

Manuscript Details

Manuscript number	ENVPOL_2019_2421
Title	Adverse effects of beached plastic litter on the establishment of coastal dune vegetation: first experimental evidences
Article type	Research Paper

Abstract

Plastic pollution is a global issue involving all environments, including coastal dunes which are among the most valuable habitats worldwide. Vegetation plays a fundamental role in dune formation and maintenance, and it is subjected to macro-plastic entanglement. Yet the effects of plastics on dune plant communities and dune system stability are unknown. To fill this knowledge gap, we investigated the impact of macro-plastics of non-biodegradable bags and compostable starch-based bags deposited in foredunes on seedling emergence, morphology and biomass production of three ecologically relevant dune species *Thinopyrum junceum*, *Ammophila arenaria* and *Glaucium flavum*. Non-biodegradable bags are a large component of beach litter. Compostable bags have recently been introduced to reduce plastic pollution, but their behavior on natural habitats is poorly known. We hypothesized that bags would affect seedling emergence and development, and that the magnitude of their effects would vary due to different material behavior and species sensitivity. Both virgin bags and bags exposed to beach conditions were used to examine whether weathering would reduce or exacerbate bag effects. Both bag types reduced seedling emergence percentage irrespectively of weathering and species. Virgin MB bags also delayed the emergence of *A. arenaria* seedlings. Six months after emergence, *T. junceum* seedlings from MB and PE bags had lower aboveground biomass, and those from PE bags had also altered root system compared to controls. Few *A. arenaria* seedlings from PE bags survived, and those emerged from MB ones were smaller than controls. No *G. flavum* seedlings survived regardless of treatments. These findings showed that not only traditional bags, but also compostable ones are a threat to foredunes. They also highlight the need of informing people and managers about the impact of the incorrect disposal of compostable bags and of establishing more appropriate beach cleaning practices on dune ecosystem health.

Keywords Coastal environment; dune plants; ecology; plastic pollution; seedlings

Corresponding Author Elena Balestri

Corresponding Author's Institution University of Pisa

Order of Authors Virginia Menicagli, Elena Balestri, Flavia Vallerini, Alberto Castelli, Claudio Lardicci

Suggested reviewers john griffin, Anna Traveset, ELIZABETH LACHTER, Roberto Simonini, Italo Castro

Submission Files Included in this PDF

File Name [File Type]

Cover letter.docx [Cover Letter]

Highlights.docx [Highlights]

Graphical abstract.tif [Graphical Abstract]

Manuscript.docx [Manuscript File]

To view all the submission files, including those not included in the PDF, click on the manuscript title on your EVISE Homepage, then click 'Download zip file'.

Dear Editor,

My co-authors (Virginia Menicagli, Flavia Vallerini, Alberto Castelli and Claudio Lardicci) and I are pleased to submit the enclosed manuscript “Adverse effects of beached plastic litter on the establishment of coastal dune vegetation: first experimental evidences” as Research Article in Environmental Pollution.

The manuscript focuses on plastic litter and its effects on sexual recruitment and growth of dune plants which play fundamental role in structuring and maintaining dune systems. The impact of plastics on dune vegetation has been neglected up to date. Previous studies have investigated the composition and distribution of plastics on beaches and dunes and their effects on biota. However, increasing evidence indicate that plastics can influence sand physical/chemical properties and can release in sand plasticizers and other additives included during manufacture that are toxic to a variety of organism, including plants. In addition, macro-fragments can entangle plant organs and seeds.

Our study is the first that experimentally examined the effects of plastics deposited in foredune on the establishment and growth of plants. Specifically, in an eight-months field seed sowing experiment we investigated the impact of macro-fragments of two types of shopping bags, a non-biodegradable bag (that is common in beach litter) and a compostable starch-based bag (recently marketed to reduce plastic pollution), on seedling emergence, morphology and biomass production of three ecologically relevant species *Thinopyrum junceum*, *Ammophila arenaria* and *Glaucium flavum*. Both virgin bags and bags exposed to beach conditions were used to examine whether weathering would reduce or exacerbate bag effects on plants. The response of plants grown with bags were compared with that of plants grown without bags (controls).

The results show that both types of bags once buried in sand reduced the percentage of seedling emergence in all the investigated species. In addition, they caused morphological and growth alterations both at below- and aboveground level. *Thinopyrum junceum* was more sensitive to non-biodegradable bag while *A. arenaria* was sensitive to both the bag types. In general, weathering

minimized the impact of bags on plants. Overall, the results suggest that plastic litter is an actual threat to coastal dune vegetation and hence to dune stability and ecosystem health.

The manuscript fits the Aims and Scope of the Journal, as it relates for the first time the global issue of environmental pollutants (macro-plastics) to the recruitment success and growth of coastal dune plants, providing new insights into the interaction between plastics and dune environment, as well as into its effect on dune ecosystem health. The results highlight the need of informing people and producers about the impact of their incorrect disposal on natural ecosystems to prevent further dune damage. Management actions, such as those involving the burial of beach wrack in foredunes and the use of mechanical wrack removal during cleaning operations (Poeta et al., 2014) should be avoided in order to preserve the ecological functions and services of coastal dune habitats.

We suggest as potential reviewers:

Prof. Elizabeth Lachter: lachter@iq.ufrj.br

Dr. John Griffin: j.n.griffin@swansea.ac.uk

Prof. Roberto Simonini: roberto.simonini@unimore.it

Prof. Anna Traveset: atraveset@imedea.uib-csic.es

Prof. Italo Castro: italobraga@gmail.com

On behalf of all co-authors,

Yours sincerely,

the corresponding author Elena Balestri

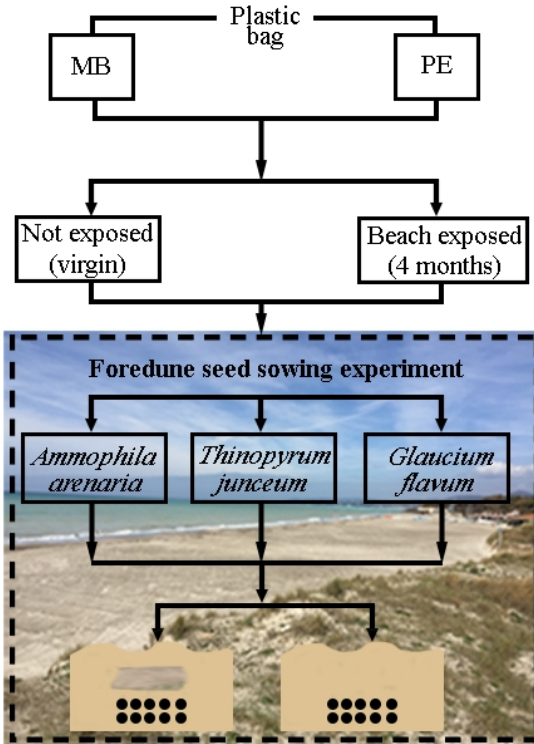
E-mail address: elena.balestri@unipi.it

- Shopping bags are a consistent fraction of plastic litter deposited in beaches.
- Effects of beached bags on dune plant recruitment from seed have been neglected.
- Non-biodegradable and compostable bags reduced the emergence of three dune plants.
- Bags differentially affected seedling growth and morphology of the study species.
- Bag effects could translate in changes in plant communities and dune structure.

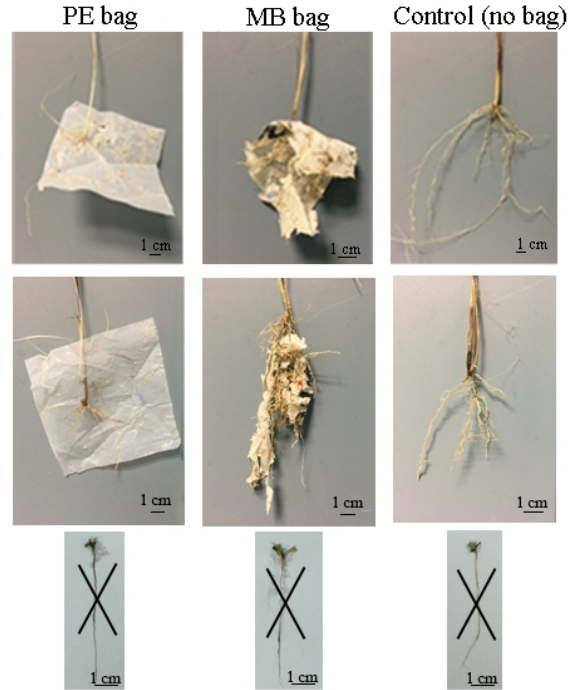
MB: Mater-bi®
 PE: high-density polyethylene

●●●●●: seeds

▭: bag fragment
 (10 x 10 cm)



Seedlings (six months after emergence)



Ammophila arenaria

Seedling emergence (%): PE = MB < C
 Mean emergence time: not exposed bag
 PE = C < MB

Seedling performance: MB ≠ C
 PE: few survived seedlings

Thinopyrum junceum

Seedling emergence (%): PE = MB < C
 Mean emergence time: PE = MB = C

Seedling performance: not exposed bag
 PE = MB ≠ C
 PE: not exposed bag ≠ exposed bag

Glaucium flavum

Seedling emergence (%): PE = MB < C
 No survived seedling

1 **Adverse effects of beached plastic litter on the establishment of coastal dune vegetation: first**
2 **experimental evidences**

3

4 Virginia Menicagli^a, Elena Balestri^{a*}, Flavia Vallerini^a, Alberto Castelli^a, Claudio Lardicci^a

5

6 ^a Department of Biology, University of Pisa, via Derna 1, 56126, Pisa, Italy

7

8 *Corresponding author

9 elena.balestri@unipi.it

10 University of Pisa, via Derna 1, 56126 Pisa, Italy

11

12

13

14

15

16

17

18

19

20

21

22

23

24

25

26

27 **Keywords**

28 Coastal environment; dune plants; ecology; plastic pollution; seedlings

29

30 **Capsule**

31 Plastic litter negatively affect the development of coastal dune vegetation by drastically reducing
32 seedling emergence and growth.

