983.
- 474 David P, Pujol B, Viard F, Castella V, Goudet J (2007) Reliable selfing rate estimates from imperfect population
475 genetic data. *Molecular Ecology* **16**, 2474-2487.
- 476 Dlugosch KM, Anderson SR, Braasch J, Cang FA, Gillette HD (2015) The devil is in the details: genetic
477 variation in introduced populations and its contributions to invasion. *Molecular Ecology* **24**, 2095-2111.
- 478 Dlugosch KM, Parker IM (2008) Founding events in species invasions: genetic variation, adaptive evolution,
479 and the role of multiple introductions. *Molecular Ecology* **17**, 431-449.
- 480 Dufaure J, Hubert J (1961) *Table de développement du lézard vivipare: Lacerta (Monotoca) vivipara Jacquin.*
- 481 Earl DA, vonHoldt BM (2011) STRUCTURE HARVESTER: a website and program for visualizing
482 STRUCTURE output and implementing the Evanno method. *Conservation Genetics Resources* **4**, 359-
483 361.
- 484 Estoup A, Guillemaud T (2010) Reconstructing routes of invasion using genetic data: why, how and so what?
485 *Molecular Ecology* **19**, 4113-4130.
- 486 Evanno G, Regnaut S, Goudet J (2005) Detecting the number of clusters of individuals using the software
487 STRUCTURE: a simulation study. *Molecular Ecology* **14**, 2611-2620.
- 488 Excoffier L, Foll M, Petit RJ (2009) Genetic Consequences of Range Expansions. *Annual Review of Ecology,*
489 *Evolution, and Systematics* **40**, 481-501.
- 490 Excoffier L, Lischer HEL (2010) Arlequin suite ver 3.5: a new series of programs to perform population genetics
491 analyses under Linux and Windows. *Molecular Ecology Resources* **10**, 564-567.

492 Excoffier L, Smouse PE, Quattro JM (1992) Analysis of molecular variance inferred from metric distances
493 among DNA haplotypes: application to human mitochondrial DNA restriction data. *Genetics* **131**, 479-
494 491.

495 Facon B, Huffbauer RA, Tayeh A, *et al.* (2011) Inbreeding depression is purged in the invasive insect *Harmonia*
496 *axyridis*. *Curr Biol* **21**, 424-427.

497 Facon B, Pointier J-P, Jarne P, Sarda V, David P (2008) High genetic variance in life-history strategies within
498 invasive populations by way of multiple introductions. *Current biology : CB* **18**, 363-367.

499 Frankham R, Bradshaw CJA, Brook BW (2014) Genetics in conservation management: Revised
500 recommendations for the 50/500 rules, Red List criteria and population viability analyses. *Biological*
501 *Conservation* **170**, 56-63.

502 Gage MJ, Surridge AK, Tomkins JL, *et al.* (2006) Reduced heterozygosity depresses sperm quality in wild
503 rabbits, *Oryctolagus cuniculus*. *Curr Biol* **16**, 612-617.

504 Gassert F, Schulte U, Husemann M, *et al.* (2013) From southern refugia to the northern range margin: genetic
505 population structure of the common wall lizard, *Podarcis muralis*. *Journal of Biogeography* **40**, 1475-
506 1489.

507 Genton BJ, Shykoff JA, Giraud T (2005) High genetic diversity in French invasive populations of common
508 ragweed, *Ambrosia artemisiifolia*, as a result of multiple sources of introduction. *Molecular Ecology* **14**,
509 4275-4285.

510 Giovannotti M, Nisi-Cerioni P, Caputo V (2010) Mitochondrial DNA sequence analysis reveals multiple
511 Pleistocene glacial refugia for *Podarcis muralis* (Laurenti, 1768) in the Italian Peninsula. *Italian Journal*
512 *of Zoology* **77**, 277-288.

513 Goudet J (1995) FSTAT (Version 1.2): A computer program to calculate F-statistics. *Heredity* **86**, 485-486.

514 Goudet J (2001) FSTAT, a program to estimate and test gene diversities and fixation indices (version 2.9.3).
515 Available from <http://www.unil.ch/izea/software/fstat.html>.

516 Heathcote RJP, Dawson DA, Uller T (2014) Characterisation of nine European wall lizard (*Podarcis muralis*)
517 microsatellite loci of utility across sub-species. *Conservation Genetics Resources* **7**, 85-87.

