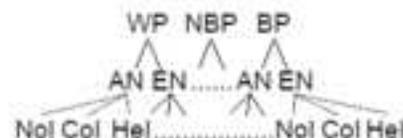




***Thinopyrum junceum*
as target**



***Sporobolus pumilus*
as target**



WP: without plastic
 BP: biodegradable plastic bag
 NBP: non-biodegradable plastic bag
 AN: ambient N deposition
 EN: elevated N deposition
 Nol: no interaction
 Col: interaction with a conspecific
 Hel: interaction with a hetero-specific

DETERIORATION STATUS OF PLASTICS



	<i>Thinopyrum junceum</i>	<i>Sporobolus pumilus</i>
COLONIZATION SUCCESS	Survival NBP < BP = WP Hel < Col = Nol	Survival NBP < BP = WP
PLANT GROWTH	Shoot biomass Col: BP < WP Nol: EN < AN Root biomass WP: EN < AN	Shoot and root biomass WP < BP Rhizome biomass Col, Hel + EN: WP < BP BP+ EN: Nol < Hel Below- to aboveground ratio BP: AN < EN

1 **Combined effect of plastic litter and increased atmospheric nitrogen deposition on vegetative**
2 **propagules of dune plants: a further threat to coastal ecosystems**

3

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13

14 **Highlights**

- 15 • Plastic litter and increased nitrogen (N) input are global environmental problems.
- 16 • Combined effect of plastic, high N and biotic condition on dune plants is unknown.
- 17 • Non-biodegradable plastics reduce *T. junceum* and *S. pumilus* survival.
- 18 • Biodegradable plastics and increased N promote *S. pumilus* growth.
- 19 • Plastics and increased N in combination could further threat dune ecosystems.

20

21

22 **Keywords:**

23 Biotic interaction; Coastal vegetation; Plastic pollution; *Sporobolus pumilus*; *Thinopyrum junceum*

24

25

26

27 **Abstract**

28 Large amounts of non-biodegradable plastics are currently deposited on beach-dune systems, and
29 biodegradable plastics could enter these already declining habitats in coming years. Yet, the impacts
30 of plastics on vegetative recruitment, a plant strategy playing a key role in dune stabilization, are
31 unknown. Whether these pollutants interact with increased atmospheric nitrogen (N) deposition, a
32 major global driver of plant biodiversity loss, in affecting plant communities of such nutrient-poor
33 habitats, and how plant-plant interactions mediate their effects need to be explored.

34 In a one-year field experiments, we examined individual and combined effects of plastic (non-
35 biodegradable, biodegradable), N deposition (ambient, elevated) and biotic condition (no
36 interaction, interaction with a conspecific or with a hetero-specific) on the colonization success and
37 growth of vegetative propagules of dune plants. *Thinopyrum junceum* and *Sporobolus pumilus* were
38 chosen as models because they co-occur along Mediterranean dunes and differ in ecological role
39 (dune- vs. non dune-building) and photosynthetic pathway (C3 vs. C4).

40 For both species, survival probability was reduced by non-biodegradable plastic and elevated N by
41 up to 100%. *Thinopyrum junceum* survival was also reduced by *S. pumilus* presence. Elevated N
42 and biodegradable plastic reduced *T. junceum* shoot biomass when grown alone and with a
43 conspecific, respectively; these factors in combination mitigated their negative individual effects on
44 root biomass. Biodegradable plastic increased *S. pumilus* shoot and root biomass, and in
45 combination with elevated N caused a greater biomass investment in belowground (root plus
46 rhizome) than aboveground organs.

47 Non-biodegradable plastic may be a further threat to dune habitats by reducing plant colonization.
48 Biodegradable plastic and increased N deposition could favour the generalist *S. pumilus* and hinder
49 the dune-building *T. junceum*. These findings highlight the urgency of implementing measures for
50 preventing plastic deposition on beaches and reducing N input.

51

52 **Capsule:** Non-biodegradable plastic litter strongly reduced the colonization of coastal dunes by
53 plants and biodegradable plastic in combination with increased nitrogen deposition favoured the
54 generalist *S. pumilus* over the typical dune-building *T. junceum* by differentially affecting their
55 growth.

56

57 **1. Introduction**

58

59 Plastic litter is a major emerging global environmental problem (Barnes et al., 2009; Maity and
60 Pramanick, 2020), and current projections suggest that coastal environments are likely to continue
61 to receive plastic waste in the future (Jambeck et al., 2015). The impact of plastics on vegetated
62 coastal habitats of conservation concern, such as seagrass beds and sand dunes (Directive
63 92/43/EEC, 1992), has only recently been addressed (Balestri et al., 2017; Menicagli et al., 2019a;
64 NOAA, 2016). Yet, whether plastic litter may interact with other global change related stressors
65 that are likely to affect these habitats needs to be explored. This issue is of high relevance, as the
66 combined effects of multiple, co-occurring environmental stressors on plant traits and processes can
67 often result in unexpected responses (Crain et al., 2008; Paine et al., 1998). In addition, how plant-
68 plant interactions could mediate the individual and combined effects of plastic and of other co-
69 occurring stressors on dune vegetation should be investigated. These interactions play a fundamental
70 role in determining plant community structure and dynamics, and they can modulate the effects of
71 environmental drivers by promoting plant survival in extreme conditions (facilitation) or replacing
72 less tolerant species to environmental changes (competition; Brooker, 2006; Arroyo et al., 2015).