33

34 **Declaration of interest: none**

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53 **1. Introduction**

54

55 Plastic pollution is a global ecological problem affecting all marine and terrestrial environments
56 (Barnes et al., 2009; Lavers and Bond, 2017; Machado et al., 2018a). Due to their location at the
57 interface between the terrestrial and the marine habitat, beaches and sand dunes are final receptors
58 of plastics waste and may represent hot-spots of plastic pollution (Calvão et al., 2013; Poeta et al.,
59 2014; UNEP, 2015; Poeta et al., 2017; Ceccarini et al., 2018; Rangel-Buitrago et al., 2018). Coastal
60 dunes are structured by a relatively low number of plant species adapted to cope with stressful
61 conditions (Maun, 2009), and they are one of the most valuable but threatened habitats worldwide
62 (Martínez et al., 2004; Defeo et al., 2009; Barbier et al., 2011; Doody 2013; Drius et al. 2019). Over
63 the last decade, increasing efforts have been made to examine the composition and abundance of
64 plastics deposited in beaches and dunes (Kordella et al., 2013; Poeta et al., 2014; Lee and Sanders,
65 2015; Munari et al., 2016; Moreira et al., 2016; Alshawafi et al., 2017; Munari et al., 2017; Šilc et
66 al., 2018). Available data indicate that beached plastic fragments can alter physical sediment
67 properties, including temperature and permeability, with potential consequences for beach biota
68 (Carson et al., 2011). However, plastic items accumulated in embryonic and mobile dunes (Poeta et
69 al., 2014; de Francesco et al., 2018) can interact with vegetation through direct mechanisms, such
70 the entanglement of plant organs and seeds (Poeta et al., 2017; Šilc et al., 2018), and indirect
71 mechanisms, including altered sand structure, nutrient mobilization and release of plasticizers and
72 additives (León et al., 2019), similarly to that observed for plastic films used in agriculture
73 (Steinmetz et al., 2016; Machado et al., 2018a; Qi et al., 2018; Rillig et al., 2019). However, no
74 information is available on the actual impact of the deposition of plastics in dunes on plant
75 development.

76 Plastic bags made of conventional non-biodegradable polymers are a common component (17-
77 23%, up to 1 bag per linear meter) of the plastic litter deposited in beaches and dunes (Munari et al.,
78 2016; Alshawafi et al., 2017; Pasternak et al., 2017; Schmuck et al., 2017; Šilc et al., 2018), and

79 they are among the most harmful plastic litter items to marine biota (Hardesty et al., 2015). Most of
80 these bags derives from terrestrial sources via rivers or marine-based activities, but some of them
81 are accidentally or intentionally discarded on beaches by beach-goers (Ryan et al., 2009; Accinelli
82 et al., 2012). Due to the durability and adverse impacts on natural environments and biota, non-
83 biodegradable bags have recently been banned in many countries over the world (Xanthos and
84 Walker, 2017). Biodegradable and compostable bags have been introduced as eco-friendly
85 substitute of conventional ones, but all of these bags are intended to be properly disposed in
86 industrial or home compost facilities at the end of their life (EN 13432, 2000). Increasing evidence
87 indicates that once entered coastal environments these bags degrade slowly and represent a new
88 emerging threat for organisms and habitats (Accinelli et al., 2012; Tosin et al., 2012; Green et al.,
89 2015; Balestri et al., 2017; Nazareth et al., 2018). In addition, recent laboratory studies have shown
90 that when in contact with water, macro-fragments derived from non-biodegradable bags and
91 compostable bags can release plasticizers and additives incorporated during manufacture in the
92 surrounding environment (Bejgarn et al., 2015; Balestri et al., 2019). The leachates generated from
93 these bags have found to be toxic to terrestrial plants such as *Lepidium sativum* L. a species used in
94 standard phytotoxicity tests (Balestri et al., 2019). Since the persistence and the integrity of coastal
95 dunes depend on preserving vegetation (Acosta et al. 2007; Drius et al., 2013), understanding the
96 effects of beached plastic bags on early life history stages that are most critical for dune plants
97 (Balestri and Cinelli, 2004; Maun, 2009; Rajaniemi and Barrett, 2018), it is fundamental and may
98 help to guide more appropriate management actions.

99 The aim of the present study was to evaluate the impact of the deposition of two types of
100 shopping bags, a non-biodegradable bag and a compostable starch-based bag, in foredunes on plant
101 populations and communities. We hypothesized that these bags would affect the emergence and
102 development of dune plant seedlings, but the magnitude of their effects would differ due to different
103 bag composition and behavior in nature, and plant species sensitivity. To test these hypotheses, a
104 field seed sowing experiment was performed using as models three plant species widely distributed

105 along Mediterranean and Atlantic foredunes, *Thinopyrum junceum* (L.) Á. Love, *Ammophila*
106 *arenaria* (L.) Link and *Glaucium flavum* Crantz, that differ in their ecological role (Acosta et al.,
107 2007; van Puijenbroek et al., 2017). The emergence, morphology and biomass production of
108 seedlings in experimentally bag polluted and not polluted plots (controls) were examined eight
109 months after seed sowing. Both new bags and bags exposed to beach conditions were assessed to
110 investigate whether weathering would reduce or exacerbate bag effects on plants.

111

112 **2. Materials and methods**

113

114 2.1. Plant material collection and bag types

115

116 *Thinopyrum junceum* is a monocotyledon rhizomatous pioneer species that contributes to the
117 formation of embryonic shifting dunes (Drius et al., 2016; van Puijenbroek et al., 2017). *Ammophila*
118 *arenaria* is also a monocotyledon rhizomatous species, and it is involved in the formation of mobile
119 shifting dunes (Hesp, 2002; van Puijenbroek et al., 2017). Both these species are clonal and can
120 propagate through sexually reproduction and clonal growth (Harris and Davy, 1986; Del Vecchio et
121 al., 2013; Hilton et al., 2019). *Glaucium flavum* is a dicotyledon non-clonal herb (Scott, 1983)
122 common along coastal dunes (Thanos, 1989; Walmsley and Davy, 1997; Cambrollè et al., 2015).

123 Seeds of *T. junceum*, *A. arenaria* and *G. flavum* were collected from natural plant populations
124 along the dune system of Rosignano Solvay (43°21'06.11''N 10°27'21.64''E - Italy). Seeds of each
125 species were separately placed in clean glass jars and stored outdoor in darkness until the start of
126 the experiment. Before sowing, the viability of seeds was checked through a pressure test (Borza et
127 al., 2007), and non-viable seeds were discarded.

128 Two type of plastic bags were used in this experiment: high-density polyethylene (PE) bag and
129 Mater-bi® bag (MB). High-density polyethylene is one the most abundant polymers found on beach
130 litter (Corcoran et al., 2009; Munari et al., 2017; Ceccarini et al., 2018). Mater-bi® is based on

131 starch and vinyl-alcohol copolymers (Sforzini et al., 2016) and is certified as compostable in
132 accordance with the main European and international standards (EN13432, 2000). This certification
133 guarantees the complete degradation of MB bags in industrial composting and the absence of any
134 phytotoxic effect of compost. However, the behavior and degradation rate of MB bags in dune
135 habitats is unknown. All the bags were purchased from Italian retailers and belong to the same lots
136 used in a previous laboratory experiment on effects of bag leachates on higher plants (Balestri et al.,
137 2019).

138

139 2.2. Set up and design of seed sowing experiment

140

141 To investigate the effect of the two different types of plastic bags on recruitment and seedling
142 development, a seed sowing experiment was set up in a foredune area of Rosignano Solvay. During
143 the experimental period, air temperature varied from 4.7 to 29.8 °C and total precipitation was
144 574.2 mm. Before the experiment (October 2017), half of bags (hereafter referred to as not exposed
145 or virgin bags, NE) was left in laboratory at a temperature of 22 °C (± 1 SE) while the other half
146 (hereafter referred to as exposed bags, E) was placed on the embryo dune for 4 months to allow
147 bags experiencing weathering before their burial in the seed banks. All bags were then cut into
148 fragments of approximately 10 x 10 cm with scissor, a size within the range of that of macro-
149 plastics found on beaches (van Cauwenberghe et al., 2013; Williams et al., 2016; Prevenios et al.,
150 2018). The average weight of MB bag fragments was 0.286 g (± 0.001) and that of PE ones was
151 0.241 g (± 0.001). Bag thickness was about 25 μm .

