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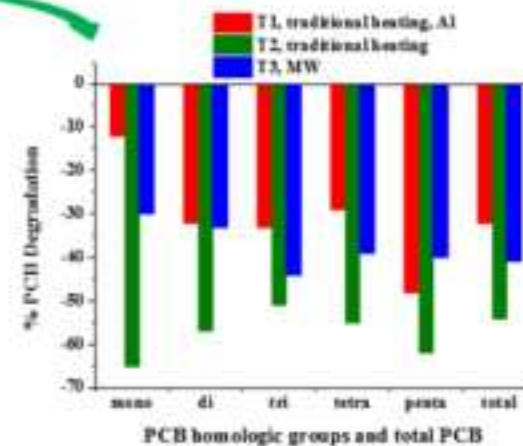
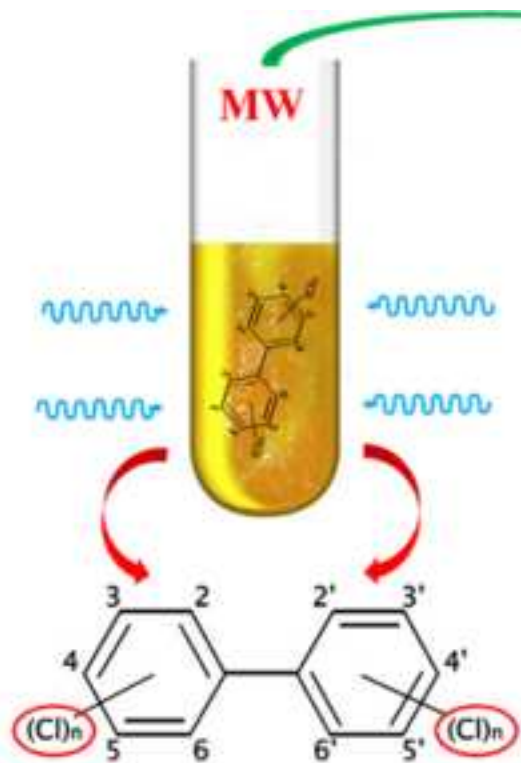
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## Transformer and Engine Oils



Remarkable dechlorination degree was achieved when microwave irradiation was employed as the only driving force!

## Abstract

The removal of polychlorinated biphenyls (PCBs) both from **siloxane** transformer **oil** and **hydrocarbon** engine oil was investigated through the application of microwave (MW) irradiation **and** a reaction system based on polyethyleneglycol (PEG) and potassium hydroxide. The influence of the main reaction parameters (MW irradiation time, molecular weight of PEG, amount of added reactants and **temperature**) on the dechlorination **behavior** was studied. Promising **performances** were **reached**, **allowing** about 50 % of dechlorination **under** the best experimental conditions, **together** time and energy **saving** compared to conventional heating systems. Moreover, **an interesting** dechlorination degree (up to 32 %) was achieved for **siloxane** transformer oil when MW irradiation was employed as the unique driving force. **To the best of our knowledge, this is the first time in which MW irradiation is tested as the single driving force for the dechlorination of these two types of PCB-contaminated oils.**

# Application of microwave irradiation for the removal of polychlorinated biphenyls from siloxane transformer and hydrocarbon engine oils

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## Abstract

The removal of polychlorinated biphenyls (PCBs) both from siloxane transformer oil and hydrocarbon engine oil was investigated through the application of microwave (MW) irradiation and a reaction system based on polyethyleneglycol (PEG) and potassium hydroxide. The influence of the main reaction parameters (MW irradiation time, molecular weight of PEG, amount of added reactants and temperature) on the dechlorination behavior was studied. Promising performances were reached, allowing about 50 % of dechlorination under the best experimental conditions, together time and energy saving compared to conventional heating systems. Moreover, an interesting dechlorination degree (up to 32 %) was achieved for siloxane transformer oil when MW irradiation was employed as the unique driving force. To the best of our knowledge, this is the first time in which MW irradiation is tested as the single driving force for the dechlorination of these two types of PCB-contaminated oils.

**Keywords:** microwave, PCBs, dechlorination, transformer oil, engine oil.

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## 25 **1. Introduction**

26 Since the late 1920, polychlorinated biphenyls (PCBs) have been used in a wide variety of  
27 applications thanks to their specific properties (Ivanov and Sandell, 1992). Unfortunately, these  
28 compounds are toxic and persistent pollutants that threaten not only the environment but also the  
29 human health (Wang D.G. et al., 2008) and due to their lipophilic nature, they can bio-accumulate  
30 in cells up to the food chain. Today, the production of PCBs has been banned all over the world, but  
31 there are many residual devices containing these compounds, such as transformer and engine oils,  
32 whose improper management practices can cause serious problems, making urgent their  
33 replacement and their proper disposal. However, the decontamination of oils from PCBs in  
34 electrical devices is extremely difficult and also the replacement of the contaminated oil with a new  
35 one is not a feasible solution because PCBs remain absorbed on electrical coils and layers of  
36 magnetic material, making impossible their complete removal, thus causing the inevitable  
37 contamination even of the new oil. Until now, the only proven and widely used technology for  
38 treating big amounts of polluted matrices is their incineration which can destroy over 99.9999 % of  
39 PCBs (Liu et al., 2011). However, there is a widespread public opposition to this approach due to  
40 the potential release of Cl<sub>2</sub>, HCl and polychlorinated dibenzo-p-dioxins and dibenzofurans via flue  
41 gas stream (Davila et al., 1994). In order to overcome these problems, a patented method (Tumiatti,  
42 1997), industrially employed, removes PCBs from different oils in the presence of  
43 polyethylene/polyalkylene glycols characterized by high molecular weight, hydroxides or alkali  
44 alcoholates and non-alkali metals working around 200°C. Two different degradation pathways may  
45 occur, the nucleophilic substitution and the hydrodechlorination. In fact, the first mechanism is  
46 typical of the polyethyleglycol/polyalkyleneglycols-base process where polyethylenglycolate,  
47 formed as a base is added to PEG, selectively attacks the active sites (halogen groups) of the  
48 polychlorobiphenyl molecule. In addition, according to this pathway, PEG also acts as phase  
49 transfer catalyst due to its high lipophilicity, moving polychlorobiphenyl molecules from the apolar

50 phase (oil) to the polar one (PEG/base) where the reaction does occur. The second mechanism  
51 (hydrodechlorination) is attributed to the presence of metal catalyst and hydrogen donors,  
52 enhancing the hydrodechlorination of PCBs. The industrially adopted method allows good removal  
53 yields but requires long reaction times, up to 2 h (Tumiatti, 1997). More recently, other approaches  
54 have been developed for remediation of PCBs, such as thermal, oxidative, reductive photocatalytic  
55 and microbial technologies (Beless et al., 2014; Borja et al., 2005; De Filippis et al., 1997; Field and  
56 Sierra-Alvarez, 2008; Habekost and Aristov, 2012; Jelic et al., 2015; Kanbe and Shibuya, 2001;  
57 Kume et al., 2008; O'Brien et al., 2005; Ohbayashi et al., 2002; Ohmura et al., 2007; Peng et al.,  
58 2014; Rastogi et al., 2009; Seok et al., 2005; Shaban et al., 2016; Van Aken and Bhalla, 2001;  
59 Villalobos-Maldonado et al., 2015; Wang et al., 2016; Wu et al., 2012; Wu et al., 2005; Yao et al.,  
60 2014; Zhao et al., 2015). In addition to these techniques, also microwave irradiation (MW) has been  
61 investigated due to its advantages ascertained in many different applications (Abramovitch et al.,  
62 1999a, 1999b, 1998; Cravotto et al., 2007; Huang et al., 2011; Kamarehie et al., 2014; Lin et al.,  
63 2013; Liu and Yu, 2006; Liu et al., 2008; Tajik et al., 2014). In fact, in the past few years the  
64 interest in the employment of MW irradiation has considerably increased because it represents a  
65 sustainable and green tool for many applications, improving selectivity together with a significant  
66 reduction of the reaction time (often by orders of magnitude) and of energy consumption (Antonetti  
67 et al., 2015, 2012, 2010; Appleton et al., 2005; Raspolli Galletti et al., 2013, 2010, 2008). As a  
68 novel and efficient approach, MW irradiation has been especially used to treat contaminated soils,  
69 obtaining interesting results. A series of pioneer studies were reported by Abramovitch on MW  
70 irradiation for the remediation of PCBs from polluted soil in the presence of microwave absorbers  
71 ( $\text{Cu}_2\text{O}$ , graphite) and NaOH. He reported that most of the chlorinated aromatics were decomposed  
72 and the majority of the dechlorinated products were possibly mineralized into the soil (Abramovitch  
73 et al., 1999a, 1999b, 1998). Liu and Yu investigated the combined effects of MW irradiation and  
74 granular activated carbon on the removal of PCBs from soils and indicated that the addition of