73 Being located at the interface of land and sea, beaches and dunes have become “hotspots” of
74 plastic litter accumulation (Andriolo et al., 2020; Ceccarini et al., 2018; De Francesco et al., 2019;
75 Šilc et al., 2018). These systems are ecologically and economically important since they perform
76 critical ecological functions and provide valuable goods and services (Barbier et al., 2011; Drius et
77 al., 2019), but anthropogenic pressures and sea level rise are leading to a dramatic dune habitat

78 decline worldwide (Feagin et al., 2005; Malavasi et al., 2016). Dune formation is driven by the
79 interaction between sediment deposition and the establishment and spread of a few plant species
80 (Maun, 2009) adapted to cope with stressful harsh conditions (i.e. low nutrient and water
81 availability, García-Mora et al., 1999; Hesp, 1991). Plant-plant interactions can vary from
82 competitive to neutral to facilitative, depending on the stress and disturbance levels (Franks, 2003;
83 Martínez and García-Franco, 2008; Vaz et al., 2015). The accumulation of plastics made of
84 conventional non-biodegradable polymers, in particular polyethylene and polypropylene, on
85 beaches could pose a novel type of stress to dune plants (Menicagli et al., 2019a,b; Poeta et al.,
86 2017; Šilc et al., 2018). Recent studies have shown that non-biodegradable plastic bags, which are
87 currently a relevant component of the beach litter (Alshawafi et al., 2017; Andriolo et al., 2020; Šilc
88 et al., 2018), can change sediment pore-water physical and chemical parameters (e.g., pH, salinity
89 and oxygen-reduction potential) and release chemical compounds (e.g., bisphenol A) into soil,
90 inducing variations in the seedling emergence pattern of sensitive species (Balestri et al., 2019;
91 Menicagli et al., 2019a,b). Changes in the emergence and growth pattern of seedlings exposed to
92 biodegradable/compostable bags have also been observed, raising concerns about their
93 harmlessness for natural environments (Menicagli et al., 2019a,b). These bags are considered as an
94 eco-friendly alternative to non-biodegradable ones and are designed to be disposed at the end of
95 their life cycle in proper industrial facilities. But, when released in natural habitats they can cause
96 sediment and water acidification and release compounds well before being biodegraded (Balestri et
97 al., 2019; Bandopadhyay et al., 2018; Narancic et al., 2018; Serrano-Ruíz et al., 2018; Serrano-Ruíz
98 et al., 2020).

99 In addition, dune plants have to face increased deposition of nitrogen (N), one of the major
100 threats to terrestrial plant biodiversity (BassiriRad, 2015; Bobbink et al., 2003; Pakeman et al.,
101 2016). During the past century, emissions of nitrogenous atmospheric pollutants has increased
102 markedly, from an estimated N input of about 1-5 kg ha⁻¹ year⁻¹ to 20-60 kg ha⁻¹ year⁻¹ in some
103 European ecosystems (Bobbink, 1998; Galloway et al., 2008). Impacts of elevated N deposition on

104 plant communities are mainly mediated by N enrichment and changes in soil chemistry, such as
105 acidification and nutrient imbalance (Bobbink et al., 2010; Stevens et al., 2018). Enhanced
106 sensitivity to secondary environmental drivers such as extreme weather, herbivory and disease is
107 also an important mechanism (Phoenix et al., 2012; Stevens et al., 2018). In addition, there is a
108 quite variation in plant ability to respond to increased soil N availability. For example, plants with a
109 C3 photosynthetic pathway are less efficient than those with a C4 pathway in using N (Ripley et al.,
110 2010; Sage and Pearcy, 1987; Yuan et al., 2007), and the competitive balance between these plants
111 can be influenced by soil N variations through changes in the relative importance of belowground
112 and aboveground competition (Tilman, 1988). The impacts of increased N deposition are likely to
113 be greatest in nitrogen-deficient systems, such as coastal dunes, because a higher N availability can
114 change community composition by increasing the dominance of more nutrient-demanding species
115 over those stress-tolerant adapted to low nutrient conditions (Ceulemans et al., 2017; McLendon
116 and Redente, 1992 Wang et al., 2019), as already occurred along U.S. eastern and European coasts
117 (Day et al., 2018; Remke et al., 2009). However, large variations in vegetation responses to N
118 inputs have been reported among geographical areas and calcareous and acidic dunes (Aggenbach
119 et al., 2017; Bird and Choi, 2017; Day et al., 2018; Pakeman et al., 2016; Plassmann et al., 2009).
120 For European shifting dunes, the estimated critical N threshold, the lower exposure level at which
121 significant detrimental effects begin to occur, ranges from 10 to 20 kg N ha⁻¹year⁻¹ (Bobbink et al.,
122 2010). Chronic atmospheric N deposition and plastic pollution have the potential to change some
123 sediment chemical/physical parameters (Bobbink et al., 2010; Carson et al., 2010), and they will
124 probably continue to affect coastal habitats in the coming decades (BassiriRad, 2015; Stevens et al.,
125 2018; Windsor et al., 2019). Therefore, understanding whether and if so, how this novel
126 combination of stressors may affect dune plant communities may help to provide a more holistic
127 view of dune habitat vulnerability and identify priority conservation/mitigation actions.

128 In a one-year field experiment, we examined the individual and combined effects of macro-
129 plastic, N deposition (current and predicted in the Mediterranean for 2050) and biotic condition (no

130 interaction, interaction with a conspecific or with a hetero-specific neighbour) on the performance
131 (in terms of colonization success and growth) of vegetative propagules of dune plants. A traditional
132 non-biodegradable and a biodegradable/compostable plastic were chosen for testing. This latter was
133 included as the global production of biodegradable plastics is predicted to increase in the next future
134 (EU Parliament, 2018; European bioplastics, 2018) and reach the level of traditional ones. We
135 focused on vegetative propagules because the stabilization of new dune areas strongly depends on
136 their successful establishment and growth (Harris and Davy, 1986; Maun, 2009). As models, we
137 selected two clonal grasses with different photosynthetic pathway (C3 vs. C4) and ecological role
138 that can co-occur along Mediterranean dunes; *Thinopyrum junceum* (L.) Á. Love, a typical dune-
139 building C3 species (van Puijenbroek et al., 2017; Vogel et al., 1986), and *Sporobolus pumilus*
140 (Roth) P.M. Peterson & Saarela, a generalist C4 species not involved in the dune formation process
141 (Brantley et al., 2014; Pyankov et al., 2010). We hypothesized that (i) non-biodegradable and
142 biodegradable/compostable plastic would affect differentially the colonization success and plant
143 growth based on their different behaviour in natural environments and potential influence on
144 physical and chemical substrate properties observed in previous studies; (ii) elevated N deposition
145 would influence differentially the growth of the two species based on their different water and N
146 use efficiency; (iii) plastic and increased N deposition would influence interactively plant
147 performance due to their ability to alter substrate properties, and (iv) the presence of a neighbour
148 would modulate (alleviate or exacerbate) their effects, depending on the nature of plant-plant
149 interactions (facilitative or competitive).