152 In February 2018, 25 cm x 25 cm plots (tens of centimeters each from other) were randomly
153 established in the study area and the sand within each plot was removed. Ten seeds of one of the
154 selected species were sown at 6-7 cm depth within each plot, which correspond to the depth of
155 viable seeds found in foredunes (Hilton et al., 2019). Then, a fragment of PE or MB bag, either not
156 exposed or beach exposed, was placed on seeds and covered with sand previously sieved through a

157 0.5 mm mesh-sieve to remove extraneous seeds and materials. Additional plots in which seeds were
158 sown and covered with sand as treated ones, but no bag fragment was added over, were also
159 established and used as controls. The experiment thus followed a full factorial design with three
160 factors, plastic bag (fixed, three levels: PE bag, MB bag and no bag or control), bag exposure
161 (fixed, two levels: no exposure and beach exposure) and plant species (fixed, three levels: *T.*
162 *junceum*, *A. arenaria* and *G. flavum*). There were eight replicates per each treatment combination.
163 In total there were 144 plots.

164 Plots were regularly inspected over the experimental period, and the number of new emerged
165 seedlings of each species within each plot was recorded until the end of the emergence period. The
166 total percentages of seedling emergence and mortality were computed for each species and
167 treatment, and the mean emergence time (MET) was calculated as

$$168 \quad t = \frac{\sum_{i=1}^k n_i t_i}{\sum_{i=1}^k n_i},$$

169 where t_i was the time elapsed from seed sowing to the i^{th} day of observation; n_i was the number
170 of newly emerged seedlings in the day i , and k was the day in which the last seedlings emerged.
171 At the end of the experiment (October 2018), plots were excavated, and seedlings were carefully
172 extracted from sand and transported to the laboratory for morphological measurements (number and
173 length of adventitious roots, number of culms and plant height for *T. junceum* and *A. arenaria*, and
174 length of the primary root and rosette diameter for *G. flavum*). The length of adventitious roots was
175 computed by averaging the length of three intact roots of each plant. Plants were then separated into
176 above-and below-ground organs, dried at 70 °C and weighted to determine the above- and
177 belowground biomass. The below- to aboveground biomass ratio was also computed as considered
178 as an indicator of plant investment into allocation of below- or aboveground resources (Poorter
179 et al., 2012, 2015). Since a different number of seedlings were available at the end of the
180 experiment for each species, the analyses of morphological and growth data were conducted on a
181 sample of five seedlings selected at random among those survived for each species. Only very few
182 seedlings of *G. flavum* emerged from sand, and hence MET and morphological and growth data

183 were not included in the analyses. Finally, all bag fragments in plots were extracted from sand,
184 dried and weighted to calculate weight loss relative to initial value, as this variable is considered a
185 qualitative indicator of plastic degradation (Accinelli et al., 2012; Tosin et al., 2012).

186

187 2.2. Statistical analysis

188

189 Separate three-way analyses of variance (ANOVA) were conducted to examine the effects of
190 treatments (plastic bag, exposure and species) on the percentage of emergence, the mean emergence
191 time and percentage of mortality of seedlings. Due to the low number of seedlings of *A. arenaria*
192 emerged from plots in the presence of exposed and not exposed PE bags, as well as in the presence
193 of not exposed MB bags, separate analyses on morphological and growth variables were conducted
194 for each species. Two-way permutational multivariate analysis of variance (PERMANOVA) was
195 performed on morphological and growth data, and biomass ratio of *T. junceum* to investigate the
196 overall effect of treatments on seedling performance. Since significant treatment effects were
197 detected in PERMANOVA, separate two-way ANOVAs were performed on each variable. One-
198 way PERMANOVA was conducted only on morphological data of *A. arenaria* seedlings to
199 examine the effects of the presence of beach exposed MB on seedling performance, as biomass data
200 were highly correlated (> 0.95) with other morphological variables. Since PERMANOVA revealed
201 significant treatment effects, separate one-way ANOVAs were conducted on each seedling variable.
202 Three-ways ANOVA was also performed on the percentage of bag weight loss to compare the
203 effects of treatments (plastic, exposure and species) on the extent of plastic degradation after burial
204 in the experimental seed bank. A Student-Newman-Keuls (SNK) test was applied to perform post
205 hoc comparisons among levels of significant terms.

206 PERMANOVA analyses were carried out on the Euclidean distance of previously normalized
207 data using 9999 permutations of the residuals under a reduced model for *T. junceum* and using
208 unrestricted permutation of raw data for *A. arenaria* (Anderson et al., 2008). Statistically significant

209 terms ($p < 0.05$) were checked for differences in multivariate group dispersion through
210 permutational analysis of multivariate dispersion (PERMDISP). Prior to ANOVA analyses, data
211 were checked for normality and homogeneity of variance using Shapiro-Wilk test and Cochran's *C*
212 test, respectively, and transformed when necessary to meet ANOVA assumptions. Residuals of data
213 of percentage of emergence, number of roots for *T. junceum* and number of culms for *T. junceum*
214 and *A. arenaria* were not normally distributed, but ANOVA analysis was performed anyway as it is
215 quite robust to violation of this assumption (Underwood, 1997). PERMANOVA analyses were
216 conducted using PERMANOVA + for PRIMER 6 statistical software (Anderson et al., 2008;
217 Clarke and Gorley, 2006) while ANOVA analyses by using GMAV version 5.0 for Windows
218 (Underwood and Chapman, 1998).

219

220 **3. Results**

221

222 Seedlings started to emerge from sand about four weeks after seed sowing and ceased to emerge
223 two months later. The presence of a plastic bag fragment over seeds, irrespectively of the type and
224 the exposure of bag, significantly reduced the percentage of emergence of seedlings for all the
225 investigated species (Table 1; Fig. 1). *Thinopyrum junceum* showed a significant higher percentage
226 of seedling emergence compared to the other two species, and *G. flavum* exhibited the lowest values
227 (Table 1; Fig. 1). No *G. flavum* seedling emerged from MB bag plots. No difference in the mean
228 seedling emergence time was detected among treatments for *T. junceum*. Seedlings of *A. arenaria*
229 from plots with not exposed MB bag emerged significantly later compared to those from plots with
230 PE bags and controls (Table 1; Fig. 1). They also emerged later than seedlings of *T. junceum* from
231 seeds exposed to the same MB treatment (Table 1).

232 Multivariate PERMANOVA analysis on all examined variables of *T. junceum* detected a
233 significant effect of the main factors, plastic bag and exposure, and of their interaction (Table 2).
234 ANOVA analysis on the number of roots revealed a significant effect of the interaction between the

235 two main factors (Table 2). Seedlings emerged from MB bag plots, irrespectively of bag
236 weathering, had a similar number of roots to that of controls. Instead, seedlings emerged from
237 virgin PE bags had a significantly lower number of roots than controls while the opposite was found
238 for those grown with beach exposed PE bags (Table 2; Fig. 2). The length of roots of seedling
239 grown with MB bags was similar to that of controls and significantly greater than that of seedlings
240 grown with PE bags, irrespectively of weathering (Table 2; Fig. 2). The presence of both PE or MB
241 bags significantly reduced seedling height and aboveground biomass as compared to controls. No
242 significant difference was observed in the number of culms, belowground biomass and biomass
243 ratio among treatments (Table 2; Fig. 2).

244 Multivariate PERMANOVA analysis on all examined seedling variables of *A. arenaria* detected
245 significant differences between seedlings grown in the presence of beach exposed MB bags and
246 controls (Table 3). Seedlings grown with MB bags had a significantly lower number and shorter
247 roots, and their height was significantly lower than those of controls (Table 3; Fig. 3). In addition,
248 they showed significantly smaller below- and aboveground biomasses compared to controls (Table
249 3; Fig. 3). The number of culms and the biomass ratio were not significantly affected by the
250 investigated factors (Table 3; Fig. 3). No statistically significant effect was detected on percentage
251 of seedling mortality among treatments (Table 1). However, only very few emerged seedlings of *A.*
252 *arenaria* survived with PE bags (Fig. 1). No *G. flavum* seedling survived to the end of the
253 experiment, regardless of treatments. All plastic fragments were still in place and showed holes and
254 breakages on their surface caused by the perforation of emerged seedlings. Irrespectively of
255 weathering and species, MB bags exhibited a significant higher percentage of weight loss relative to
256 their initial value ($14.5 \% \pm 2$) compared to PE bags ($5.2 \% \pm 0.8$; Table 4).

257

258 **4. Discussion**

259

260 Macro-plastics pollution in coastal environments is a serious global issue, and its potential
261 consequences on marine biota, habitats and human health are well described. The present study is
262 the first aimed at assessing the effects of the occurrence of macro-plastics, both not biodegradable
263 and compostable, in coastal sand dunes on the recruitment by seed and the early growth of plants
264 inhabiting these environments.