75 carbon to soil effectively increases its ability to absorb MW energy, **this heating** resulting in the  
76 enhanced degradation of PCBs in the soil (Liu and Yu, 2006). The employment of **a MW** absorber  
77 has been also studied by Huang which investigated the removal efficiency in PCB-contaminated  
78 soils by microwave irradiated MnO<sub>2</sub>, absorber of microwave energy, obtaining a removal  
79 percentage above 95 % (Huang et al., 2011). Cravotto combined solid Fenton-like reagents (sodium  
80 percarbonate and the urea/hydrogen peroxide complex) and MW irradiation for the decomposition  
81 **of organic pollutants of the soil**, highlighting that 4-chloronaphthol, 2,4-dichlorophenoxyacetic acid  
82 and *p*-nonylphenol were completely degraded (Cravotto et al., 2007). **On the other hand**, the use of  
83 **MW** irradiation for the removal of PCBs from oil matrices has been less investigated (Kamarehie et  
84 al., 2014; **Kastanek et al., 2011**; Lin et al., 2013; Liu et al., 2011; Liu et al., 2008; Tajik et al., 2014).  
85 Liu studied the applicability of MW irradiation for the removal of PCBs from the soil heavily  
86 contaminated by capacitor oil in the presence of sodium hypophosphite, iron powder and granular  
87 activated carbon as MW-adsorbing materials. In the presence of sodium hypophosphite and carbon,  
88 about 80 % of PCBs in soils were effectively removed by MW irradiation for 10 minutes and the  
89 further addition of iron powder raised the average removal efficiency up to 95 % (Liu et al., 2008).  
90 Lin investigated also the effect of MW in the removal of PCBs from soil polluted with capacitor oil  
91 in the presence of MnO<sub>2</sub> as MW absorber and oxidizer. The removal efficiencies for di-, tri-, tetra-,  
92 penta-, hexa-, hepta- and octa- chlorobiphenyls were 95.9, 82.5, 52.0, 71.6, 62.5, 28.6 and 16.1 %,   
93 respectively, applying MW irradiation for 45 minutes to a mixture of MnO<sub>2</sub>, contaminated soil and  
94 water (Lin et al., 2013). **Kastanek examined the influence of MW field on the effectiveness of**  
95 **KPEG method (nucleophilic substitution by alkaline polyethylene glycol, PEG) in the removal of**  
96 **PCBs from highly contaminated mineral oils. The obtained results show that MW irradiation**  
97 **significantly increases the reaction rate and the method's effectiveness. Moreover, the authors**  
98 **highlight that the addition of a small amount of the ionic liquid 1-butyl-3-methylimidazolium**  
99 **hexafluorophosphate markedly positively affects the results achieved under MW irradiation**

100 (Kastanek et al., 2011). However, the paper does not deal with the influence of some important  
101 parameters on the achieved results, such as temperature, MW irradiation time, molecular weight of  
102 PEG, amount of added reactants on the effectiveness of the KPEG method. Finally, MW irradiation  
103 was employed as tool to induce the hydrothermal reaction between iron powder, NaOH and H<sub>2</sub>O as  
104 reactants in order to produce hydrogen, responsible then for the dechlorination of PCBs. The  
105 efficiency of this method was verified on a simulated oil based on hexadecane containing 100 mg/L  
106 of Aroclor 1254, achieving the almost complete dechlorination after MW treatment for 10 minutes  
107 (Liu et al., 2011).

108 Now, this study investigates the efficiency of MW irradiation compared to the traditional heating  
109 and the influence of some selected reaction conditions on the removal of PCBs from two types of  
110 oils: siloxane transformer oil and hydrocarbon engine one, adopting the industrially applied reaction  
111 mixture made of PEG and potassium hydroxide in the presence or in the absence of aluminum  
112 powder. In addition, to the best of our knowledge, this is the first application in a systematic  
113 investigation of MW efficiency when applied as unique driving force in the dechlorination of these  
114 two different oil matrices, in the absence of other chemicals.

115

## 116 **2. Material and methods**

### 117 **2.1 Materials**

118 Aroclor 1242 pure (1mL neat - F109 – lotto 399-6B) was purchased from Chemservice srl (MI-  
119 Italy). PCB 19\* (D), PCB 156\* (D), PCB 81\* (D) (50 µg/mL in toluene) were provided from  
120 Wellington Laboratories and PCB mixtures of individual congeners (PCB 525; 525.1; PCB mix.  
121 500 µg/mL each in hexane) were obtained from Supelco and used as received.  
122 Polydimethylsiloxane transformer oil was provided from Chimica San Fedele, whereas hydrocarbon  
123 engine oils are commercial products: Q8 Formula Plus (cod. SAE 15W-40) and Q8 Formula F1  
124 (cod. SAE 10W-50). The model polluted siloxane transformer oil sample was prepared by adding



125 400  $\mu$ L of Aroclor 1242 in isooctane to 40 mL of the transformer oil, whereas the model  
126 **hydrocarbon** engine oil was prepared by adding 400  $\mu$ L of Aroclor 1242 in isooctane to 40 mL of a  
127 mixture of Q8 Formula Plus and Q8 Formula F1 oils (1:1, v/v). Polyethylene glycol 8000,  
128 polyethylene glycol 200, ethylene glycol, aluminum powder, potassium hydroxide, potassium  
129 bicarbonate and hexane were purchased from Sigma Aldrich and used as received.

130

## 131 **2.2 Methods**

132 The MW dechlorination experiments were carried out using a commercially available mono-mode  
133 microwave unit (CEM Discover S-class system) equipped with a calibrated external infrared sensor  
134 and a PC control. The instrument consisted of a single-mode, self-tuning cavity where the correct  
135 position of the reaction vial was automatically tuned to ensure reproducible conditions at each run.

136 The oven was operated by a continuous power generator capable of supplying an irradiation power  
137 from 0 to 300W (with 1W-step increment) with a fine and automated control of the sample  
138 temperature based on the CEM PowerMax technology. Initial reaction time ( $t = 0$ ) was taken when  
139 the temperature set point was reached. Air simultaneous cooling of the sample during MW  
140 irradiation (cooling system PowerMAX) was adopted for selected runs, as specified in the text.