150

151 **2. Material and methods**

152

153 *2.1. Study species*

154

155 *Thinopyrum junceum* and *S. pumilus* are perennial rhizomatous grasses inhabiting European
156 dune systems. *Thinopyrum junceum*, previously known as *Elymus farctus* (common name: sand
157 couch), is a pioneer species responsible for the early dune formation (van Puijenbroek et al., 2017).
158 *Sporobolus pumilus*, previously known as *Spartina versicolor* (common name: cordgrass), grows in
159 a range of habitats including disturbed dune depressions occasionally waterlogged by storm surge
160 (Sarmati et al., 2019) and does not substantially contribute to dune formation (Brantley et al., 2014).
161 This species was likely introduced in Europe from North Atlantic populations during the past
162 century (Sarmati et al., 2019). In Mediterranean dune systems, the two species can co-exist forming
163 the *Elymo farcti-Spartinetum Junceae* association (Sarmati et al., 2019).

164

165 2.2 Site description and experimental design

166

167 The experiment was conducted in a dune system of the Tuscany (Italy, 43°21'06.11''N
168 10°27'21.64''E) from 2018 to 2019. At this site, the average minimum and maximum monthly
169 temperature ranged from 4.3 (in January) to 29.8 °C (in August), and the cumulative precipitation
170 was approximately 670 mm during the experimental period (data from SIR Toscana-Servizio
171 Idrologico della Regione Toscana, www.sir.toscana.it). Total N deposition rate along the Tuscan
172 coast ranged from 4.7 to 5 kg N ha⁻¹ year⁻¹ (Marchetto et al., 2014). Prior to the start of the
173 experiment (April 2018), sand pH was 8.81 ± 0.01, and total N and P content was 0.084 ± 0.000
174 mg/g and 0.059 ± 0.001 mg/g, respectively (full details on sand sample collection and chemical-
175 physical analyses are provided in the Supplementary Materials).

176 At the site, a relatively homogenous area (approximately 660 m²) along embryonic shifting
177 dunes was selected based on the presence of typic Mediterranean dune plants including both the
178 study species. The area was fenced to avoid the intrusion of herbivores and humans. Before the start
179 of the experiment, the area was ideally divided in a grid of 1 m² quadrats, and plots of 25 cm x 25
180 cm were attributed to quadrats according to a completely randomized design. The average distance

181 between the nearest plots was approximately 1.25 m (minimum and maximum distances were 0.75
182 and 1.75 m respectively). The substrate within each plot was manually excavated and sieved (1 mm
183 mesh) to remove seedlings, seeds and other materials. Plots were randomly assigned to a factorial
184 combination of the following treatments; Plastic litter (P): without plastic (WP), non-biodegradable
185 bag (NBP), biodegradable/compostable bag (BP); Nitrogen (N): ambient nitrogen (AN), elevated
186 nitrogen (EN); Biotic condition (B): no interaction with a neighbour (NoI), interaction with a
187 conspecific neighbour (CoI), interaction with a hetero-specific neighbour (HeI). The “control” was
188 the condition WPxNoIxAN (Fig. 1). Each of the 18 treatment combinations (3x2x3) was replicated
189 12 times, in total 216 plots for each of the two species as the target. In plots assigned to the plastic
190 treatment, a piece of bag (25 x 25 cm), either non-biodegradable or biodegradable, was placed on
191 the bottom at a depth of 10 cm and covered with previously sieved sand. The bag size was in the
192 range of that of plastic macro-fragments commonly found along sandy shores (Vlachogianni et al.,
193 2020), and the burial depth was in the range of that of plastic fragments buried in sand dunes
194 (Ceccarini et al., 2018). Plots without plastic bags were filled with sieved sand. All the bags were
195 purchased from Italian retailers. Non-biodegradable bags were made of high-density polyethylene
196 (HDPE), and biodegradable/compostable bags were made of starch and vinyl-alcohol copolymers
197 (Mater-Bi[®]) and complied with the European standard EN13432:2000. Vegetative propagules
198 (rhizomes or stolons naturally detached from established plants by wave action) were harvested
199 from local populations and consisted of similar-sized rhizome fragments of *T. junceum* (34.6 ± 1.8
200 cm tall, mean \pm SE, with 3.5 ± 0.3 culms and 4 nodes) and *S. pumilus* (40.8 ± 2.3 cm tall with $3.8 \pm$
201 0.3 culms and 4 nodes). The size of propagules, in terms of number of nodes, was in the range of
202 that of naturally detached propagules found along sandy shores (Harris and Davy, 1986; Maun,
203 2009). To recreate different biotic interactions that the species may experience, one propagule of the
204 target species, either *S. pumilus* or *T. junceum*, was planted alone in the plot (no interaction) or with
205 another propagule of the same species (i.e., in interaction with a conspecific neighbour, *T. junceum*
206 + *T. junceum* or *S. pumilus* + *S. pumilus*) or with a propagule of the other species (i.e., in interaction