265 The results of the study partly supported our first hypothesis, i.e. plastic bags would influence
266 seedlings emergence, and that the magnitude of their effects would vary according to different bag
267 behavior and plant species sensitivity. Indeed, for all the examined species the number of seedlings
268 emerged from plots with buried bag macro-fragments, irrespectively of bag type and weathering,
269 was lower compared to that of controls. Since all bag fragments placed over seeds were not
270 degraded at the end of the experimental period, the observed reduction in the percentage of seedling
271 emergence could be due to the mechanical resistance of bag films as previously observed for
272 seedlings of a crop species grown in the presence of plastic fragments (Liu et al., 2014). In fact, the
273 coleoptile (or the hypocotyl) of seedlings had to break and penetrate the plastic film to move
274 upward and emerge from sand. In addition, alterations of the physical/chemical characteristics of
275 the micro-environment, for example oxygen, temperature, moisture, sand texture and gas diffusion
276 rate induced by bag presence might have played a role. The percentage of emergence of *T. junceum*
277 and *A. arenaria* seedlings observed in control plots was relatively high and similar to that reported
278 for these species in other studies (Harris and Davy, 1986; Van der Putten, 1990; Del Vecchio et al.,
279 2013). Instead, the percentage of seedling emergence of *G. flavum* was lower than that previously
280 recorded (Thanos, 1989; Walmsley and Davy, 1997), and the high seedling mortality prevented us
281 to assess the effects of bags on plants. The observed low emergence capability of this species was
282 probably due to the imposed sowing depth, as burial of seeds deeper than 0.5 cm has found to
283 inhibit seed germination of *G. flavum* (Harris and Davy, 1986; Thanos, 1989; Maun, 2009). Our
284 results also showed that MB bags delayed the time of seedling emergence of *A. arenaria* but this
285 effect was consistent only for virgin bags. This suggests that *A. arenaria* seedlings are more

286 sensitive than *T. junceum* ones to chemical compounds migrated from MB bag to the surrounding
287 environment (Lithner et al., 2012; Li et al., 2016; Balestri et al., 2019). Overall, these findings
288 indicate that the burial of MB bags in coastal dunes can adversely affect seedling emergence as well
289 as that of PE bags.

290 Our prediction concerning the differential influence of the two bag types on seedling
291 development in the study species was partly confirmed. Indeed, the performance of *T. junceum*
292 seedlings was influenced by PE bags regardless of weathering while that of *A. arenaria* by exposed
293 MB bags. However, only a limited number of *A. arenaria* seedlings emerged from plots with PE
294 bags and not exposed MB bags, thus this species seemed to be vulnerable to both types of bag.
295 As concerning root development of *T. junceum*, root length decreased regardless of weathering.
296 Instead, root number decreased with virgin bags and increased with exposed ones. This latter
297 response could be due to the mitigation of the inhibitory growth effect of additives released from
298 exposed bags during weathering (Balestri et al. 2019). Indeed, previous studies have shown that
299 these compounds can either inhibit (at high concentration) or stimulate (at low concentration) root
300 development in crops (Staples et al., 2010; Li et al., 2018). Instead, the low number, and the
301 reduced length and biomass of roots observed in *A. arenaria* seedlings grown with exposed MB
302 bags could be explained by changes in soil properties and rhizosphere bacteria community
303 (Bandopadhyay et al., 2018; Machado et al., 2018b; Rilling et al., 2019) beneath bag fragments.
304 Previous studies on crop species have found that soil property changes associated to the presence of
305 biodegradable plastic films (mulches) in soils could be responsible for below- and aboveground
306 biomass reduction (Qi et al., 2018). The production of smaller root systems may not allow plants to
307 efficiently uptake nutrients and water from sand in coastal dunes. This could explain the reduced
308 plant height and aboveground biomass observed in *T. junceum* seedling grown in the presence of
309 both MB and PE bags as well as in *A. arenaria* seedlings with MB bags. These findings are in
310 agreement with results of recent studies on mulches made of biodegradable plastics that have

311 revealed adverse effects of this material on agroecosystem and soil health (Steinmetz et al., 2016;
312 Qi et al., 2018; Sintim et al., 2019).

313

314 **5. Conclusions**

315

316 Results of this study demonstrate that the occurrence of plastic bags in coastal dunes can reduce
317 the success of sexual recruitment and the growth of newly established seedlings, hindering the
318 formation of a well-developed vegetation cover that is critical to maintain dune system integrity.
319 Thus, plastic bags represent a further threat to these important but vulnerable habitats. Importantly,
320 the results of this study indicate that even new generation of bags made of biodegradable and
321 compostable polymers can affect plants similarly to, or even more than, traditional ones made of
322 non-biodegradable polymers. As compostable bags are currently proposed as an environmental
323 solution to plastic pollution and their production will increase in the future (UNEP, 2015), it is
324 essential informing people and producers about the impact of their incorrect disposal on natural
325 ecosystems to prevent further dune damage. Management actions, such as those involving the burial
326 of beach wrack in foredunes and the use of mechanical wrack removal during cleaning operations
327 (Poeta et al., 2014) should be avoided in order to preserve the ecological functions and services of
328 coastal dune habitats.

329

330 **Acknowledgements**

331

332 We sincerely thank Viviana Ligorini and Andrea Torre for their support in the set-up of the field
333 experiment. We also thank the Municipality of Rosignano Marittimo and Solvay Chimica Italia
334 SPA of Rosignano Solvay for providing technical support during the study. This work is part of the
335 PhD research project of Virginia Menicagli.

336

337 **Funding**

338

339 This work was supported by the University of Pisa (PRA and FA).

340

341 **References**

342

343 Acosta, A., Ercole, S., Stanisci, A., Pillar, V.D.P., Blasi, C., 2007. Coastal vegetation zonation and
344 dune morphology in some mediterranean ecosystems. *J. Coast. Res.* 236, 1518–1524.

345 <https://doi.org/10.2112/05-0589.1>

346 Accinelli, C., Saccà, M.L., Mencarelli, M., Vicari, A., 2012. Deterioration of bioplastic carrier bags
347 in the environment and assessment of a new recycling alternative. *Chemosphere.* 89, 136-143.

348 <https://doi.org/10.1016/j.chemosphere.2012.05.028>.

349 Alshawafi, A., Analla, M., Alwashali, E., Aksissou, M., 2017. Assessment of marine debris on the
350 coastal wetland of Martil in the North-East of Morocco. *Mar. Pollut. Bull.* 117, 302–310.

351 <https://doi.org/10.1016/j.marpolbul.2017.01.079>

352 Anderson, M.J., Gorley, R.N., Clarke, K.R., 2008. PERMANOVA+ for PRIMER: Guide to
353 software and statistical methods. PRIMER-E: Plymouth, UK.

354 Balestri, E., Cinelli, F., 2004. Germination and early-seedling establishment capacity of *Pancreatium*
355 *maritimum* L. (Amaryllidaceae) on coastal dunes in the north-western Mediterranean. *J. Coast.*

356 *Res.* 20, 761-770. [https://doi.org/10.2112/1551-5036\(2004\)20\[761:GAEECO\]2.0.CO;2](https://doi.org/10.2112/1551-5036(2004)20[761:GAEECO]2.0.CO;2)

357 Balestri, E., Menicagli, V., Vallerini, F., Lardicci C., 2017. Biodegradable plastic bags on the
358 seafloor: A future threat for seagrass meadows? *Sci. Total Environ.* 605-606, 755-763.

359 <https://doi.org/10.1016/j.scitotenv.2017.06.249>

360 Balestri, E., Menicagli, V., Ligorini, V., Fulignati, S., Raspolli Galletti, A.M., Lardicci, C., 2019.

361 Phytotoxicity assessment of conventional and biodegradable plastic bags using seed

362 germination test. *Ecol. Indic.* 102, 569-580. <https://doi.org/10.1016/j.ecolind.2019.03.005>

363 Bandopadhyay, S., Martin-Closas, L., Pelacho, A.M., DeBruyn, J.M., 2018. Biodegradable plastic
364 mulch films: Impacts on soil microbial communities and ecosystem functions. *Front.*
365 *Microbiol.* 9, 1–7. <https://doi.org/10.3389/fmicb.2018.00819>

366 Barbier, E.B., Hacker, S.D., Kennedy, C., Koch, E.W., Stier, A.C., Silliman, B.R., 2011. The value
367 of estuarine and coastal ecosystem services. *Ecol. Monogr.* 81, 169–193.
368 <https://doi.org/10.1890/10-1510.1>

369 Barnes, D.K.A., Galgani, F., Thompson, R.C., Barlaz, M., 2009. Accumulation and fragmentation
370 of plastic debris in global environments. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* 364, 1985–
371 1998. <https://doi.org/10.1098/rstb.2008.0205>

372 Bejgarn, S., MacLeod, M., Bogdal, C., Breitholtz, M., 2015. Toxicity of leachate from weathering
373 plastics: an exploratory screening study with *Nitocra spinipes*. *Chemosphere.* 132, 114-119.
374 <https://doi.org/10.1016/j.chemosphere.2015.03.010>

375 Borza, J.K., Westerman, P.R., Liebman, M., 2007. Comparing estimates of seed viability in three
376 foxtail (*Setaria*) species using the imbibed seed crush test with and without additional
377 tetrazolium testing. *Weed Technol.* 21, 518–522. <https://doi.org/10.1614/wt-06-110>