141 This is a new technology which allows the simultaneous cooling of the reactor with a stream of air  
142 on the outside of the reaction vessel driving its heating with a substantial amount of microwave  
143 energy, up to 300 W. In this way, when the power is increased, it is possible to avoid the risk of  
144 overheating temperature-sensitive reactants, enabling more energy to be introduced into the reaction  
145 without the degradation of the product.

146 In a typical procedure, 2.5 g of the selected oil and the right amounts of chosen reagents were  
147 **introduced** in a 10 mL reactor equipped with a magnetic teflon stir bar. The reactor was capped and  
148 heated to the desired temperature under maximum stirring. After reaction time was over, the sample

149 was rapidly cooled to room temperature, filtered and washed several times with hexane in order to  
150 recover the total amount of the oil for the successive PCB analysis.

151 Traditionally heated dechlorination reactions were performed in 50-mL two-neck round-bottom  
152 flask equipped with a condenser. In a typical procedure, 10 g of the selected oil and the **selected**  
153 amounts of chosen reagents were loaded in the round-bottom flask equipped with a magnetic teflon  
154 stir bar. The run was carried out at the desired temperature by heating in oil bath previously  
155 thermostated at the stated temperature. Upon completion of the reaction, the sample was quenched  
156 in a cold water bath and worked-up as reported above. For both approaches reproducibility of  
157 triplicate experiments resulted within 3 %. **Detailed experimental conditions adopted in the**  
158 **dechlorination reaction on the siloxane transformer and hydrocarbon engine oils are reported in the**  
159 **Supplementary Information (see S1 and S2 respectively in the Supplementary Information).**

160

### 161 **2.3 Analysis**

162 Before the analysis, purification of the oil samples was carried out according to the clean-up  
163 technique SPE (Solid Phase Extraction) using a column Supelco Supelclean Sulfoxide PPSPE  
164 3g/6mL consisting of a silica-bonded sulfoxide (-SO) phase. The column was washed with acetone  
165 and then equilibrated with hexane. At the same time, the oil sample was diluted with hexane and  
166 two internal standards of PCB congener, PCB 19\* and PCB 156\*, were added in order to evaluate  
167 the average recovery percentages of PCBs from each oil sample during the clean-up procedure.  
168 Later, the treated sample was layered onto the upper part of the SPE column. Then, the sample was  
169 eluted with hexane and two fractions were collected for the subsequent analysis. The first one  
170 mainly includes apolar hydrocarbons, whereas the second one primarily contains polar  
171 hydrocarbons and PCBs. This second fraction was subsequently diluted to the right volume and  
172 loaded with the internal standard (PCB 81\*) for the following quantitative determination. The  
173 recovery percentages of the internal standards (PCB 19\* and PCB 156\*) for each oil sample and the

174 corresponding average values are reported in the **Supplementary Information** (see **S3** in the  
175 **Supplementary Information**). The average values were employed for the calculation of PCB  
176 concentrations after GC-MS analysis.

177 **Identification and** quantitative analysis of PCBs and **reaction media products** was accomplished by  
178 a Thermo Finnigan MS Trace DSQ gas chromatography/mass spectrometry (GC-MS) equipped  
179 with a G.C. column Agilent DB5-MS Ultra Inert (column length 30 m, inner diameter 0.25 mm and  
180 thickness of the stationary phase 0.25  $\mu\text{m}$ ). The transport gas was helium 5.5 **with a flow of** 1.3  
181 mL/min. The temperature of the injection port was set at 250 °C, carrier pressure at 100 kPa and the  
182 injection was carried out in the splitless mode. The column temperature was programmed as  
183 follows: the oven was heated at 45 °C for 3 minutes, then the temperature was raised by 30 °C/min  
184 up to 130 °C remaining at this temperature for 2 minutes, then the temperature was raised by 8  
185 °C/min up to 250 °C and as soon as the temperature had reached the desired value, it was again  
186 raised by 30 °C/min up to 300 °C remaining at this temperature for 6 minutes.

187 **The determination of water amount in the oil samples was carried out according to the ASTM D**  
188 **6304-07 method (ASTM International, 2007).**

189

### 190 **3. Results and Discussion**

#### 191 **3.1 Dechlorination activity adopting traditional heating**

192 The experiments were carried out using a commercial **siloxane** transformer oil and a commercial  
193 hydrocarbon engine one. In preliminary experiments, we replicated PCB degradation under the  
194 reaction conditions of a patented industrial method (Tumiatti, 1997) using traditional heating at  
195 100°C for 2 h. The reaction involves the contemporary presence of high molecular weight  
196 polyethyleglycol (PEG 8000), potassium hydroxide (KOH) 3 % and Al powder.

197 Alternatively, we tested the same reaction without Al, **considering that this metal is unsuitable in a**  
198 **successive study under MW irradiation.**

199 Both reactions were carried out with the two different matrices, i.e. **siloxane** transformer oil (TO)  
200 and **hydrocarbon** engine oil (MO), spiked with Aroclor 1242 (Table 1).

201 Table 1, near here

202 The comparison between runs T1 and T2 shows a partially inhibiting role of Al; the total  
203 degradation efficiency of PCBs was in fact 54% in the runs without Al, significantly higher than  
204 32% in the runs with Al. On the contrary, Al plays a negligible role in PCB degradation  
205 efficiencies, both in the homologous groups and in the total degradation, in the M1 and M2 runs  
206 (engine oil).

207 The comparison of PCB degradation in runs carried out under the same reaction conditions (T1-M1  
208 and T2-M2) highlights a significant matrix effect: the experiments involving TO show higher  
209 efficiencies. These results suggest that the different chemical and physical properties of the two oils  
210 influence the ability of the systems to create specific interactions between the different mixture  
211 compounds, modifying the reactivity of the involved species. **In particular, the different  
212 dechlorination activity could be ascribed to the different viscosity of the two oil matrices. In fact,  
213 the employed siloxane transformer oil is characterized by a viscosity of about  $50 \text{ mm}^2 \text{ s}^{-1}$ , whereas  
214 the hydrocarbon engine one shows an average viscosity of about  $100 \text{ mm}^2 \text{ s}^{-1}$ . The lower viscosity  
215 of siloxane transformer oil can help the mixing and the contact between the reagents, thus favoring  
216 the dechlorination reaction, leading to higher % PCB degradations, both in the homologous groups  
217 and in the total degradation.**

218 These preliminary experiments give also some insight into the involved reaction mechanisms.  
219 According to the literature, the main degradation pathway which occurs should be the nucleophilic  
220 substitution dechlorination, although in the presence of Al **also** the hydrodechlorination mechanism  
221 can work.

222 In our tests, we have not observed accumulation of low chlorine PCBs in **the** runs working with Al;  
223 this **suggests** that the major mechanism is the nucleophilic substitution, whose reaction pathway is  
224 reported in Scheme 1.

225 Scheme 1, near here

226 Furthermore, the data of Table 1 suggest that, when TO is employed, Al can fine-tune the  
227 mechanism of dechlorination, favoring the reaction from highly chlorinated congeners to less  
228 chlorinated ones. **In fact**, run T1 in the presence of Al shows a higher removal degree for  
229 pentachlorinated biphenyls that **decrease** by 48 % **respect to** monochlorinated ones, whose reduction  
230 was only 12 %. On the contrary, in the absence of Al (run T2, Table 1), the degradation of  
231 chlorinated species appears more statistically distributed on an average value of 55-60 % for all  
232 congeners.

233

### 234 **3.2 Dechlorination activity adopting microwave irradiation**

235 In order to make the PCB removal greener and more sustainable, we tested the employment of MW  
236 in the PCB degradation on both oil matrices.