518 Heber S, Briskie JV (2010) Population bottlenecks and increased hatching failure in endangered birds.
519 *Conservation Biology* **24**, 1674-1678.

520 Hemmings NL, Slate J, Birkhead TR (2012) Inbreeding causes early death in a passerine bird. *Nat Commun* **3**,
521 863.

522 Huisman J, Kruuk LEB, Ellis PA, Clutton-Brock T, Pemberton JM (2016) Inbreeding depression across the
523 lifespan in a wild mammal population. *Proceedings of the National Academy of Sciences*
524 10.1073/pnas.1518046113, 201518046.

525 Jakobsson M, Rosenberg NA (2007) CLUMPP: a cluster matching and permutation program for dealing with
526 label switching and multimodality in analysis of population structure. *Bioinformatics* **23**, 1801-1806.

527 Kardos M, Allendorf FW, Luikart G (2014) Evaluating the role of inbreeding depression in heterozygosity-
528 fitness correlations: how useful are tests for identity disequilibrium? *Mol Ecol Resour* **14**, 519-530.

529 Keller LF, Waller DM (2002) Inbreeding effects in wild populations. *Trends in Ecology and Evolution* **17**, 230-
530 241.

531 Keller SR, Taylor DR (2008) History, chance and adaptation during biological invasion: separating stochastic
532 phenotypic evolution from response to selection. *Ecology Letters* **11**, 852-866.

533 Kempenaers B, Adriaensen F, Noordwijk AJV, Dhondt AA (1996) Genetic Similarity, Inbreeding and Hatching
534 Failure in Blue Tits: Are Unhatched Eggs Infertile? *Proceedings: Biological Sciences* **263**, 179-185.

535 Kolbe JJ, Glor RE, Rodríguez Schettino L, *et al.* (2004) Genetic variation increases during biological invasion
536 by a Cuban lizard. *Nature* **431**, 177-181.

537 Kolbe JJ, Glor RE, Schettino LR, *et al.* (2007) Multiple sources, admixture, and genetic variation in introduced
538 anolis lizard populations. *Conservation Biology* **21**, 1612-1625.

539 Lambrot R, Xu C, Saint-Phar S, *et al.* (2013) Low paternal dietary folate alters the mouse sperm epigenome and
540 is associated with negative pregnancy outcomes. *Nat Commun* **4**, 2889.

541 Lee CE (2002) Evolutionary genetics of invasive species. *Trends in Ecology and Evolution* **17**, 386-391.

542 Lever C (1977) *The Naturalized Animals of the British Isles* Hutchinson, London, UK.

543 Michaelides S, While G, Bell C, Uller T (2013) Human introductions create opportunities for intra-specific
544 hybridization in an alien lizard. *Biological Invasions* **15**, 1101-1112.

545 Michaelides SN, While GM, Zajac N, Uller T (2015) Widespread primary, but geographically restricted
546 secondary, human introductions of wall lizards, *Podarcis muralis*. *Molecular Ecology* **24**, 2702-2714.

547 Miller JM, Coltman DW (2014) Assessment of identity disequilibrium and its relation to empirical
548 heterozygosity fitness correlations: a meta-analysis. *Molecular Ecology* **23**, 1899-1909.

549 Miller JM, Malenfant RM, David P, *et al.* (2014) Estimating genome-wide heterozygosity: effects of
550 demographic history and marker type. *Heredity (Edinb)* **112**, 240-247.

551 Nakagawa S, Cuthill IC (2007) Effect size, confidence interval and statistical significance: a practical guide for
552 biologists. *Biol Rev Camb Philos Soc* **82**, 591-605.

553 Nei M, Maruyama T, Chakraborty R (1975) The Bottleneck Effect and Genetic Variability in Populations.
554 *Evolution* **29**, 1.

555 Olsson M, Shine R (1997) The seasonal timing of oviposition in sand lizards (*Lacerta agilis*): why early clutches
556 are better. *Journal of Evolutionary Biology* **10**, 369-381.

557 Olsson M, Shine R, Madsen T, Gullberg A, Tegelström H (1996) Sperm selection by females. *Nature* **383**, 585-
558 585.