378 Calvão, T., Pessoa, M.F., Lidon, F.C., 2013. Impact of human activities on coastal vegetation - A
379 review. *Emirates J. Food Agric.* 25, 926–944. <https://doi.org/10.9755/ejfa.v25i12.16730>

380 Cambrollé, J., Muñoz-Vallés, S., Mancilla-Leytón, J.M., Andrades-Moreno, L., Luque, T.,
381 Figueroa, M.E., 2014. Effects of soil physicochemical properties on plant performance of
382 *Glaucium flavum* Crantz. *Plant Soil.* 386, 185–193. <https://doi.org/10.1007/s11104-014-2258-7>

383 Carson, H.S, Colbert, S.L., Kaylor, M., McDermid, K.J., 2011. Small plastic debris changes water
384 movement and heat transfer through beach sediments. *Mar. Pollut. Bull.* 62, 1708-1713.
385 <https://doi.org/10.1016/j.marpolbul.2011.05.032>

386 Ceccarini, A., Corti, A., Erba, F., Modugno, F., La Nasa, J., Bianchi, S., Castelvetro, V., 2018. The
387 hidden microplastics: new insights and figures from the thorough separation and

388 characterization of microplastics and of their degradation by-products in coastal sediments.
389 Environ. Sci. Technol. 52, 5634-5643. <https://doi.org/10.1021/acs.est.8b01487>

390 Clarke, K.R., Gorley, R.N., 2006. *PRIMER* v6: User manual/tutorial. PRIMER-E: Plymouth.

391 Corcoran, P.L., Biesinger, M.C., Grifi, M., 2009. Plastics and beaches: A degrading relationship.
392 Mar. Pollut. Bull. 58, 80–84. <https://doi.org/10.1016/j.marpolbul.2008.08.022>

393 de Francesco, M.C., Carranza, M.L., Stanisci, A., 2018. Beach litter in Mediterranean coastal
394 dunes: an insight on the Adriatic coast (central Italy). Rend. Lincei. 29, 825–830.
395 <https://doi.org/10.1007/s12210-018-0740-5>

396 Defeo, O., McLachlan, A., Schoeman, D.S., Schlacher, T.A., Dugan, J., Jones, A., Lastra, M.,
397 Scapini, F., 2009. Threats to sandy beach ecosystems: A review. Estuar. Coast. Shelf Sci. 81,
398 1–12. <https://doi.org/10.1016/j.ecss.2008.09.022>

399 Del Vecchio, S., Marbà, N., Acosta, A., Vignolo, C., Traveset, A., 2013. Effects of *Posidonia*
400 *oceanica* beach-cast on germination, growth and nutrient uptake of coastal dune plants. PLoS
401 One 8. <https://doi.org/10.1371/journal.pone.0070607>

402 Doody, J.P., 2013. Sand dune conservation, management and restoration. Coastal Research Library.
403 Springer, Netherlands.

404 Drius, M., Carranza, M.L., Stanisci, A., Jones, L., 2016. The role of Italian coastal dunes as carbon
405 sinks and diversity sources. A multi-service perspective. Appl. Geogr. 75, 127–136.
406 <https://doi.org/10.1016/j.apgeog.2016.08.007>

407 Drius, M., Jones, L., Marzialetti, F., de Francesco, M.C., Stanisci, A., Carranza, M.L., 2019. Not
408 just a sandy beach. The multi-service value of Mediterranean coastal dunes. Sci. Total Environ.
409 668, 1139–1155. <https://doi.org/10.1016/j.scitotenv.2019.02.364>

410 Drius, M., Malavasi, M., Acosta, A.T.R., Ricotta, C., Carranza, M.L., 2013. Boundary-based
411 analysis for the assessment of coastal dune landscape integrity over time. Appl. Geogr. 45, 41–
412 48. <https://doi.org/10.1016/j.apgeog.2013.08.003>

413 EN 13432, 2000. Packaging - Requirements for packaging recoverable through composting and
414 biodegradation - Test scheme and evaluation criteria for the final acceptance of packaging.

415 Green, D.S., Boots, B., Blockley, D.J., Rocha, C., Thompson, R.C., 2015. Impacts of discarded
416 plastic bags on marine assemblages and ecosystem functioning. *Environ. Sci. Technol.* 49,
417 5380-5389. <https://doi.org/10.1021/acs.est.5b00277>

418 Hardesty, B.D., Good, T.P., Wilcox, C., 2015. Novel methods, new results and science-based
419 solutions to tackle marine debris impacts on wildlife. *Ocean Coast. Manag.* 115, 4–9.
420 <https://doi.org/10.1016/j.ocecoaman.2015.04.004>

421 Harris, D., Davy, A.J., 1986. Regenerative potential of *Elymus farctus* from rhizome fragments and
422 seed. *J. Ecol.* 74, 1057-1067. <https://doi.org/10.2307/2260233>

423 Hesp, P., 2002. Foredunes and blowouts: Initiation, geomorphology and dynamics.
424 *Geomorphology.* 48, 245–268. [https://doi.org/10.1016/S0169555X\(02\)00184-8](https://doi.org/10.1016/S0169555X(02)00184-8)

425 Hilton, M., Konlechner, T., McLachlan, K., Lim, D., Lord, J., 2019. Long-lived seed banks of
426 *Ammophila arenaria* prolong dune restoration programs. *J. Coast. Conserv.* 23, 461–471.
427 <https://doi.org/10.1007/s11852-018-0675-0>

428 Kordella, S., Geraga, M., Papatheodorou, G., Fakiris, E., Mitropoulou, I.M., 2013. Litter
429 composition and source contribution for 80 beaches in Greece, Eastern Mediterranean: A
430 nationwide voluntary clean-up campaign. *Aquat. Ecosyst. Heal. Manag.* 16, 111–118.
431 <https://doi.org/10.1080/14634988.2012.759503>

432 Lavers, J.L., Bond, A.L., 2017. Exceptional and rapid accumulation of anthropogenic debris on one
433 of the world’s most remote and pristine islands. *Proc. Natl. Acad. Sci.* 114, 6052–6055.
434 <https://doi.org/10.1073/pnas.1619818114>

435 Lee, R.F., Sanders, D.P., 2015. The amount and accumulation rate of plastic debris on marshes and
436 beaches on the Georgia coast. *Mar. Pollut. Bull.* 91, 113–119.
437 <https://doi.org/10.1016/j.marpolbul.2014.12.019>

438 León, V.M., García-Agüera, I., Moltó, V., Fernández-González, V., Llorca-Pérez, L., Andrade,
439 J.M., Muniategui-Lorenzo, S., Campillo, J.A., 2019. PAHs, pesticides, personal care products
440 and plastic additives in plastic debris from Spanish Mediterranean beaches. *Sci. Total Environ.*
441 670, 672–684. <https://doi.org/10.1016/j.scitotenv.2019.03.216>

442 Li, H.X., Getzinger, G.J., Ferguson, P.L., Orihuela, B., Zhu, M., Rittschof, D., 2016. Effects of
443 toxic leachate from commercial plastics on larval survival and settlement of the barnacle
444 *Amphibalanus amphitrite*. *Environ. Sci. Technol.* 50, 924–931.
445 <https://doi.org/10.1021/acs.est.5b02781>

446 Li, X., Wang, L., Wang, S., Yang, Q., Zhou, Q., Huang, X., 2018. A preliminary analysis of the
447 effects of bisphenol A on the plant root growth via changes in endogenous plant hormones.
448 *Ecotoxicol. Environ. Saf.* 150, 152–158. <https://doi.org/10.1016/j.ecoenv.2017.12.031>

449 Lithner, D., Nordensvan, I., Dave, G., 2012. Comparative acute toxicity of leachates from plastic
450 products made of polypropylene, polyethylene, PVC, acrylonitrile-butadiene-styrene, and
451 epoxy to *Daphnia magna*. *Environ. Sci. Pollut. Res.* 19, 1763–1772.
452 <https://doi.org/10.1007/s11356-011-0663-5>

453 Liu, E.K., He, W.Q., Yan, C.R., 2014. “White revolution” to “white pollution” - agricultural plastic
454 film mulch in China. *Environ. Res. Lett.* 9. <https://doi.org/10.1088/1748-9326/9/9/091001>

455 Machado, A.A., Kloas, W., Zarfl, C., Hempel, S., Rillig, M.C., 2018a. Microplastics as an emerging
456 threat to terrestrial ecosystems. *Glob. Chang. Biol.* 24, 1405–1416.
457 <https://doi.org/10.1111/gcb.14020>

458 Machado, A.A., Lau, C.W., Till, J., Kloas, W., Lehmann, A., Becker, R., Rillig, M.C., 2018b.
459 Impacts of microplastics on the soil biophysical environment. *Environ. Sci. Technol.* 52, 9656–
460 9665. <https://doi.org/10.1021/acs.est.8b02212>

461 Martínez M.L., Psuty, N.P., Lubke, R.A., 2004. A Perspective on Coastal Dunes, in: Martínez,
462 M.L., Psuty, N.P. (Eds.), *Coastal dune - ecology and conservation*. Springer, pp. 3-10.