237

#### 238 **3.2.1 Dechlorination activity adopting MW irradiation on the siloxane transformer oil**

239 **The comparison between the employment of traditional heating and MW irradiation on siloxane TO**  
240 **is shown in Figure 1, where the runs T1 and T2 (reaction conditions in Table 1) were carried out**  
241 **adopting traditional heating, whereas T3 was performed under MW irradiation at 100°C and 300 W**  
242 **for 15 minutes in the same reaction conditions (PEG 8000 4.5 % w/w, KOH 3 %).**

243 Figure 1, near here

244 The use of MW (run T3) gives intermediate PCB degradation efficiency between runs T1 (with Al)  
245 and T2 (without Al) **adopting** a reaction time of 15 minutes **under the same reaction conditions**.

246 Figure 2A shows that, doubling the reaction time (run T3L, **30 minutes**) **under the same reaction**

247 conditions adopted in run T3 (PEG 8000 4.5 % w/w, KOH 3 %, 100°C, 300 W), the efficiency of  
248 PCB degradation is comparable to that of run T2 (reaction time: 2 h).

249 Figure 2, near here

250 The comparison between the runs T3 and T3L suggests that the selectivity seems to be addressed  
251 towards the degradation of the highly chlorinated congeners, at least with shorter reaction times.  
252 From this point of view, MW irradiation works similarly to the metal catalyst. As the reaction  
253 proceeds, all the congeners are degraded in the same amount and the reaction has reached an  
254 advanced degradation degree.

255 Employing the simultaneous cooling of the sample (run T3R) with 15 minutes of reaction time, the  
256 effect of microwave is maximized, resulting into a total degradation efficiency of 54 %, analogous  
257 to that obtained in 2 h adopting traditional heating (run T2). The simultaneous cooling of the reactor  
258 with a stream of air on the outside of the reaction vessel allows the heating with a substantial  
259 amount of MW energy. It is well-established that the ability of MW to quickly superheat the  
260 reactants increases the yield decreasing side reactions. In fact, MW energy is transferred kinetically  
261 to the reactants and when higher amount of energy is released to the reactants, greater yields with  
262 fewer side reactions were achieved. However, when the power is significantly increased, the risk of  
263 overheating temperature-sensitive reactants and of causing degradation of the product can occur. It  
264 is possible to solve this problem by cooling the reactor with compressed gas on the outside of the  
265 reaction vessel, while simultaneously irradiating the sample with MW energy up to 300 W.

266 The effect of the molecular weight of polyethyleneglycol on the dechlorination efficiency was  
267 investigated. For this purpose, two experiments were performed on TO employing PEG with a  
268 lower molecular weight, PEG 200 4.5 % w/w in run T4 and the monomer ethylene glycol (EG) 4.5  
269 % w/w in run T5, both runs carried out under the same reaction conditions (KOH 3 %, 100°C, 15  
270 minutes, 300 W). The reduction of the molecular weight of the reagent causes a significant decrease  
271 of the dechlorination activity (Figure 2B), which reaches the lowest value in the presence of EG

272 (run T5). The reason of this trend can be attributed to the different repartition coefficient of PCBs  
273 between the oil phase and the glycol: the increase of the molecular weight of PEG can cause an  
274 increased solubility of PCB species in the PEG, thus favoring the dechlorination reaction. Regarding  
275 the selectivity, it is possible to observe that independently from the molecular weight of PEG, the  
276 system at the start of the reaction is more active in the dechlorination of highly chlorinated  
277 congeners which lose one Cl atom at a time, thus increasing the concentration of less chlorinated  
278 species until these last ones are in turn activated and dechlorinated. This behavior is markedly  
279 evident for run T5 where the concentrations of low-chlorinated biphenyls (mono- and di-  
280 chlorobiphenyls) increase compared to their levels in the starting oil. This can be rationalized  
281 considering that the activity of the reaction, although low, is aimed at the degradation of high-  
282 chlorinated biphenyls (tri-, tetra- and penta-chlorobiphenyls) which give lighter chlorinated species.  
283 It is also possible to observe the same behavior in run T4, where the percentages of removal of the  
284 first and the second class of congeners were significant lower than those observed for the other  
285 more chlorinated classes (tri-, tetra- and penta-chlorobiphenyls).

286 It is important to highlight that each GC-MS analysis does not evidence the production of  
287 polychlorobiphenyl alcohols, precursors of dangerous and toxic polychlorinated dibenzofurans  
288 (Choudhary et al., 1983).

289 To better investigate the role of MW irradiation, three runs were performed on TO employing only  
290 MW; this approach has not been reported in the literature so far. Only few blank experiments on  
291 real contaminated oils and soils without the addition of any reagents in the only presence of MW  
292 are reported in the literature as reference runs (Kamarehie et al., 2014; Tajik et al., 2014; Liu et al.  
293 2008) and evidenced little efficiencies towards PCB removal. The effects of the reaction time, runs  
294 T6 performed for 15 minutes and T6L for 30 minutes, and of the simultaneous cooling, run T6R  
295 carried out for 15 minutes with cooling, were investigated at 100°C and 300 W and the results are  
296 shown in Figure 2C.

297 The correlation between the increase of the reaction time and the higher dechlorination activity  
298 appears clear in **siloxane matrix**, moving from a total degradation degree of 22 % after 15 minutes  
299 up to 26 % after 30 minutes. Moreover, a further improvement in the PCB removal was achieved  
300 when the reaction was performed with the simultaneous cooling of the sample, **supplying more**  
301 **energy and** reaching a total dechlorination value of 32 %. **In the literature it has been reported that**  
302 **higher MW power enhances the removal of PCBs (Liu et al., 2008).**

303 The possibility to perform PCB degradation by **this MW-assisted** process in **the** liquid phase is quite  
304 **interesting**, because **the same** does not work under conventional heating conditions. **The higher**  
305 **degradation efficiency of MW irradiation respect to conventional heating has been related to the**  
306 **direct interaction of the reaction medium and of the PCB molecules with the MW irradiation, these**  
307 **last being more degraded when the dipole moment is higher. Moreover, the eventual formation of**  
308 **micro/nanoscale hotspots, causing a not uniform temperature of the irradiated medium, can occur**  
309 **(Huang et al., 2011).**

310

### 311 3.2.2 Dechlorination activity adopting MW irradiation on the hydrocarbon engine oil

312 **The influence of MW on dechlorination** is always positive also for this type of oil in comparison to  
313 conventional heating. Figure 3A shows **the influence of microwave irradiation time and of cooling**  
314 **system adopting the already tested reaction conditions (PEG 8000 4.5 % w/w, KOH 3 %, 100°C,**  
315 **300 W), highlighting as run M3 performed under MW irradiation for 15 minutes achieves the same**  
316 **dechlorination efficiencies of run M2 carried out under conventional heating for 2 h.**

317 Figure 3, near here

318 According to previously described results of experiments with TO, doubling the reaction time **under**  
319 **the same reaction conditions (PEG 8000 4.5 % w/w, KOH 3 %, 100°C, 300 W)** as **evidenced by run**  
320 **M3L, which was carried out for 30 minutes**, the total PCB degradation degree increases by 7 %, that



321 is approximately the same amount observed in parallel experiments on TO (run T3 and T3L, **Figure**  
322 **2A**).

323 However, **in this case**, the effect of the MW combined with simultaneous cooling (run M3R) is not  
324 **remarkable**: runs M3 and M3R **after 15 minutes of reaction time under the same reaction conditions**  
325 show comparable results, as again shown in Figure 3A.