207 with a hetero-specific neighbour, *T. junceum* + *S. pumilus* for *T. junceum* as the target and *S.*
208 *pumilus* + *T. junceum* for *S. pumilus* as the target). Hence, *T. junceum* and *S. pumilus* alternated in
209 the role of the target and the hetero-specific neighbour. Propagules were planted in defined
210 positions, and the distance between propagules was about 8 cm. In plots assigned to EN, a N
211 amount of 5.2 Kg N ha⁻¹ over a background deposition was added to simulate a total N input (10.2
212 Kg N ha⁻¹year⁻¹) in the range of the atmospheric N inputs predicted in the Mediterranean for 2050
213 (10-15 kg N ha⁻¹ year⁻¹, Phoenix et al., 2006). Nitrogen was supplied as 1:1 nitrate to ammonium
214 fertilizer (NH₄NO₃, Petrokemija, Croatia). Since atmospheric N deposition may occur as pulses
215 adding a relatively small amount of N to natural systems with each pulse (Galloway et al., 2008;
216 Wang et al., 2019), the amount of fertilizer was divided in four doses that were separately applied
217 during the experimental period. Before each application, the fertilizer was dissolved in 50 mL of
218 distilled water, and the resulting solution was evenly sprayed on the plot substrate after a moderate-
219 heavy precipitation event (≥ 30 mm day⁻¹, Alpert et al., 2002; Caloiero et al., 2016) to maintain the
220 natural N input frequency. Plots assigned to AN treatment received each time an equal amount of
221 distilled water only.

222

223 2.3 Observed variables

224

225 One month after planting, each target propagule within a plot was examined *in situ* to check for
226 the presence of at least one new shoot, considered as an indicator of successful plant regeneration
227 (de la Peña et al., 2011), and to determine the probability of plant establishment (yes, no). Since in
228 some plots the target propagule failed to establish, the development of established target plants was
229 monitored on an equal number (six) of plots randomly selected among those with alive target plant
230 per each treatment. Plants were monitored weekly for the whole duration of the experiment. At the
231 end of the experiment (April 2019), the status (dead or alive) of each target plant within plots was
232 evaluated to determine the probability of plant survival, and all plants and plastic materials still

233 present within plots were harvested and transported to the laboratory. For each plastic type, a
234 subsample (20) of the retrieved materials was inspected using a stereomicroscope (Leica WILD
235 M3C, Germany) to assess the deterioration status. As all biodegradable plastics were broken in
236 small pieces, their deterioration status was estimated in terms of number and size fragments. As all
237 non-biodegradable plastics were still entire, the number of holes /perforations in each plastic was
238 counted. Plants were washed with tap water to remove sand, and those still alive were separated into
239 shoot, rhizome and root tissues and weighted after drying at 70° C for at least 72 h to determine
240 their biomass. These metrics were considered as they provide different useful indications about the
241 effects of the investigated factors on the horizontal space colonization ability (for rhizome), and
242 nutrient/water uptake and sand stabilization efficiency (for roots). The below (rhizome plus roots) -
243 to aboveground (shoot) biomass ratio was calculated as it is considered as an important indicator of
244 changes in relative resource allocation that reflects plant adaptive responses to surrounding abiotic
245 and biotic conditions (nutrient/water/space limitation and neighbour presence) (Qi et al., 2019). As
246 only few *T. junceum* plants and none of *S. pumilus* ones flowered during the study period,
247 reproductive effort was not evaluated.

248

249 2.4 Data analysis

250

251 The effects of plastic, N deposition and biotic condition on vegetative propagule recruitment and
252 plant growth were analysed separately for each of the two species as the target. The probability of
253 target propagule establishment and plant survival were analysed using generalized linear models
254 (GLMs) with binomial error distributions and a logit link due to the binary nature of these variables
255 (yes, no). For each variable, the statistical significance of fixed terms in the model was assessed
256 with a Likelihood-ratio Test (LRT) of the change in deviance between a model with and a model
257 without the term of interest (Zuur et al., 2009). The results of the models retaining those terms
258 identified as being significant to the model according to LRT were reported in the Supplementary

259 Materials. When significant effects were detected in the analyses, multiple post-hoc Tukey's HSD
260 tests were used for pairwise comparisons of means.

261 To assess the effects the experimental factors on the growth of each species, separate univariate
262 three-ways ANOVAs were performed on each variable. These analyses were carried out on three
263 replicates, as in some treatments only three plants survived at the end of the experiment. In
264 treatments with more alive plants, the three plants were selected at random. As few plants survived
265 on buried non-biodegradable bags, the level non-biodegradable of the factor Plastic was excluded
266 from the analyses for both the species. In addition, as few target *T. junceum* plants survived in the
267 presence of biodegradable bag and hetero-specific neighbour, for this species as the target the level
268 hetero-specific interaction of the factor Biotic condition was excluded from the analyses. Post-hoc
269 Student-Newman-Keuls (SNK) tests were used to identify differences among levels of significant
270 factors. Prior to ANOVA analyses, data were checked for normality and homogeneity of variances
271 by using Shapiro-Wilk test and Cochran C test, respectively, and transformed to meet assumptions
272 when necessary. For *T. junceum* rhizome biomass, no transformation was effective in removing the
273 non-normal error distribution, and ANOVAs were performed on untransformed data as this analysis
274 is robust to departures from the normality assumption (Underwood, 1997).

275 Generalized linear models were carried out in R software (v. 3.5.1; R Development Core Team,
276 2018) by using the glm function within the stats package and the lrtest function within lmtest
277 package (Zeileis and Hothorn, 2002) while *post-hoc* Tukey's HSD tests were conducted by using
278 the glht function within the multcomp package (Hothorn et al., 2008). ANOVA analyses were
279 performed by using the lm function within the GAD package of R environment (Sandrini-Neto and
280 Camargo, 2014).

281

282 **3. Results**

283

284 *3.1 Plastic deterioration status*

285

286 All non-biodegradable plastics retrieved from sand at the end of the experiment were still entire and
287 showed various holes (28 ± 3 , mean \pm SE) with a diameter lower than 1 mm caused by root
288 perforation. Instead, all biodegradable plastics were broken in smaller pieces (mean 24 ± 2 , mean \pm
289 SE) of various size (from 0.6 cm to 9 cm), and most of them adhered to roots (Fig. S1).