559 Peakall R, Smouse PE (2012) GenA1Ex 6.5: genetic analysis in Excel. Population genetic software for teaching
560 and research--an update. *Bioinformatics* **28**, 2537-2539.

561 Pinheiro J, Bates D, DebRoy S, Sarkar D, Team RC (2015) nlme: Linear and Nonlinear Mixed Effects Models

562 Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using multilocus genotype data.
563 *Genetics* **155**, 945-959.

564 Pujol B, Zhou SR, Sanchez Vilas J, Pannell JR (2009) Reduced inbreeding depression after species range
565 expansion. *Proceedings of the National Academy of Sciences* **106**, 15379-15383.

566 Queiros J, Godinho R, Lopes S, *et al.* (2015) Effect of microsatellite selection on individual and population
567 genetic inferences: an empirical study using cross-specific and species-specific amplifications. *Mol*
568 *Ecol Resour* **15**, 747-760.

569 R Development Core Team (2015) R: A language and environment for statistical computing. R Foundation for
570 Statistical Computing, Vienna, Austria.

571 Radford EJ, Ito M, Shi H, *et al.* (2014) In utero effects. In utero undernourishment perturbs the adult sperm
572 methylome and intergenerational metabolism. *Science* **345**, 1255903.

573 Reed DH, Frankham R (2003) Correlation between fitness and genetic diversity *Conservation Biology* **17**, 230-
574 237.

575 Richard M, Stevens VM, Hénanff ML, Coulon A (2012) Fourteen new polymorphic microsatellite loci for the
576 wall lizard *Podarcis muralis* (Sauria : Lacertidae). *Molecular Ecology Resources*, 1-5.

577 Sakai AK, Allendorf FW, Holt JS, *et al.* (2001) The Population Biology of Invasive Species. *Annual Review of*
578 *Ecology and Systematics* **32**, 305-332.

579 Salvi D, Harris DJ, Kaliontzopoulou A, Carretero Ma, Pinho C (2013) Persistence across Pleistocene ice ages in
580 Mediterranean and extra-Mediterranean refugia: phylogeographic insights from the common wall
581 lizard. *BMC Evolutionary Biology* **13**, 147.

582 Schulte U, Hochkirch A, Loetters S, *et al.* (2012) Cryptic niche conservatism among evolutionary lineages of an
583 invasive lizard. *Global Ecology and Biogeography* **21**, 198-211.

584 Short KH, Petren K (2011) Multimodal dispersal during the range expansion of the tropical house gecko
585 *Hemidactylus mabouia*. *Ecol Evol* **1**, 181-190.

586 Simberloff D (2009) The Role of Propagule Pressure in Biological Invasions. *Annual Review of Ecology,*
587 *Evolution, and Systematics* **40**, 81-102.

588 Slate J, David P, Dodds KG, *et al.* (2004) Understanding the relationship between the inbreeding coefficient and
589 multilocus heterozygosity: theoretical expectations and empirical data. *Heredity (Edinb)* **93**, 255-265.

590 Spottiswoode C, Moller AP (2004) Genetic similarity and hatching success in birds. *Proc Biol Sci* **271**, 267-272.

591 Szulkin M, Bierne N, David P (2010) Heterozygosity-fitness correlations: a time for reappraisal. *Evolution* **64**,
592 1202-1217.

593 Taylor DR, Keller SR (2007) Historical range expansion determines the phylogenetic diversity introduced during
594 contemporary species invasion. *Evolution* **61**, 334-345.

595 Tonione MA, Reeder N, Moritz CC (2011) High genetic diversity despite the potential for stepping-stone
596 colonizations in an invasive species of gecko on Moorea, French Polynesia. *PLoS One* **6**, e26874.

597 Uller T, Leimu R (2011) Founder events predict changes in genetic diversity during human-mediated range
598 expansions. *Global Change Biology* **17**, 3478-3485.

599 Van Oosterhout C, Hutchinson WF, Wills DPM, Shipley P (2004) Micro-Checker: Software for Identifying and
600 Correcting Genotyping Errors in Microsatellite Data. *Molecular Ecology Notes* **4**, 535-538.

601 Velando A, Barros A, Moran P (2015) Heterozygosity-fitness correlations in a declining seabird population.
602 *Molecular Ecology* **24**, 1007-1018.