326 The molecular weight of PEG on dechlorination efficiency **evidences** the same behavior already  
327 observed for the TO matrix, as **shown** in Figure 3B **where run M4 was carried out in the presence of**  
328 **PEG 200, whereas run M5 of ethylene glycol (EG), adopting the same reaction conditions (KOH 3**  
329 **%, 100°C, 15 minutes, 300 W)**. Also in the MO system, a progressive reduction in the activity  
330 moving from PEG 80000 to PEG 200 and, finally, to EG **is ascertained**.

331 The selectivity of PCB degradation in MO seems to follow similar rules as in TO **and the**  
332 **concentrations** of low-chlorinated biphenyls (mono-, di- and tri-chlorobiphenyls for run M5)  
333 increase **respect to the initial ones**.

334 Figure 3C shows the results of the experiments carried out in the presence of MW irradiation as the  
335 only driving force to explore the effects of reaction time, **runs M6 performed for 15 minutes and**  
336 **M6L for 30 minutes**, and of the simultaneous cooling, **run M6R carried out for 15 minutes with**  
337 **cooling, all runs tested at 100°C and 300 W**.

338 **Also in the case of MO oil, GC-MS analysis does not evidence the production of**  
339 **polychlorobiphenyl alcohols**.

340 In this case, the pure MW thermal treatment does not give very encouraging results; the total  
341 dechlorination with 15 minutes of reaction time is only 2 % and rises to 7 % by doubling the  
342 reaction time. The simultaneous cooling gives poor results too, because the highest efficiency is  
343 only 4 %.

344

345

346 3.2.3 Effect of temperature on the dechlorination activity under MW irradiation as the unique  
347 driving force on both oil matrices.

348 In order to investigate the influence of temperature on the dechlorination activity adopting MW  
349 irradiation as the only driving force, other two runs, M7 and T8 on MO and TO respectively, have  
350 been carried out at 180°C in the presence of the unique MW irradiation. After 15 minutes of  
351 reaction, the achieved results are shown in Figure 4 together with the results of the corresponding  
352 tests carried out at 100°C (M6 and T7 on MO and TO respectively).

353 Figure 4, near here

354 From the figure, it is possible to appreciate an increase of PCB removal efficiencies with  
355 temperature for MO oil (compare run M6 with run M7) and a lower increase for the more efficient  
356 TO oil (compare run T6 with run T7).

357 Although in the presence of only MW irradiation the reaction mechanism is not clear and studies  
358 are in progress, it is possible to hypothesize that a hydrodechlorination mechanism can occur, as  
359 proposed by other authors in the literature for the decomposition of PCBs in the soil (Tanirmchi et  
360 al., 1997). To better clarify the degradation pathways, the identification of trace-intermediate  
361 products by GC-MS during the process is now in progress testing different reaction conditions.

362 The comparison of the experiments carried out under MW irradiation on MO confirms that this type  
363 of oil is less suitable for PCB degradation, highlighting again a matrix effect. In the presence of  
364 MW, this behavior could be rationalized taking into account that the commercial transformer oil  
365 based on polydimethylsiloxane has a dielectric constant of 2.8, whereas the engine mineral oil  
366 composed of hydrocarbons shows a lower value of the dielectric constant (about 2.0), thus resulting  
367 in a slightly lower microwave absorption (Farcich N.J. et al., 2008). Moreover the siloxane oil has a  
368 water content of 200 ppm which can also slightly contribute to MW absorption by the medium,  
369 whereas the engine hydrocarbon oil has the moisture content less than 10 ppm, as determined by  
370 Karl Fischer method (ASTM International, 2007).

#### 371 **4. Conclusions**

372 The MW-assisted removal of polychlorinated biphenyls on siloxane transformer oil and  
373 hydrocarbon engine oil was studied and compared with traditional heating. The time of MW  
374 irradiation, the type and amount of added co-reactants and the temperature result important  
375 parameters influencing the dechlorination efficiency. Interesting results in the presence of MW  
376 were achieved, reaching up to about 50 % of dechlorination degree, saving time and energy  
377 compared to conventional heating. In addition, a remarkable dechlorination degree, up to 32 %, was  
378 reached for siloxane transformer oil when MW irradiation was employed in the absence of co-  
379 reactants. To the best of our knowledge, this is the first application of the MW irradiation as the  
380 unique driving force in a systematic investigation on the dechlorination of heavily polluted siloxane  
381 transformer and hydrocarbon engine oils, in the absence of other chemicals. Work is now in  
382 progress to improve this MW-approach method, extending the study to other PCB-containing  
383 wastes and modifying the reaction conditions, also with the aim of clarifying the involved reaction  
384 mechanism.

385

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389

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518 **Table 1:** Influence of Al on the % PCB degradation in homologic groups and % PCB total  
 519 degradation in the dechlorination of siloxane transformer oil (runs T1 and T2) and hydrocarbon  
 520 engine oil (runs M1 and M2) compared to the corresponding starting ones, TO (siloxane  
 521 transformer oil) and MO (hydrocarbon engine oil) added with amounts of Aroclor 1242. The  
 522 reaction was carried out in the presence of PEG 8000 4.5 % w/w, KOH 3 % for 2 h at 100°C under  
 523 traditional heating.

524

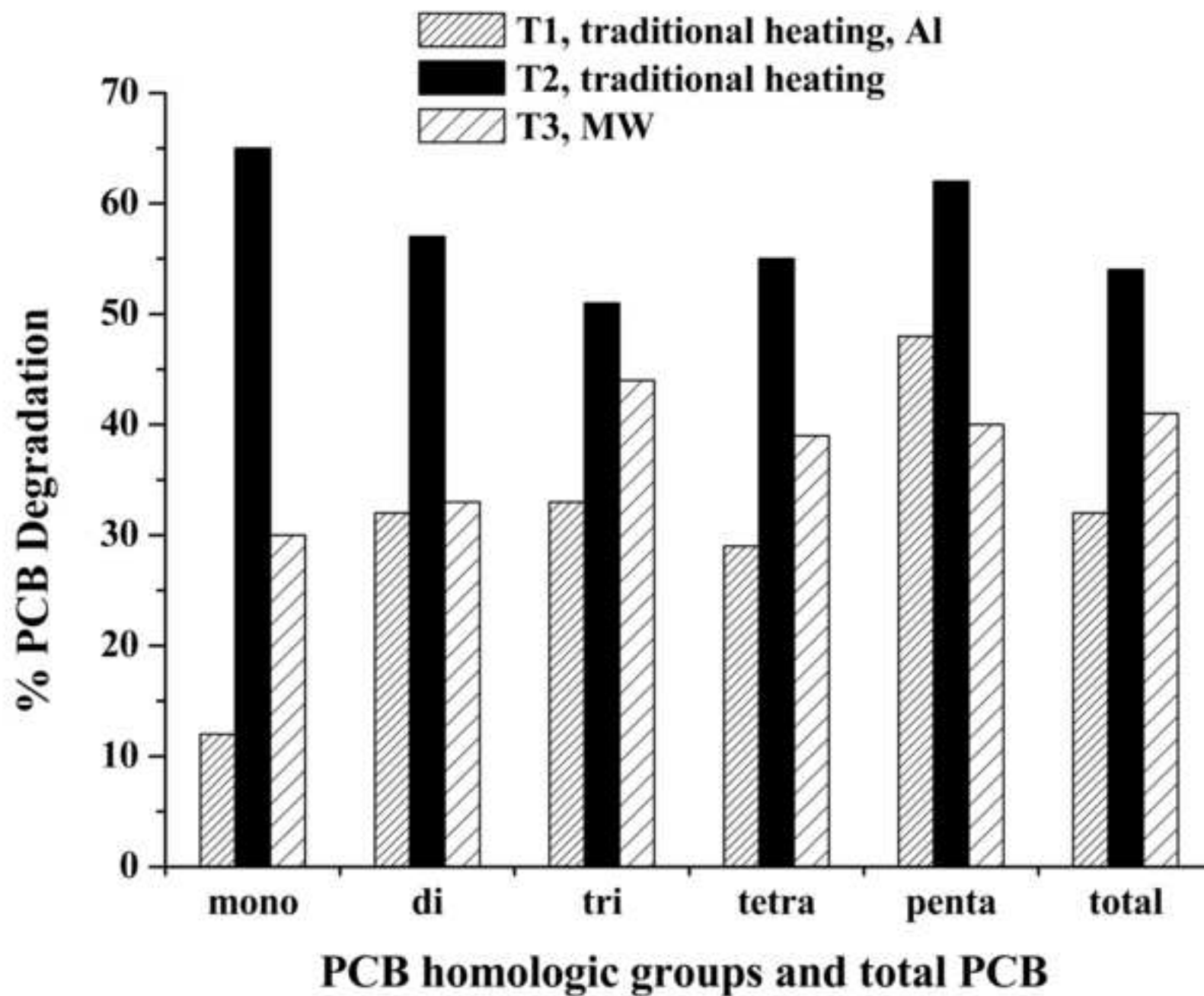
Run	Al w/w	Added PCBs (ppb)					Total added PCBs (ppb)
		mono	di	tri	tetra	penta	
TO	-	485	9108	37102	28238	6244	81177
MO	-	306	6598	19298	20352	4307	50861
		% PCB Degradation in homologic groups					% PCB Total Degradation
T1	0.5 %	12%	32%	33%	29 %	48%	32%
T2	-	65%	57%	51%	55%	62%	54%
M1	0.5 %	27%	24%	15%	20%	20%	19%
M2	-	34%	16%	20%	19%	22%	19%