290

291 3.2 *Colonization success*

292

293 One month after planting, the percentage of propagules of *T. junceum* and *S. pumilus* that had
294 established a new plant ranged from 50 to 100% across all the treatments (Fig. 2a,b). For both the
295 species, the probability of establishment was not significantly influenced by any factors or their
296 interactions (Tables S1). However, many established plants died in summer (June-July, Table S3).
297 For both the species as target, the probability of survival at the end of the experiment was
298 negatively affected by the main factor Plastic. The reduction of survival was larger for plants
299 established on NBP compared to those grown on BP and without plastic (Fig. 2c,d; Tables S1, S2).
300 The probability of target survival was also influenced by the main effect Biotic condition for *T.*
301 *junceum* and by N deposition for *S. pumilus*. For the former species, the likelihood of survival was
302 lower when grown with a *S. pumilus* than with a conspecific neighbour regardless of plastic and N
303 treatments (Fig. 2c; Table S2). For the latter species, the chance of survival was lower at the EN
304 deposition than at the AN deposition (Fig. 2d; Table S2). This effect appeared to be influenced by
305 NBP as the exclusion of this level of the factor Plastic from the survival analysis did not show any
306 significant effect of N (Table S4).

307

308 3.3 *Plant growth*

309

310 The low number of plants survived in the presence of NBP did not allow us to compare their
311 biomass variables with those of plants exposed to the other treatments. However, at a visual
312 inspection the root system of these plants considerably differed from that of plants grown in WP
313 and BP treatments, being it constituted by a tightly packed system of thin roots developed in the
314 sand layer just above the bag fragment (Fig. S1).

315 ANOVA on shoot and root biomass of *T. junceum* detected a significant effect of the interaction
316 between N deposition and Biotic condition. Specifically, plants grown alone had smaller biomasses
317 at EN than at AN deposition, and those grown at AN had smaller biomasses in the presence of a
318 conspecific than alone (Fig. 3a,c; Table 1). For shoot biomass, there was also a significant
319 interaction between Plastic and Biotic condition. In the presence of BP, plants established with a
320 conspecific were smaller than those grown alone (Fig. 3a; Table 1). Instead, for root biomass there
321 was also a significant interaction between Plastic and N deposition. Without BP, root biomass was
322 smaller at EN than at AN deposition, and in this latter condition it was smaller in the presence than
323 in the absence of BP (Fig. 3c; Table 1). No significant differences for rhizome biomass and below-
324 to aboveground biomass ratio were detected among treatments (Fig. 3b,d; Table 1).

325 For *S. pumilus*, shoot biomass was significantly affected by the main factor Plastic and Biotic
326 condition (Table 1). Shoot biomass was greater in plants grew with than without BP, and it was also
327 greater in plants grew with a hetero-specific or alone than with a conspecific neighbour (Fig. 4a;
328 Table 1). Root biomass was significantly affected by the main factor Plastic resulting greater with
329 than without BP (Fig. 4c; Table 1). For rhizome biomass, there was a significant interaction among
330 Plastic, N deposition and Biotic condition. At EN deposition, the rhizome biomass of plants grown
331 with a neighbour, either conspecific or hetero-specific, was greater in the presence than in the
332 absence of BP. In addition, with BP and a conspecific neighbour rhizome biomass was larger under
333 EN than AN deposition (Fig. 4b; Table 1). At EN deposition, the rhizome biomass of plants grown
334 in the presence of BP with a hetero-specific was also larger than that of plants grown alone while
335 the opposite effect was found in plants grown without BP (Fig. 4b; Table 1). The interaction

336 between Plastic and N deposition had a significant effect on the below-to aboveground biomass
337 ratio. At EN, a higher ratio was observed in plants grown with than without BP, and when plants
338 grew with BP the ratio was higher at EN than AN deposition (Fig. 4d; Table 1).

339

340 **4. Discussion**

341

342 Plastic litter is a global-scale problem, and the number of studies revealing the negative impact of
343 this pollutant on coastal and marine animals has increased rapidly in the last decade (Bergmann et
344 al., 2015). Yet, plants have received little attention, although the crucial role they play in building,
345 stabilizing and maintaining coastal dune ecosystems. Our study is the first providing experimental
346 evidence of the individual and combined effects of macro-plastic (either non-biodegradable or
347 biodegradable), increased N deposition and biotic interaction on dune habitat colonization by
348 vegetative propagules of dune plants.

349

350 *4.1. Colonization success*

351

352 For both the two species as target, the establishment of propagule was not influenced by any of
353 the investigated factors or their interactions. This could be due to the ability of vegetative
354 propagules of dune plants to regenerate new plants using carbohydrates reserves stored in their
355 rhizome (Maun, 2009). However, the survival of newly established plants was influenced by plastic.
356 The survival probability of plants established on NBP resulted up to three times lower compared to
357 that of plants grown without NBP, confirming our hypothesis about the importance of this pollutant
358 for dune plants. These negative effects could be due to the well-known long durability and
359 mechanical resistance of the polymer HDPE (Chamas et al., 2020) and the inability of plant roots to
360 develop enough mechanical forces to fragment this material and acquire essential belowground
361 resources for sustaining growth. Another possible explanation could be a reduction of water

362 availability in soil caused by increased evaporation rate and more negative water potentials induced
363 by the presence of buried plastic (De Jong, 1979). This is because most of plants died during the
364 warmest months (June and July, Table S3). Other changes of sediment chemical-physical properties
365 that could be due to the persistence of this plastic, for example soil pH, electrical conductivity,
366 temperature and water permeability (Carson et al., 2011; Qi et al., 2020; Steinmetz et al., 2016;
367 Zhang et al., 2020), cannot be excluded. The development of an extensive but shallow root system
368 observed in the few plants survived in the presence of NBP indicates that probably only the most
369 vigorous plants might be able to circumnavigate the bag obstacle by reorienting root growth, an
370 avoidance strategy adopted by some species (Santisree et al., 2012). However, these plants could be
371 less efficient in trapping and consolidating sediment, and also more susceptible to physical
372 disturbances, such as sand erosion and uprooting by wind, at least during their first year of growth,
373 because of their shallow root system (Balestri and Lardicci, 2013). As the occurrence of macro-
374 plastics on beaches can be relevant, up to 1 fragment (from 2.5 to 50 cm in size) per linear meter
375 (Vlachogianni et al., 2020), the negative effects of buried NBPs on plant survival could translate
376 into the formation of numerous bare sand areas. This reduction in vegetation cover could affect
377 dune mobility and promote under certain circumstances the formation of blowouts which increases
378 the vulnerability of the dune systems to storms.