463 Maun, M. A., 2009. *The biology of coastal sand dunes*. Oxford University Press, New York.

464 Moreira, F.T., Balthazar-Silva, D., Barbosa, L., Turra, A., 2016. Revealing accumulation zones of
465 plastic pellets in sandy beaches. *Environ. Pollut.* 218, 313–321.
466 <https://doi.org/10.1016/j.envpol.2016.07.006>

467 Munari, C., Corbau, C., Simeoni, U., Mistri, M., 2016. Marine litter on Mediterranean shores:
468 Analysis of composition, spatial distribution and sources in north-western Adriatic beaches.
469 *Waste Manag.* 49, 483–490. <https://doi.org/10.1016/j.wasman.2015.12.010>

470 Munari, C., Scoponi, M., Mistri, M., 2017. Plastic debris in the Mediterranean Sea: Types,
471 occurrence and distribution along Adriatic shorelines. *Waste Manag.* 67, 385–391.
472 <https://doi.org/10.1016/j.wasman.2017.05.020>

473 Nazareth, M., Marques, M. R. C., Leite, M. C. A., Braga Castro, I., 2019. Commercial plastics
474 claiming biodegradable status: Is this also accurate for marine environment? *J. Hazard Mater.*
475 366, 714-722. <https://doi.org/10.1016/j.jhazmat.2018.12.052>

476 Pasternak, G., Zviely, D., Ribic, C.A., Ariel, A., Spanier, E., 2017. Sources, composition and spatial
477 distribution of marine debris along the Mediterranean coast of Israel. *Mar. Pollut. Bull.* 114,
478 1036-1045. <https://doi.org/10.1016/j.marpolbul.2016.11.023>

479 Poeta, G., Battisti, C., Acosta, A. T. R., 2014. Marine litter in Mediterranean sandy littoral: spatial
480 distribution patterns along central Italy coastal dunes. *Mar. Pollut. Bull.* 89, 168-173.
481 <http://dx.doi.org/10.1016/j.marpolbul.2014.10.011>

482 Poeta, G., Fanelli, G., Pietrelli, L., Acosta, A.T.R., Battisti, C., 2017. Plasticsphere in action:
483 evidence for an interaction between expanded polystyrene and dunal plants. *Environ. Sci.*
484 *Pollut. Res.* 24, 11856–11859. <https://doi.org/10.1007/s11356-017-8887-7>

485 Poorter, H., Niklas, K.J., Reich, P.B., Oleksyn, J., Poot, P., Mommer, L., 2012. Biomass allocation
486 to leaves, stems and roots: meta-analyses of interspecific variation and environmental control.
487 *New Phytol.* 193, 30–50. <https://doi.org/10.1111/j.1469-8137.2011.03952.x>

488 Poorter, H., Jagodzinski, A.M., Ruiz-Peinado, R., Kuyah, S., Luo, Y., Oleksyn, J., Usoltsev, V.A.,
489 Buckley, T.N., Reich, P.B., Sack, L., 2015. How does biomass distribution change with size

490 and differ among species? An analysis for 1200 plant species from five continents. *New Phytol.*
491 208, 736–749. <https://doi.org/10.1111/nph.13571>

492 Prevenios, M., Zeri, C., Tsangaris, C., Liubartseva, S., Fakiris, E., Papatheodorou, G., 2018. Beach
493 litter dynamics on Mediterranean coasts: distinguishing sources and pathways. *Mar. Pollut.*
494 *Bull.* 129, 448–457. <https://doi.org/10.1016/j.marpolbul.2017.10.013>

495 Qi, Y., Yang, X., Pelaez, A.M., Huerta Lwanga, E., Beriot, N., Gertsen, H., Garbeva, P., Geissen,
496 V., 2018. Macro- and micro- plastics in soil-plant system: Effects of plastic mulch film residues
497 on wheat (*Triticum aestivum*) growth. *Sci. Total Environ.* 645, 1048–1056.
498 <https://doi.org/10.1016/j.scitotenv.2018.07.229>

499 Rajaniemi, T.K., Barrett, D.T., 2018. Germination responses to abiotic stress shape species
500 distributions on coastal dunes. *Plant Ecol.* 219, 1271–1282. [https://doi.org/10.1007/s11258-](https://doi.org/10.1007/s11258-018-0877-4)
501 [018-0877-4](https://doi.org/10.1007/s11258-018-0877-4)

502 Rangel-Buitrago, N., Castro-Barros, J. D., Gracia, A., Villamil Villadiego, J. D., Williams, A. T.,
503 2018. Litter impacts on beach/dune systems along the Atlantico Department, the Caribbean
504 coastline of Colombia. *Mar. Pollut. Bull.* 137, 35-44.
505 <https://doi.org/10.1016/j.marpolbul.2018.10.009>

506 Rillig, M.C., Lehmann, A., de Souza Machado, A.A., Yang, G., 2019. Microplastic effects on
507 plants. *New Phytol.* <https://doi.org/10.1111/nph.15794>

508 Ryan, P.G., Moore, C.J., van Franeker, J.A., Moloney, C.L., 2009. Monitoring the abundance of
509 plastic debris in the marine environment. *Phil. Trans. R. Soc. B.* 364, 1999–2012.
510 <https://doi.org/10.1098/rstb.2008.0207>

511 Schmuck, A.M., Lavers, J.L., Stuckenbrock, S., Sharp, P.B., Bond, A.L., 2017. Geophysical
512 features influence the accumulation of beach debris on Caribbean islands. *Mar. Pollut. Bull.*
513 121, 45–51. <https://doi.org/10.1016/j.marpolbul.2017.05.043>

514 Scott, G.A.M., 1963. *Glaucium flavum* Crantz. *J. Ecol.* 51, 743-754.

515 Sforzini, S., Oliveri, L., Chinaglia, S., Viarengo, A., 2016. Application of biotest for the
516 determination of soil ecotoxicity after exposure to biodegradable plastics. *Front. Env. Sci.* 4:
517 68. <https://doi.org/10.3389/fenvs.2016.00068>

518 Šilc, U., Kuzmic, F., Cakovic, D., Stesevic, D., 2018. Beach litter along various sand dune habitats
519 in the southern Adriatic (E Mediterranean). *Mar. Pollut. Bull.* 128, 353-360.
520 <https://doi.org/10.1016/j.marpolbul.2018.01.045>

521 Sintim, H.Y., Bandopadhyay, s., English, M.E., Bary, A.I., DeBruyn, J.M., Schaeffer, S.M., Miles,
522 C.A., Reganold, J.P., Flury, M., 2019. Impacts of biodegradable plastic mulches on soil health.
523 *Agric. Ecosyst. Environ.* 3, 36–49. <https://doi.org/10.1016/j.agee.2018.12.002>

524 Staples, C., Friederich, U., Hall, T., Klečka, G., Mihaich, E., Ortego, L., Caspers, N., Hentges, S.,
525 2010. Estimating potential risks to terrestrial invertebrates and plants exposed to bisphenol A
526 in soil amended with activated sludge biosolids. *Environ. Toxicol. Chem.* 29, 467-475.
527 <https://doi.org/10.1002/etc.49>

528 Steinmetz, Z., Wollmann, C., Schaefer, M., Buchmann, C., David, J., Tröger, J., Muñoz, K., Frör,
529 O., Schaumann, G.E., 2016. Plastic mulching in agriculture. Trading short-term agronomic
530 benefits for long-term soil degradation? *Sci. Total Environ.* 550, 690–705.
531 <https://doi.org/10.1016/j.scitotenv.2016.01.153>

532 Thanos, C. A., Georghiou, K., Skarou, F., 1989. *Glauclium flavum* seed germination- an
533 ecophysiological approach. *Ann. Bot.* 63, 121-130.
534 <https://doi.org/10.1093/oxfordjournals.aob.a087714>

535 Tosin, M., Weber, M., Siotto, M., Lott, C., Degli Innocenti, F., 2012. Laboratory test methods to
536 determine the degradation of plastics in marine environmental conditions. *Front Microbiol.* 3,
537 225. <https://doi.org/10.3389/fmicb.2012.00225>

538 Underwood, A.J., 1997. *Experiments in ecology: their logical design and interpretation using*
539 *analysis of variance.* Cambridge University press. <https://doi.org/10.1017/CBO9780511806407>

540 Underwood, A. J., Chapman, M. G., 1998. GMAV, Version 5.0 for Windows.

541 UNEP, 2015. Biodegradable Plastics and Marine Litter: Misconceptions, Concerns and Impacts on
542 Marine Environments. United Nations Environment Programme, Nairobi (2015), pp. 2015.

543 van Cauwenberghe, L., Claessens, M., Vandegehuchte, M.B., Mees, J., Janssen, C.R., 2013.
544 Assessment of marine debris on the Belgian Continental Shelf. Mar. Pollut. Bull. 73, 161–169.
545 <https://doi.org/10.1016/j.marpolbul.2013.05.026>

546 van Der Putten, W.H., 1990. Establishment of *Ammophila arenaria* (Marram Grass) from culms,
547 seeds and rhizomes. J. Appl. Ecol. 27, 188-199.