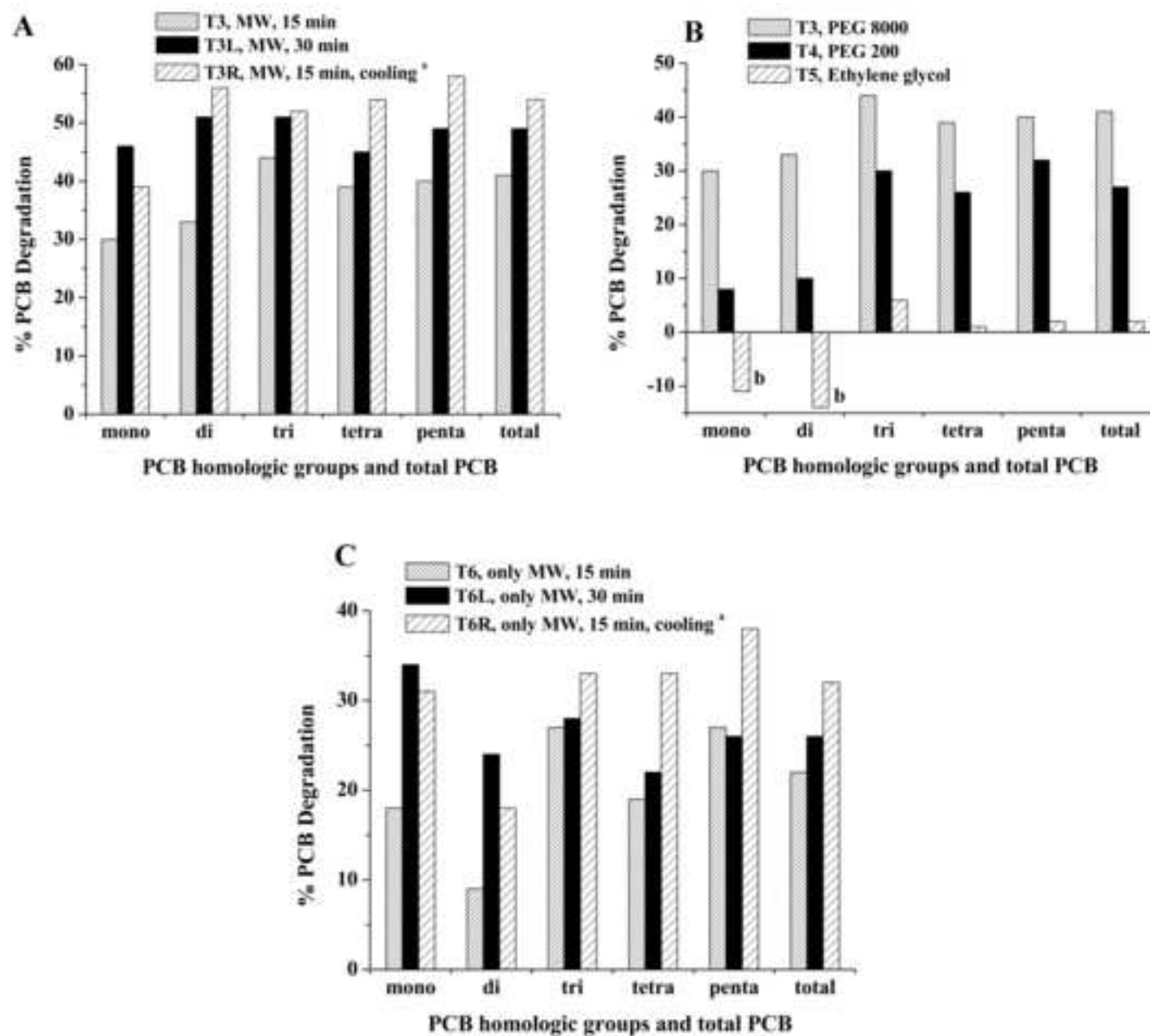
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526