379 The absence of any substantial negative effects of BP on plant survival could be explained by the
380 poor resistance of Mater-Bi[®] to perforation by roots and its fragmentation in small pieces unable to
381 hinder root expansion. This finding is not in agreement with the negative effects caused by this type
382 of bag on the emergence of seedlings of dune plants, including *T. junceum*, reported in a previous
383 study conducted close to our experimental area (Menicagli et al., 2019b). However, these
384 discrepancies could be ascribed to the greater size and vigour of vegetative propagules than
385 seedlings and the position of the bag respect to plant material (below vs. above plastic).

386 The lower survival probability of *T. junceum* when grown with *S. pumilus* than when with a
387 conspecific regardless of plastic and N deposition could be probably due to a greater competitive

388 ability for local resources and/or larger size of this latter species (on average 10 g vs. 5 g dry weight
389 of total biomass). Indeed, it has been shown that this C4 species under some circumstances may
390 completely replace C3 species such as *Calamagrostis arenaria* (L.) Roth subsp. *arundinacea*
391 (Husn.) Banfi, Galasso & Bartolucci (Fiorentin, 2007; Sarmati et al., 2019). As a result of the
392 effects of the presence of *S. pumilus* and NBP, on average only 8.3% of the established *T. junceum*
393 plants survived. Instead, the lower survival probability observed in *S. pumilus* when grown at EN
394 than at AN condition could be mainly related to drought stress. Indeed, high N availability can
395 increase the growth but also water consumption of plants (Liang et al., 2020), and in C4 species this
396 could lead to metabolic limitations and a down-regulation of photosynthetic efficiency (Ripley et
397 al., 2007; Ripley et al., 2010). The negative effect of N on plant survival appeared to be exacerbated
398 by the presence of NBP, as on average only 5.5% of the initially established *S. pumilus* plants
399 survived as a result of the combination of EN deposition and NBP.

400

401 4.2 Plant growth

402

403 For most plant variables, the effects of plastic and N deposition were mediated by biotic
404 conditions, and the direction of the effects on the two species was generally opposite, i.e., negative
405 for *T. junceum* and positive for *S. pumilus*. At AN deposition rate, the shoot biomass of both *T.*
406 *junceum* and *S. pumilus* plants were smaller when grown with a conspecific than as an isolated plant
407 regardless of BP suggesting intraspecific competition for essential resources. The reduction of root
408 biomass of *T. junceum* plants grown with BP compared to that of plants grown without plastic is in
409 agreement with results of previous studies showing a reduced root growth in *T. junceum* seedlings
410 (Menicagli et al., 2019b) and in crop species exposed to leachates obtained from Mater-Bi[®] plastics
411 (Serrano-Ruíz et al., 2018). Negative effects on roots have also been observed in crop species
412 grown in soils containing a biodegradable mulch made of starch-based plastic (Qi et al., 2018) like
413 Mater-Bi[®]. As *T. junceum* generally requires alkaline sandy substrates to grow (approximately at

414 pH 8; Greipsson, 2011; Maun, 2009), the negative effect on roots observed here could be related to
415 a reduction of sand pH, a phenomenon observed following the contact of BP with soil and water
416 (Menicagli et al., 2019a; Sintim et al., 2019), as well as to the release of degradation compounds
417 from the bag into the sand layer below roots and to alterations of microbial activity. The reduction
418 of shoot biomass observed in plants grown with a conspecific could be a result of a lower root
419 biomass production combined with intraspecific competition for aboveground resources. In
420 contrast, in *S. pumilus* BP increased biomass of roots and shoots. These positive effects indicate that
421 BP might have alleviated plants from drought for example by increasing sand moisture availability,
422 being Mater-Bi® a hydrophilic material (Milionis et al., 2014). Since *S. pumilus* grows better in
423 acid-neutral soil (pH from 4.5 to 7.1; Lonard et al., 2010), this species might also have benefited
424 from sand acidification following incorporation of BP in sand.

425 At the increased N deposition rate predicted in Mediterranean for the next decades, *T. junceum*
426 plants grown alone did not differ from those grown with a conspecific in roots and shoots biomass
427 while at AN plants grown alone had larger biomasses. This suggests that the supplied N might have
428 negatively impacted the growth of isolated plants and that the presence of the conspecific might
429 have mitigated these effects, probably by reducing the amount of N available per individual plant.
430 The negative effects of EN observed here are in agreement with the decreased standing biomass and
431 fine root production reported in other C3 species following N addition (White et al., 2012). They
432 could be related to changes in soil properties linked to the higher N input for example decreased soil
433 pH, accumulation of soil ammonium and lower efficiency in the uptake of N (especially nitrate) by
434 *T. junceum* as observed in some grassland species (Britto and Kronzunker, 2002; Brooker, 2006;
435 Chen et al., 2015). Indeed, it has been shown that a lower soil pH can suppress the growth of some
436 species and belowground microbial communities by increasing the concentrations of H⁺ and Al³⁺
437 and decreasing those of base mineral cations (e.g., Ca²⁺, Mg²⁺, Na⁺), and hence limiting nitrate
438 uptake by plants (Chen et al., 2015; Stevens et al., 2010). In contrast, EN resulted in an increased
439 rhizome biomass in *S. pumilus* but only when grown as an isolated plant. This suggests that this

440 species might have benefited from the higher N supply and that interspecific competition for
441 resources might have cancelled out positive N effects. The greater accumulation of rhizome
442 biomass was partially in accordance with the results of previous studies showing that in nitrogen-
443 limited environments increased N inputs can enhance belowground carbon allocation of C4 plants
444 and alter biomass allocation by affecting above- and belowground productivity (Chen et al., 2015;
445 Ripley et al., 2007; Sage and Pearcy, 1987; Wang et al., 2019; Zheng and Ma, 2018).