603 Weir BS, Cockerham CC (1973) Mixed self and random mating at two loci. *Genetics Research* **21**, 247-262.

604 While GM, Williamson J, Prescott G, *et al.* (2015) Adaptive responses to cool climate promotes persistence of a
605 non-native lizard. *Proceedings of the Royal Society B* **282**, 20142638.

606 **Data accessibility**

607 Sampling locations and genetic diversity data: Table S1 and S2 in supplementary information.
 608 Population average fitness trait data: Table S3 in supplementary information.
 609 Genotypes of individuals used in the genetics analyses are deposited in Dryad (XXX to be completed
 610 upon acceptance).
 611

612 **Author Contributions**

613 SNM, GMW, and TU conceived of the project, collected data, and wrote the manuscript. SNM
 614 generated and analysed the genetic data with help of NZ, and MALZ, RS, BC and FA collected
 615 samples from native populations.

616 **Table 1** Analysis of Molecular Variance (AMOVA) in the native and non-native range.

Range	Source of variation	<i>d.f.</i>	Sum of squares	Percentage of variation
Native range	Among groups (Italy – France)	1	568.14	10.6
	Among populations within groups	40	971.45	7.65
	Within populations	1940	8705.85	81.75
	<i>Total</i>	1981	10245.44	
Non-native range	Among groups (Italy – France)	1	332.87	14.55
	Among populations within groups	18	805.03	15.82
	Within populations	926	3533.18	69.63
	<i>Total</i>	945	4671.09	

617 **Table 2** GLM results for predictors of genetic diversity (expected heterozygosity, H_E , and allelic richness, A_R).
 618 $H_E \sim \text{Range} * \text{Origin}$

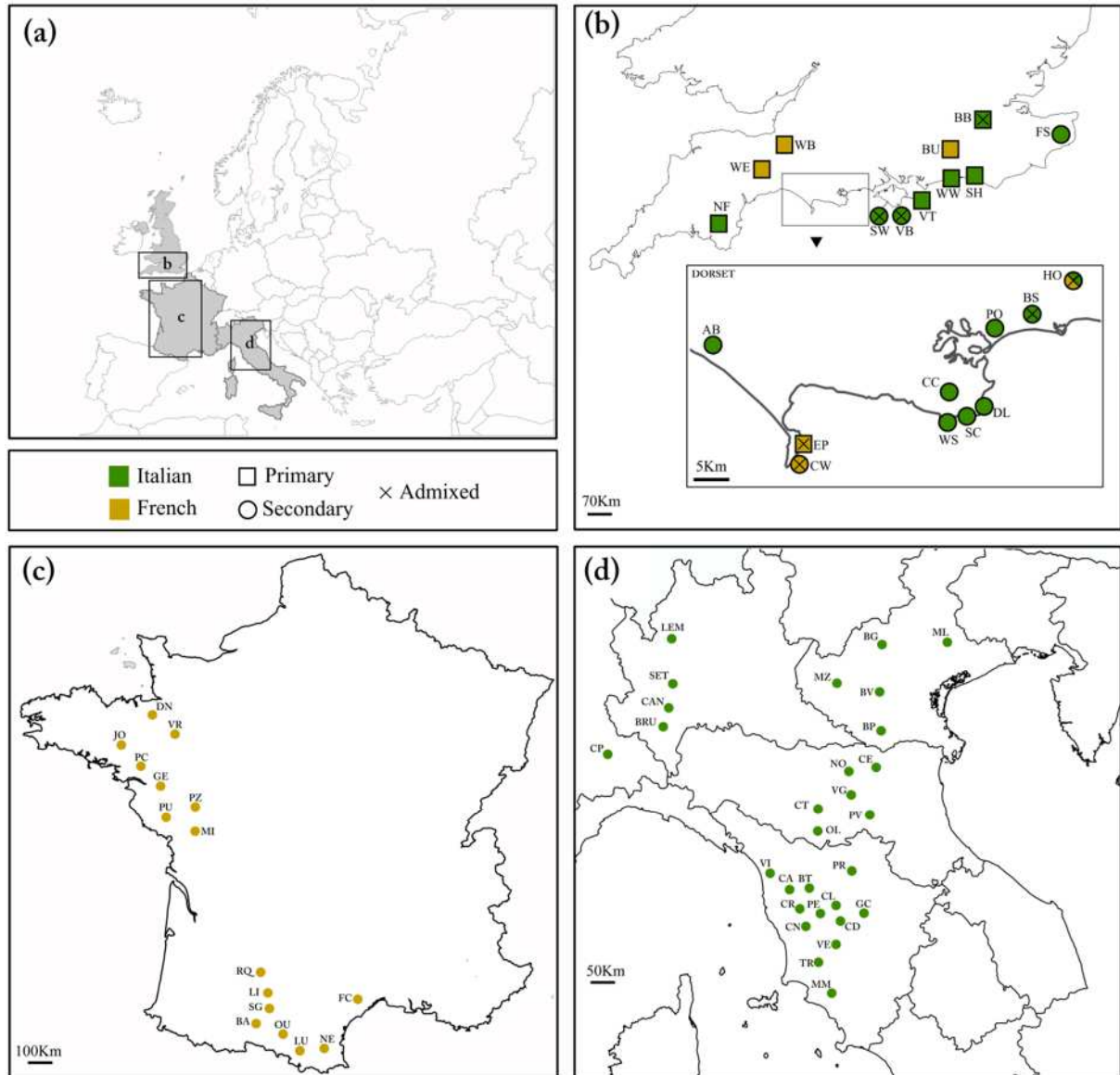
Source of variation	<i>d.f.</i>	<i>F</i>	<i>P</i>
Range (Native – Non-native)	1,61	77.32	< 0.001
Origin (Italy – France)	1,61	27.04	< 0.001
Range : Origin	1,61	11.44	< 0.001
$A_R \sim \text{Range} * \text{Origin}$			
Range (Native – Non-native)	1,61	177.95	< 0.001
Origin (Italy – France)	1,61	71.90	< 0.001
Range : Origin	1,61	24.53	< 0.001