## Highlights

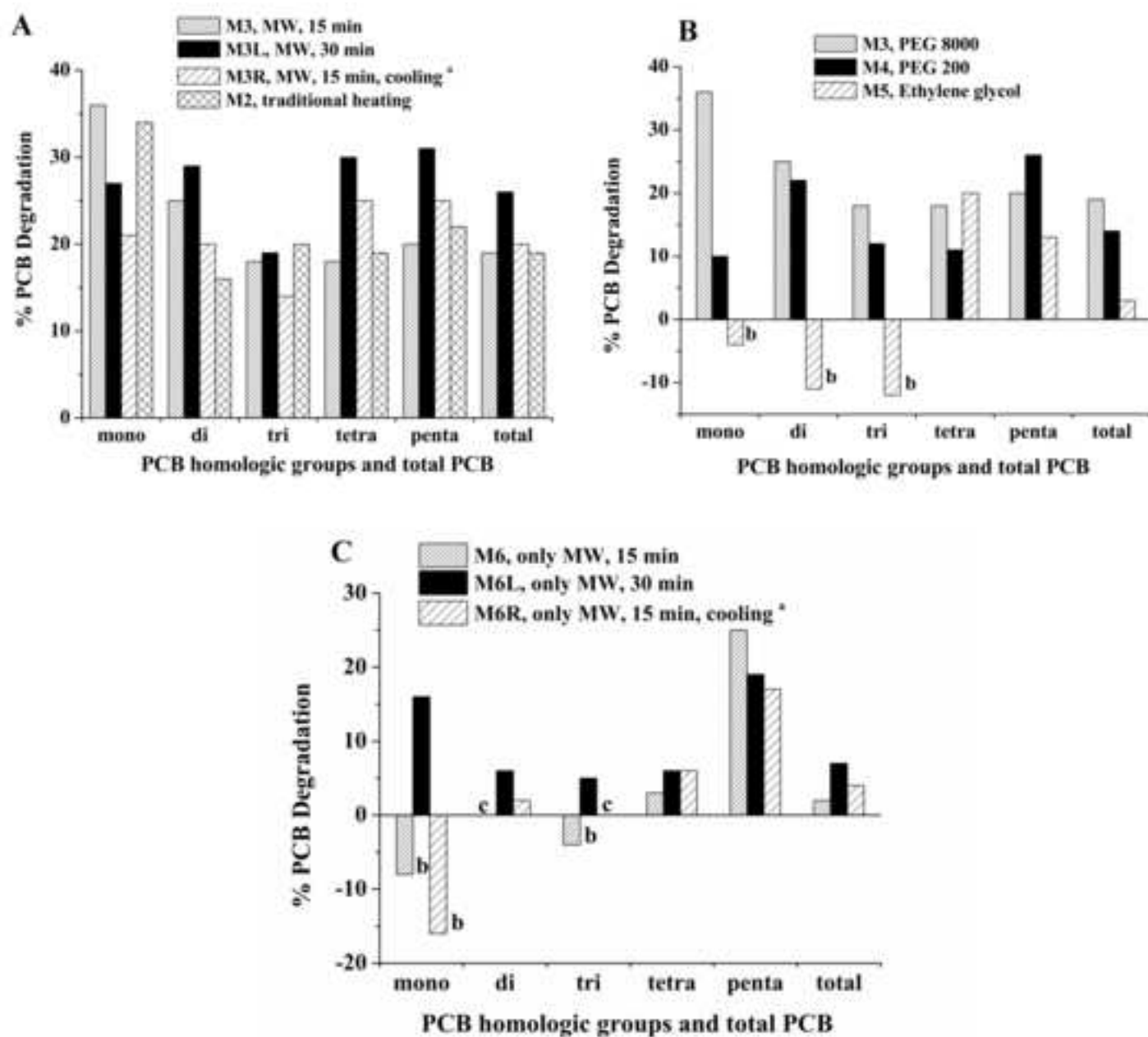
- Microwave irradiation has been applied for the removal of polychlorinated biphenyls.
- Transformer and engine oils were tested for the removal of polychlorinated biphenyls.
- For the first time microwave irradiation was employed as the only driving force.
- Microwave irradiation was used to improve the current industrial removal approach.
- Adopting microwave irradiation promising dechlorination degrees were achieved.





<sup>a</sup> in this case the sample is cooled using air and MW-irradiated at the same time (simultaneous cooling);

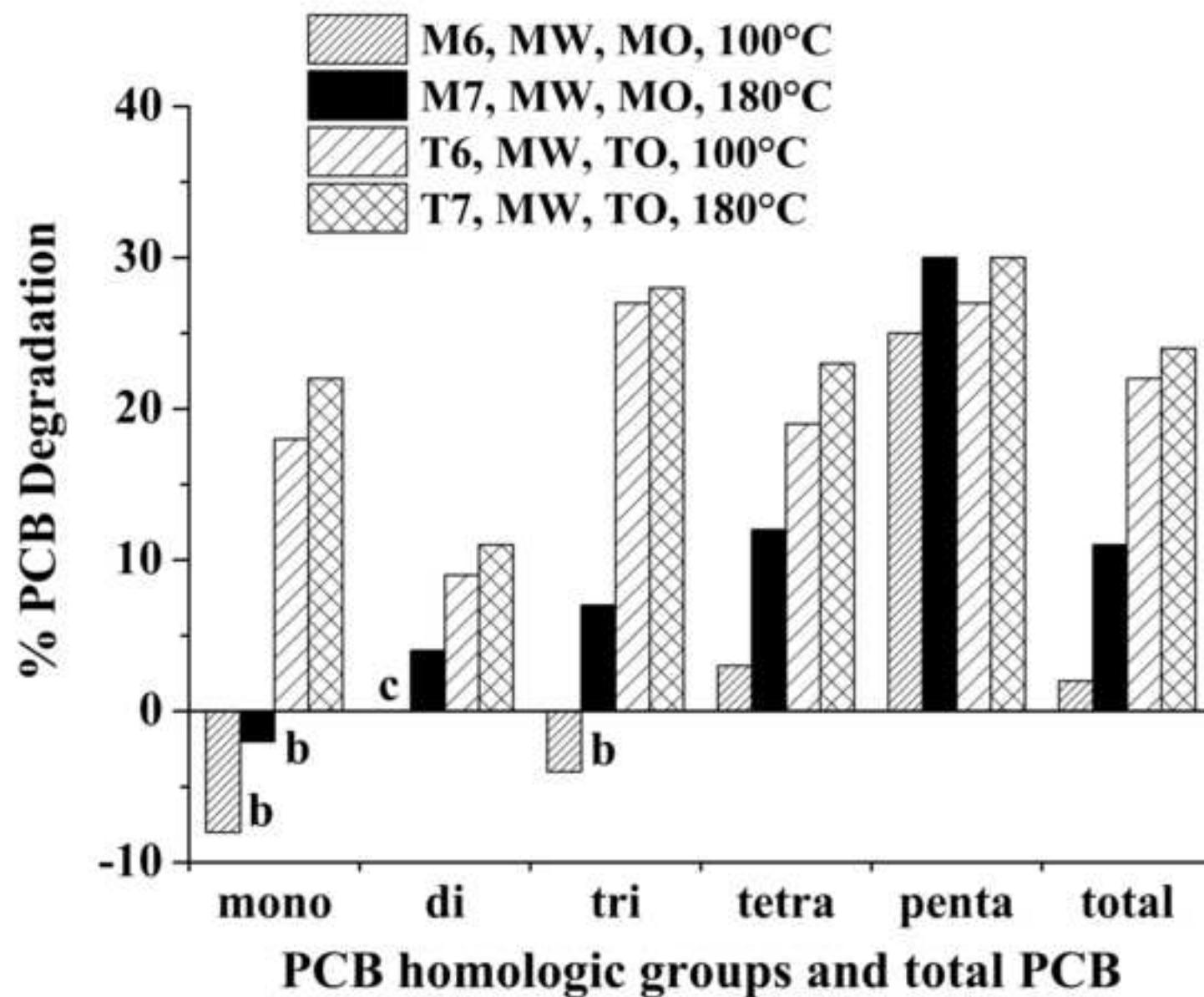
<sup>b</sup> negative values mean the increase of the concentrations of the corresponding species.



<sup>a</sup> in this case the sample is cooled using air and MW-irradiated at the same time (simultaneous cooling);

<sup>b</sup> negative values mean the increase of the concentrations of the corresponding species;