548 van Puijenbroek, M.E.B., Teichmann, C., Meijdam, N., Oliveras, I., Berendse, F., Limpens, J.,
549 2017. Does salt stress constrain spatial distribution of dune building grasses *Ammophila*
550 *arenaria* and *Elytrichia juncea* on the beach? Ecol. Evol. 7, 7290–7303.
551 <https://doi.org/10.1002/ece3.3244>

552 Walmsley, C.A., Davy, A.J., 1997. The restoration of coastal shingle vegetation: effects of substrate
553 composition on the establishment of seedlings. J. Appl. Ecol. 34, 143-153.
554 <http://doi.org/10.2307/2404855>

555 Williams, A.T., Randerson, P., Di Giacomo, C., Anfuso, G., Macias, A., Perales, J.A., 2016.
556 Distribution of beach litter along the coastline of Cádiz, Spain. Mar. Pollut. Bull. 107, 77–87.
557 <http://dx.doi.org/10.1016/j.marpolbul.2016.04.015>

558 Xanthos, D., Walker, T.R., 2017. International policies to reduce plastic marine pollution from
559 single-use plastics (plastic bags and microbeads): A review. Mar. Pollut. Bull. 118, 17–26.
560 <http://dx.doi.org/10.1016/j.marpolbul>

561
562
563
564
565
566

567
568
569
570
571
572
573
574
575
576
577
578
579
580
581
582
583
584
585
586
587
588
589
590
591
592

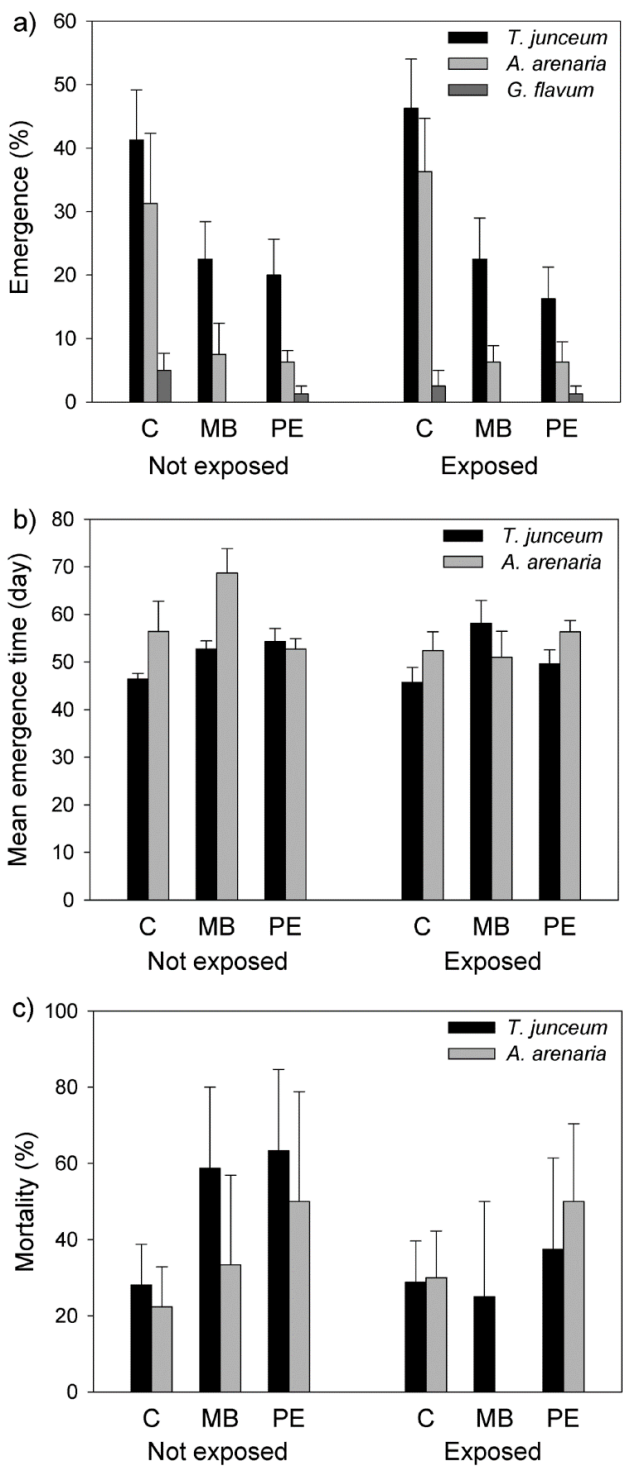


Fig. 1. Percentage of emergence (a), mean emergence time (b) and percentage of mortality (c) of seedlings of the study dune plants emerged from control plots without bags (C) and plots with not exposed and beach exposed Mater-bi® (MB) and high-density polyethylene (PE) bag fragments.

Data are means \pm SE.

Single-column fitting image

593
 594
 595
 596
 597
 598
 599
 600
 601
 602
 603
 604
 605
 606
 607
 608
 609
 610
 611
 612
 613
 614
 615
 616
 617
 618

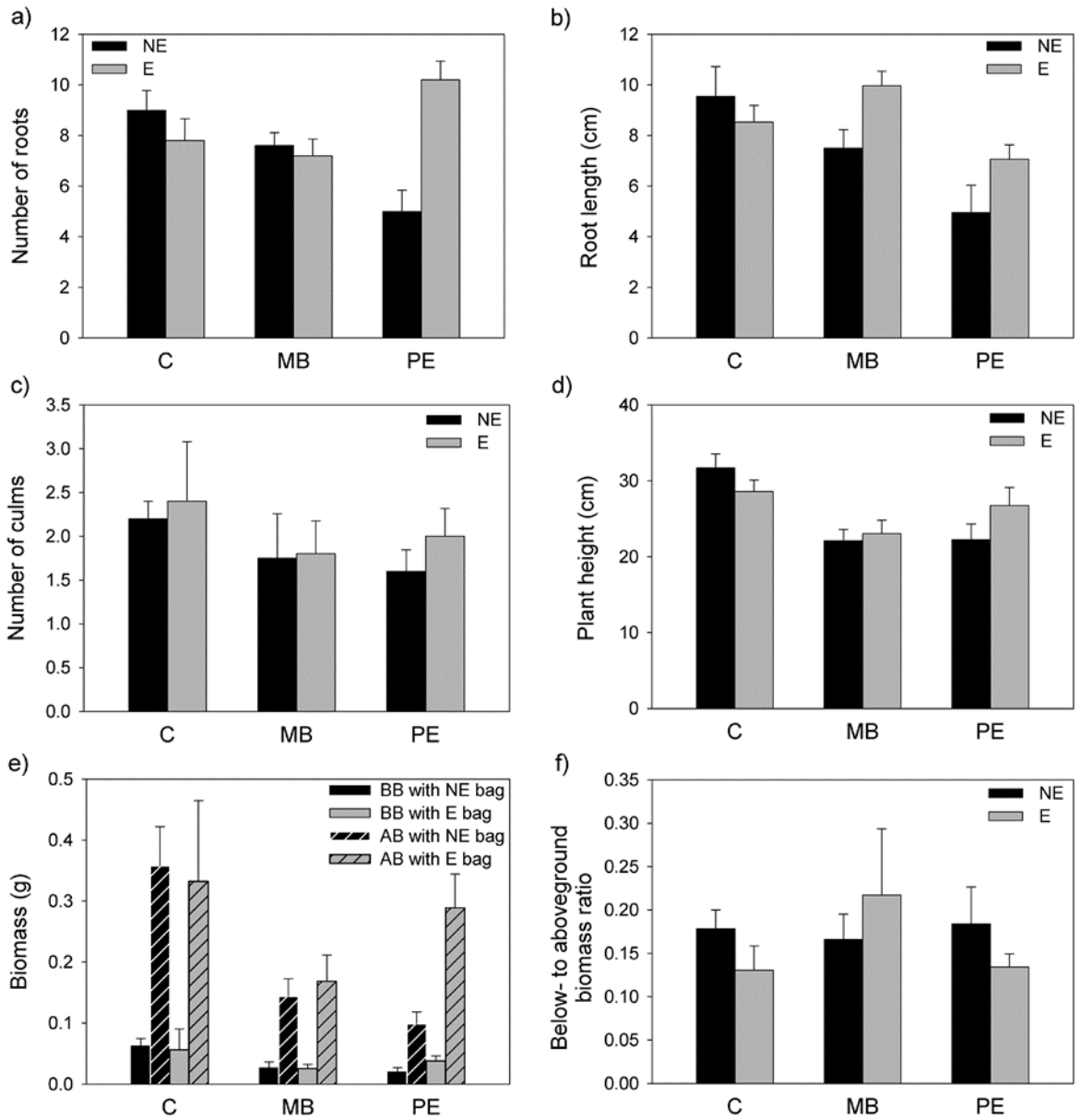
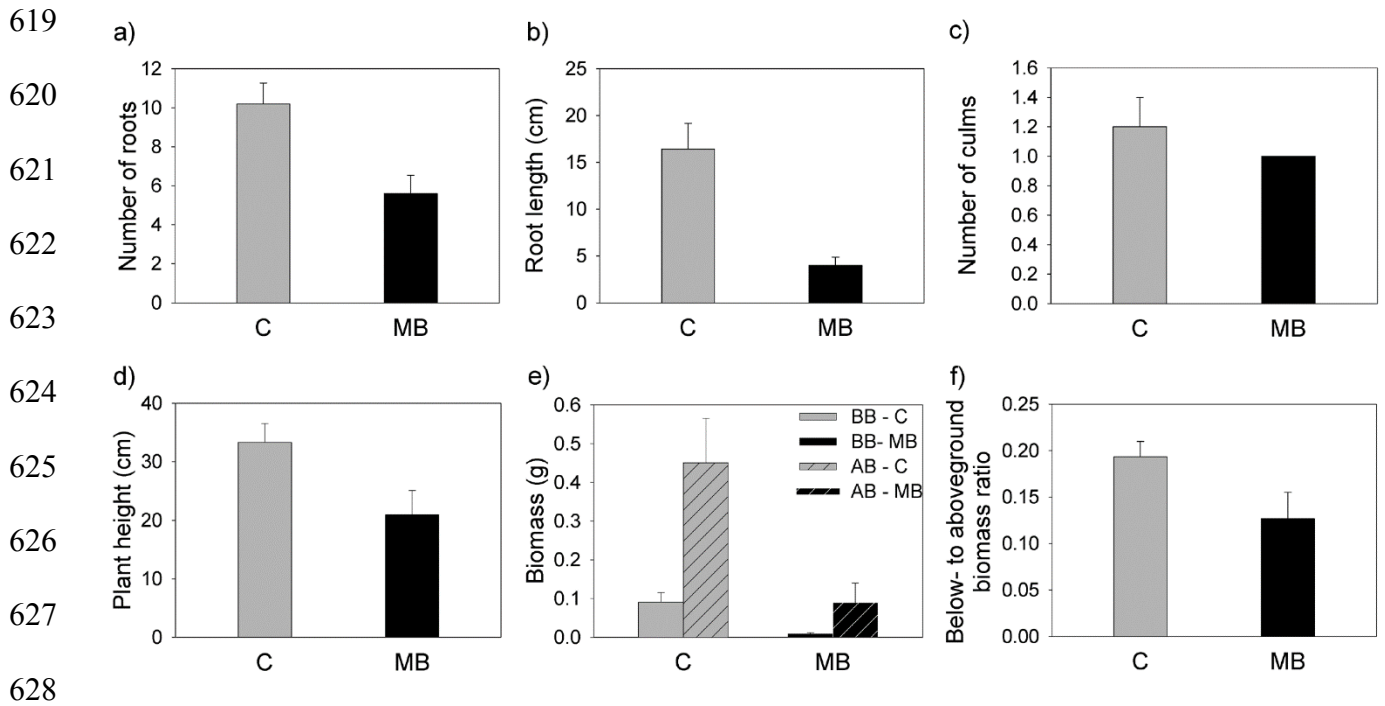