619 **Table 3** GLM results for the predictors of genetic diversity (expected heterozygosity (H_E) and allelic richness (A_R)) in the
 620 non-native range. Statistically significant p-values are in bold.
 621 $H_E \sim \text{origin} + \text{mode of introduction} + \text{admixture} + \text{years}$

Variable	<i>d.f.</i>	<i>F</i>	<i>p</i>
Origin (Italy – France)	1,19	0.13	0.72
Mode of introduction (Primary – Secondary)	1,19	1.29	0.27
Admixture (Yes – No)	1,19	0.01	0.92
Years	1,19	5.75	0.03
$A_R \sim \text{origin} + \text{mode of introduction} + \text{admixture} + \text{years}$			
Variable	<i>d.f.</i>	<i>F</i>	<i>p</i>
Origin (Italy – France)	1,19	0.21	0.64
Mode of introduction (Primary – Secondary)	1,19	0.43	0.52
Admixture (Yes – No)	1,19	0.03	0.85
Years	1,19	3.18	0.09

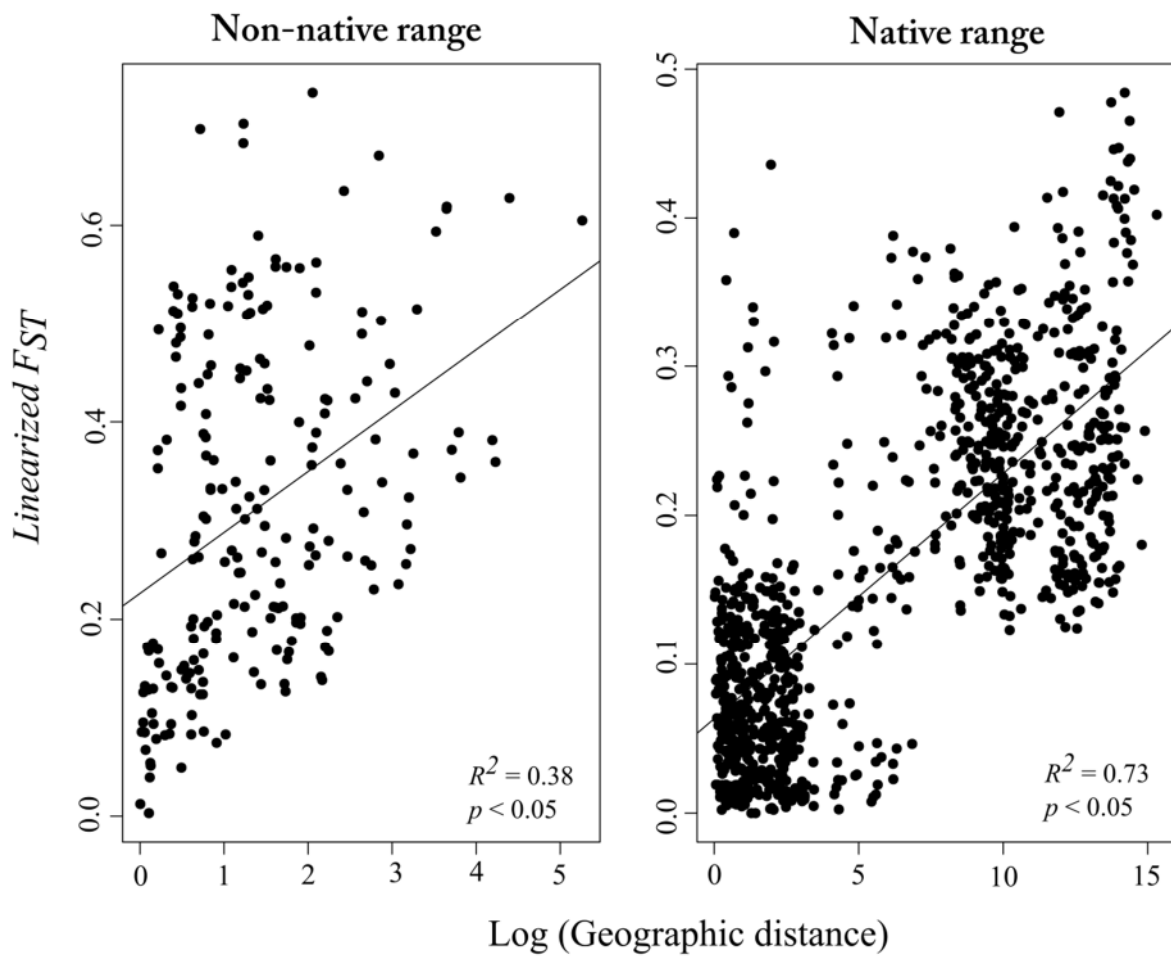
622 **Table 4** GLMM results assessing the effects of range and genetic origin on hatching failure. Statistically significant p-values
 623 are in bold.
 624

Variable	Parameter estimate (SE)	<i>p</i>	Random effects	Variance	SD
Range (Native – Non-native)	1.3187 (0.7825)	0.09	Population	0	0
Origin (Italian – French)	2.2596 (0.4866)	> 0.001	FemaleID	9.827	3.135
Origin : Range	-4.0069 (0.9536)	> 0.001			