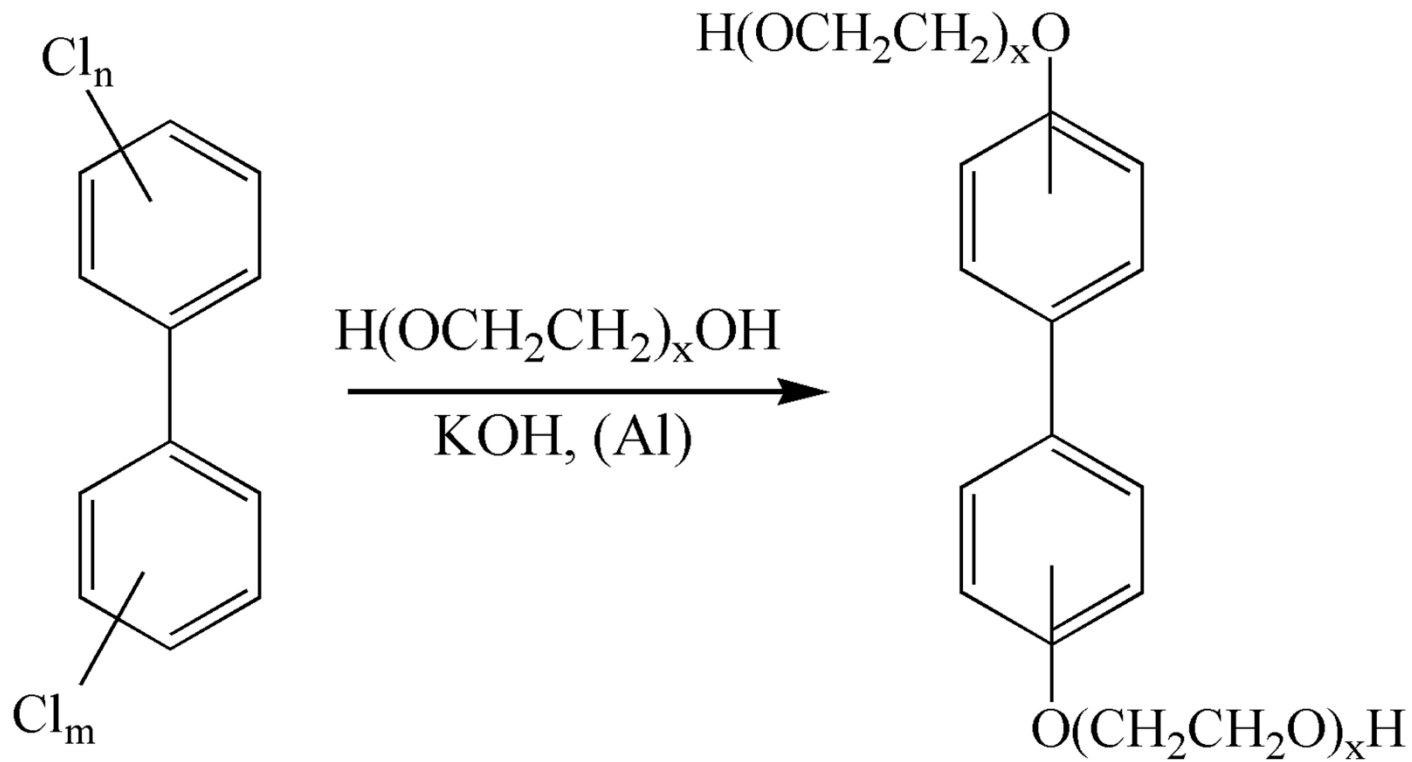
<sup>c</sup> this value is zero.



**b** negative values mean the increase of the concentrations of the corresponding species;

**c** this value is zero.

# Scheme 1



## Captions for Figures

**Figure 1:** Influence of AI adopting the traditional heating and influence of microwave irradiation compared to traditional heating on the % PCB degradation in homologous groups and % PCB total degradation in the dechlorination reaction of the model transformer oil (reaction conditions: runs T1 and T2, Table 1; run T3, PEG 8000 4.5 % w/w, KOH 3 %, 100°C, 300 W ).

**Figure 2:** % PCB degradation in homologous groups and % PCB total degradation in the dechlorination reaction of the model siloxane transformer oil: A) influence of microwave irradiation time, run T3 15 minutes, run T3L 30 minutes, and of cooling system, run T3R 15 minutes with cooling system, (reaction conditions: PEG 8000 4.5 % w/w, KOH 3 %, 100°C, 300 W); B) influence of the molecular weight of PEG, run T4 PEG 200 4.5 % w/w and run T5 Ethylene glycol 4.5 % w/w, (reaction conditions: KOH 3 %, 100°C, 15 minutes, 300 W ); C) influence of microwave irradiation at 100°C and 300 W as the unique driving force, run T6 15 minutes, run T6L 30 minutes and run T6R 15 minutes with cooling.

**Figure 3:** % PCB degradation in homologous groups and % PCB total degradation in the dechlorination reaction of the model hydrocarbon engine oil: A) influence of microwave irradiation time, run M3 15 minutes, run M3L 30 minutes, and of cooling system, run M3R 15 minutes with cooling system, (reaction conditions: PEG 8000 4.5 % w/w, KOH 3 %, 100°C, 300 W); B) influence of the molecular weight of PEG, run M4 PEG 200 4.5 % w/w and run M5 Ethylene glycol 4.5 % w/w (reaction conditions: KOH 3 %, 100°C, 15 minutes, 300 W); C) influence of microwave irradiation at 100°C and 300 W as the unique driving force, run M6 15 minutes, run M6L 30 minutes and run M6R 15 minutes with cooling.

**Figure 4:** % PCB degradation in homologous groups and % PCB total degradation in the dechlorination reaction of the model hydrocarbon engine oil and siloxane transformer oil at 100 (run M6 and T6 respectively) and 180°C (M7 and T7 respectively) after 15 minutes employing microwave irradiation as the unique driving force.

**Scheme 1:** Reaction pathway of the nucleophilic substitution mechanism in the dechlorination of PCBs.



## Supplementary Information

**S1:** Experimental conditions in the dechlorination reaction of the model **siloxane** transformer oil.

Run	Time (min)	Reagents		MW (W)	T (°C)
		glycol	KOH		
T3	15	PEG 8000 (4.5 % w/w)	3 %	300	100
T3L	30	PEG 8000 (4.5 % w/w)	3 %	300	100
T3R <sup>a</sup>	15	PEG 8000 (4.5 % w/w)	3 %	300	100
T4	15	PEG 200 (4.5 % w/w)	3 %	300	100
T5	15	Ethylene glycol (4.5 % w/w)	3 %	300	100
T6	15	-	-	300	100
T6L	30	-	-	300	100
T6R <sup>a</sup>	15	-	-	300	100

<sup>a</sup> in this case the sample is cooled using air and MW-irradiated at the same time (simultaneous cooling).

**S2:** Experimental conditions in the dechlorination reaction of the model **hydrocarbon** engine oil.

Run	Time (min)	Reagents		MW (W)	T (°C)
		glycol	KOH		
M3	15	PEG 8000 (4.5 % w/w)	3 %	300	100
M3L	30	PEG 8000 (4.5 % w/w)	3 %	300	100
M3R <sup>a</sup>	15	PEG 8000 (4.5 % w/w)	3 %	300	100
M4	15	PEG 200 (4.5 % w/w)	3 %	300	100
M5	15	Ethylene glycol (4.5 % w/w)	3 %	300	100
M6	15	-	-	300	100
M6L	30	-	-	300	100
M6R <sup>a</sup>	15	-	-	300	100

<sup>a</sup> in this case the sample is cooled using air and MW-irradiated at the same time (simultaneous cooling).

**S3:** Recovery percentages of the internal standards, PCB 19\* and PCB 156\*, for each oil sample from the clean-up procedure and the corresponding average value.

<b>Oil Sample</b>	<b>Recovery % PCB 19*</b>	<b>Recovery % PCB 156*</b>	<b>Average Recovery %</b>
TO	104	95	99
T1	67	75	71
T2	98	101	100
T3	77	90	83
T3R	99	75	87
T3L	74	91	83
T4	61	65	63
T5	65	80	72
T6	40	58	49
T6R	85	95	90
T6L	85	80	83
MO	36	73	55
M1	38	67	52
M2	93	86	90
M3	58	85	72
M3R	50	76	63
M3L	51	77	64
M4	74	90	82
M5	62	85	73
M6	53	89	71
M6R	30	66	48
M6L	34	79	57