Fig. 2. Morphological and growth variables of *Thinopyrum junceum* seedlings emerged from control plots without bags (C) and plots with not exposed (NE) and beach exposed (E) Mater-bi® (MB) and high-density polyethylene (PE) bag fragments. Data are means \pm SE. BB: belowground biomass. AB: aboveground biomass.

2-column fitting image



629 **Fig. 3.** Morphological and growth variables of *Ammophila arenaria* seedlings emerged from control
 630 plots without bags (C) and plots with beach exposed Mater-bi® (MB) and high-density polyethylene
 631 (PE) bag fragments. Data are means \pm SE. BB: belowground biomass. AB: aboveground biomass.

632 2-column fitting image

633

634

635

636

637

638

639

640

641

642

643

644

645 **Table 1**

646 Results of ANOVA analysis on the percentage of emergence, mean emergence time and percentage
 647 of mortality of seedlings of *Thinopyrum junceum* and *Ammophila arenaria* grown in the absence
 648 and in the presence of not exposed and beach exposed Mater-bi®(MB) and high-density
 649 polyethylene (PE) bags. Significant results are in bold, and results of SNK tests are showed.

Source	df	<u>Emergence (%)</u>		df	<u>Mean emergence time</u>		<u>Mortality (%)</u>	
		F	p		F	p	F	p
Plastic (P)	2	15.75	<0.001	2	3.80	0.031	1.76	0.187
Exposure (E)	1	0.03	0.861	1	1.88	0.178	1.63	0.210
Species (S)	2	63.79	<0.001	1	5.33	0.026	0.71	0.406
P x E	2	0.39	0.675	2	0.55	0.579	0.97	0.388
P x S	4	1.91	0.113	2	0.59	0.558	0.52	0.599
E x S	2	0.34	0.710	1	1.87	0.180	0.25	0.618
P x E x S	4	0.65	0.629	2	4.28	0.021	0.12	0.888
Residual	126			36				
SNK test		MB = PE < C			NE, A: PE = C < MB			
		T > A > G			MB, A: E < NE			
					MB, NE: T < A			

663 C: absence of plastic bag; or control MB: presence of Mater-bi® bag, PE: presence of high-density polyethylene bag, NE: not exposed
 664 bag, E: beach exposed bag, T: *T. junceum*, A: *A. arenaria*, G: *G. flavum*

665
 666
 667
 668
 669
 670
 671
 672
 673
 674
 675

676 **Table 2**

677 Results of multivariate PERMANOVA (a) and ANOVA (b) analyses on morphological and growth
 678 variables of *Thinopyrum junceum* seedlings grown in the absence (C) and in the presence of not
 679 exposed (NE) and beach exposed (E) Mater-bi®(MB) and high-density polyethylene (PE) plastic
 680 bags. Significant results are in bold, and results of SNK tests are showed.

681 a) PERMANOVA

Source	df	Pseudo-F	p
Plastic (P)	2	3.21	0.006
Exposure (E)	1	1.14	0.326
P x E	2	2.33	0.041
Residual	24		
Transformation		Sqrt(x)	
Pair-wise test		NE: C ≠ MB = PE; PE: NE ≠ E	

690 b) ANOVA

Source	df	Root number		Root length		Number of culms			
		F	p	F	p	F	p	F	p
Plastic (P)	2	1.02	0.374	7.96	0.002	0.94	0.406		
Exposure (E)	1	3.95	0.058	3.02	0.095	0.00	0.975		
P x E	2	11.12	<0.001	2.64	0.092	0.74	0.488		
Residual	24								
Transformation		None		None		Ln(x)			
SNK test		NE: C = MB > PE E: C = MB < PE PE: NE < E		C = MB > PE					
Source	df	Plant height		Below biomass		Above biomass		Biomass ratio	
		F	p	F	p	F	p	F	p
Plastic (P)	2	8.90	0.001	2.35	0.116	4.54	0.021	0.21	0.811
Exposure (E)	1	0.26	0.611	0.13	0.717	1.29	0.267	0.66	0.423
P x E	2	2.10	0.145	2.21	0.132	2.96	0.071	0.51	0.607
Residual	24								
Transformation		None		Ln(x)		Ln(x)		Ln(x)	
SNK test		MB = PE < C				MB = PE < C			

709

710 **Table 3**

711 Results of multivariate PERMANOVA (a) and ANOVA (b) analyses on morphological and growth
 712 variables of *Ammophila arenaria* seedlings grown in the absence (C) and in the presence of beach
 713 exposed Mater-bi®(MB) plastic bags. Significant results are in bold, and results of SNK tests are
 714 showed.

715 a) PERMANOVA

Source	df	Pseudo-F	p
Plastic (P)	1	6.86	0.015
Residual	8		
Transformation		Sqrt(x)	
Pairwise test		C ≠ MB	

721
 722 b) ANOVA

Source	df	Root number		Root length		Number of culms		F	p
		F	p	F	p	F	p		
Plastic (P)	1	10.58	0.011	19.13	0.002	1.00	0.346		
Residual	8								
Transformation		None		Ln(x)		None			
SNK test		C > MB		C > MB					
Source	df	Plant height		Below biomass		Above biomass		Biomass ratio	
		F	p	F	p	F	p	F	p
Plastic (P)	1	5.46	0.047	13.46	0.006	8.32	0.020	4.07	0.078
Residual	8								
Transformation		None		Ln(x)		None		None	
SNK test		C > MB		C > MB		C > MB			

735

736

737

738

739

740

741

742 **Table 4**
 743 Results of ANOVA analysis on the
 744 percentage of weight loss of not
 745 exposed and beach exposed Mater-bi®
 746 (MB) and high-density polyethylene
 747 (PE) bags buried in sand above seeds
 748 of *Thinopyrum junceum*, *Ammophila*
 749 *arenaria* and *Glaucium flavum*.
 750 Significant results are in bold, and
 751 results of SNK test are showed.

		<u>Weight loss (%)</u>	
Source	df	F	p
Plastic (P)	1	14.30	<0.001
Exposure (E)	1	2.88	0.093
Species (S)	2	0.09	0.914
P x E	1	0.51	0.477
P x S	2	0.30	0.739
E x S	2	0.24	0.785
P x E x S	2	0.61	0.544
Residual	84		
Transformation		Ln (x+1)	
SNK test		MB > PE	

764
 765
 766
 767
 768
 769