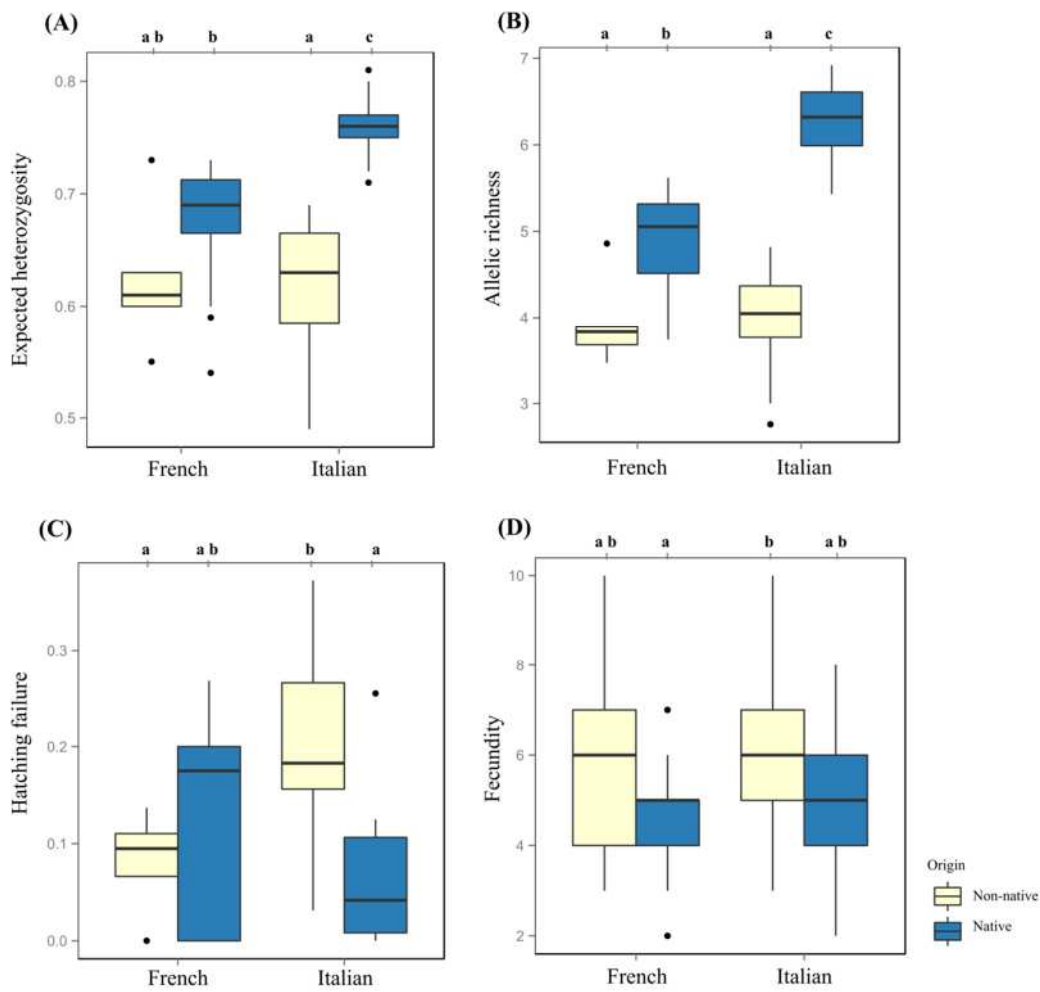
625
 626
 627
 628

Figure 1 Distribution of sampling locations in the native and non-native range. Populations in England are coded based on their introduction history (Italian or French genetic origin, primary or secondary introduction and whether there was evidence of admixture; presence of mtDNA haplotypes from two or more lineages). Map modified from Michaelides *et al.* 2015.



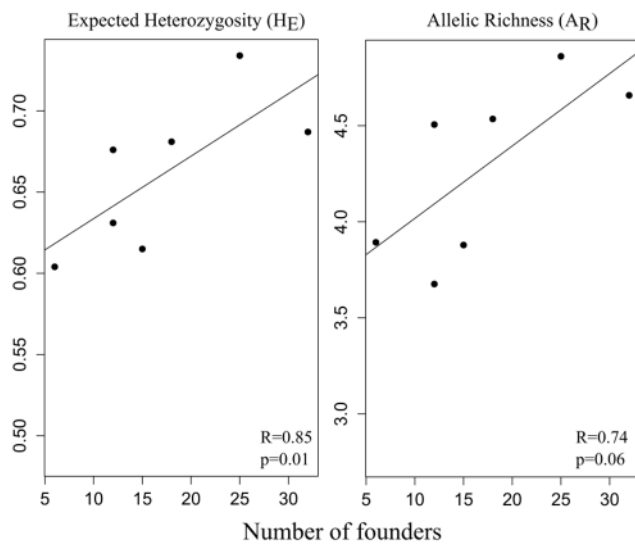
629
 630
 631
 632
 633

Figure 2 Correlation between genetic (Linearized F_{ST}) and geographic distance (log-transformed). There was evidence of isolation-by-distance in both the non-native and native range as assessed by Mantel tests (after 9999 permutations). Note different scales on the axes for the two plots.



634
 635 **Figure 3** Genetic diversity and fitness related traits (hatching failure, fecundity (clutch size)) in native and non-native
 636 populations of French and Italian ancestry. (A) Expected heterozygosity; (B) Allelic richness; (C) Hatching failure; (D)
 637 Fecundity. Different letters above the plots indicate significantly different pairwise comparisons assessed by Tukey posthoc
 638 tests (groups sharing the same letter have non-significant differences).

639



640
 641 **Figure 4** Correlation between number of founders and genetic diversity.