446 At the increased N deposition rate predicted in Mediterranean and in the presence of BP, the root
447 biomass of *T. junceum* was reduced by approximately 40% compared to plants grown at ambient N
448 deposition without BP. However, this reduction was less than that caused by the two factors when
449 they acted in isolation, suggesting an antagonistic interaction of negative type according to the
450 directional classification system proposed by Piggott et al. (2015). This could be due to the
451 mitigation of the negative individual effects of the two factors on chemical/physical sand properties
452 and a change in the balance between nitrogenous compounds. In contrast, the combined effect of
453 BP and EN on below- to aboveground biomass ratio of *S. pumilus* was positive although the effect
454 of each individual stressor was not statistically significant suggesting a synergistic interaction of
455 positive type (Piggott et al., 2015) between these factors. This higher below-to aboveground
456 biomass ratio was mainly due to an increased biomass allocation to belowground compartment. In
457 addition, we found that the effects of BP and EN on the rhizome biomass of *S. pumilus* were
458 mediated by biotic condition. The rhizome biomass of *S. pumilus* plant as the target grown with *T.*
459 *junceum* at EN and BP was six-times greater than that of plants grown alone while it was four-times
460 smaller than that of plants established alone without plastic. This could indicate that the presence of
461 BP reversed the effect of EN on rhizome and possibly shifted interspecific interactions from
462 competition to facilitation. Unfortunately, we could not examine the effect of *S. pumilus* on biomass
463 accumulation and allocation pattern of *T. junceum* as the target, as there were not enough replicates
464 for this treatment to perform statistical analyses. However, the overall low survival of *T. junceum* in
465 the presence of *S. pumilus* suggests that this former species would not benefit from the latter one.

466 Collectively, these results indicate that BP and EN could either exacerbate or mitigate their
467 individual effect on the development of belowground organs of plants depending on the species and
468 neighbour identity, favouring the spread of the generalist grass *S. pumilus* and hindering that of the
469 dune-building *T. junceum*. This could have relevant consequences for dune system health leading to
470 reduced dune accretion and sediment stability. This because *S. pumilus* does not substantially
471 contribute to dune formation by accumulating sand, as instead *T. junceum* does, and it also enhances
472 the vulnerability of dunes to repeated disturbances like over-wash (Brantley et al., 2014). Although
473 we have suggested several potential explanations, more research is needed to elucidate the specific
474 physiological and/or environmental drivers behind the responses to BP and EN deposition observed
475 here.

476

477 **5. Conclusions**

478

479 At a global scale, coastal dune ecosystems play a prominent role in coastal defence, wind and
480 aerosol protection and biodiversity support. However, these highly valuable habitats are threatened
481 by natural and anthropogenic factors including plastic pollution and N deposition. Our study reveals
482 that non-biodegradable plastic litter could be a further threat to coastal dunes as it could
483 dramatically reduce the colonization success by vegetative propagules of both dune-building and
484 generalist species by up to 100% at EN deposition rate. It also demonstrates that BPs individually
485 and in combination with EN and neighbour presence have the potential to affect differentially the
486 growth of dune plant species and then their interactions. Importantly, our findings underline the
487 relevance of better investigating the potential risks associated with the introduction of plastics in
488 coastal dune environments and suggest that assessing the nature (additive and non-additive) and the
489 direction (synergistic or antagonistic) of interactions among plastic litter, plant-plant interactions
490 and global environmental stressors in future studies could improve our ability to predict the real
491 impact of this pollutant and identify adequate interventions and restoration actions. Overall,

492 lowering N deposition and setting up effective management actions aimed at reducing the entering
493 of plastic waste in coastal environments and removing all plastics accumulated on beaches are
494 urgently needed to mitigate the impact on coastal dune systems. Given the predicted global rise of
495 the production of new generation of biodegradable/compostable plastics (EU Parliament, 2018;
496 European bioplastics, 2018), also the implementation of more effective postconsumer management
497 actions is fundamental to prevent their release into natural environments.

498

499 **Acknowledgements**

500

501 We sincerely thank Sara Corti for her support in the set-up of the field experiment, and Rachele
502 Cardella for her assistance in plant collection. We also thank the Municipality of Rosignano
503 Marittimo (Italy) and Solvay Chimica Italia SPA of Rosignano Solvay (Italy) for providing
504 technical assistance during the study. We are grateful to four anonymous reviewers for their helpful
505 suggestions which improved the quality of the manuscript. This work is part of the PhD research
506 project of Virginia Menicagli, and it was funded by *Fondi di Ateneo* (FA) and *Progetti di Ricerca di*
507 *Ateneo* (PRA) of University of Pisa (Italy). The authors declare no conflict of interest.

508

509 **CRedit authorship contribution statement**

510

511 **Virginia Menicagli:** Conceptualization, Formal analysis, Investigation, Visualisation, Writing-
512 Original draft. **Elena Balestri:** Conceptualization, Formal analysis, Investigation, Visualisation,
513 Writing- Original draft. **Flavia Vallerini:** Formal analysis, Investigation, Visualisation. **Alberto**
514 **Castelli:** Supervision, Writing- Review & Editing. **Claudio Lardicci:** Conceptualization,
515 Investigation, Supervision, Writing- Review & Editing.

516

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811 **Figure legends**

812 **Fig. 1.** Experimental design.

813 Double-column fitting image

814 Colour image only in online version

815 **Fig. 2.** Probability of propagule establishment (a, b) and plant survival (c, d) of *Thinopyrum*

816 *junceum* (left panels) and *Sporobolus pumilus* (right panels) as the target grown without a neighbour

817 (NoI) or with a conspecific (CoI) or a hetero-specific (HeI) neighbour in the presence of non-

818 biodegradable plastic bag fragment (NBP), or biodegradable/compostable plastic bag macro-

819 fragment (BP) or without plastic (WP) at ambient and elevated nitrogen (N) deposition. Mean \pm se,

820 $n = 12$ (a,b), $n = 6$ (c,d).

821 Double-column fitting image

822 Colour image only in online version

823 **Fig. 3.** Shoot biomass (a), rhizome biomass (b), root biomass (c) and below- to aboveground

824 biomass ratio (d) of *Thinopyrum junceum* as the target grown without a neighbour (NoI) or with a

825 conspecific (CoI) neighbour in the presence of biodegradable/compostable plastic bag macro-

826 fragment (BP) or without plastic (WP) at ambient and elevated nitrogen (N) deposition. Mean \pm se,

827 $n = 3$.

828 Double-column fitting image

829 Colour image only in online version

830 **Fig. 4.** Shoot biomass (a), rhizome biomass (b), root biomass (c) and below- to aboveground

831 biomass ratio (d) of *Sporobolus pumilus* as the target grown without a neighbour (NoI) or with a

832 conspecific (CoI) neighbour or with a hetero-specific (HeI) neighbour in the presence of

833 biodegradable/compostable plastic bag macro-fragment (BP) or without plastic (WP) at ambient

834 and elevated nitrogen (N) deposition. Mean \pm se, $n = 3$.

835 Double-column fitting image

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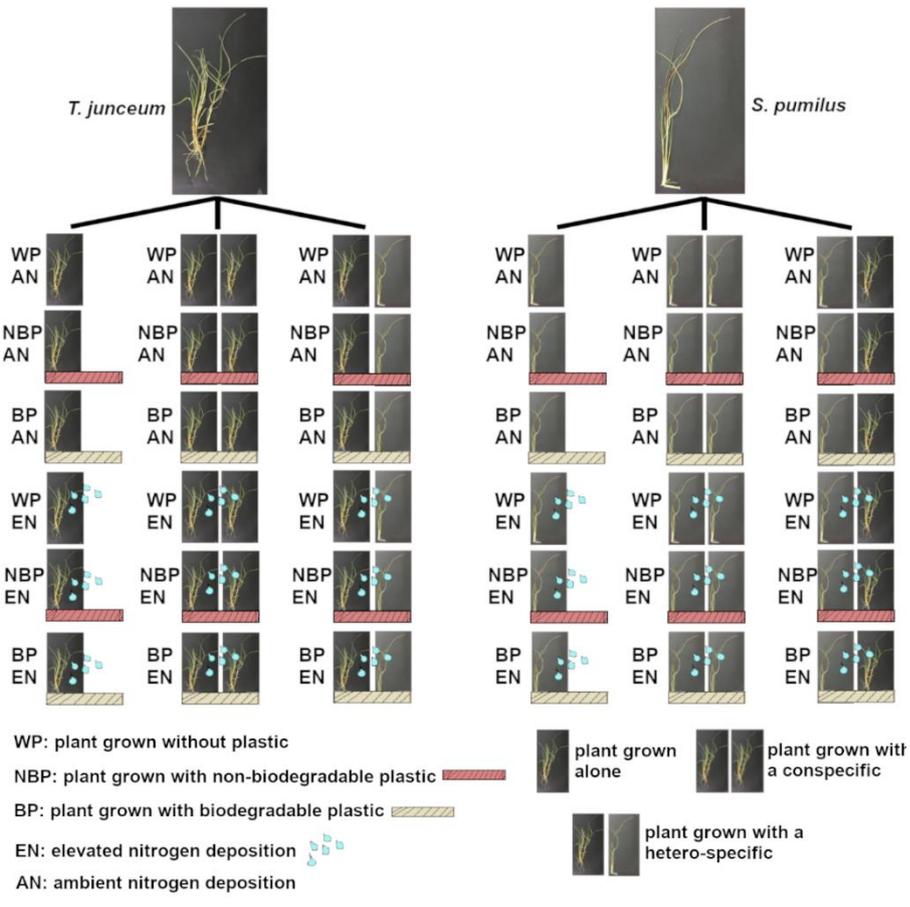
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867 **Fig.1**

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887 **Fig.2**

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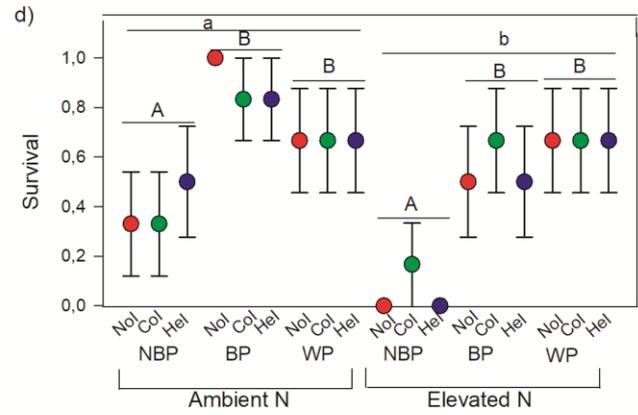
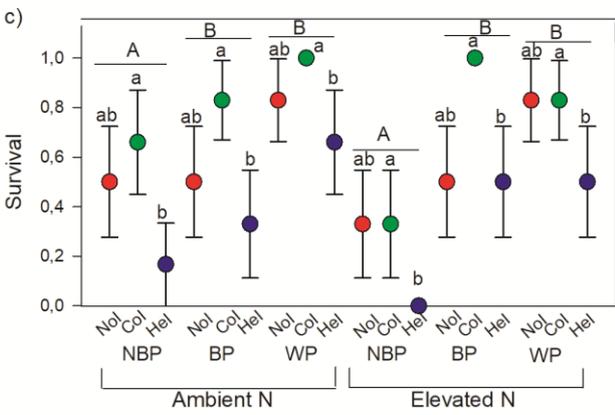
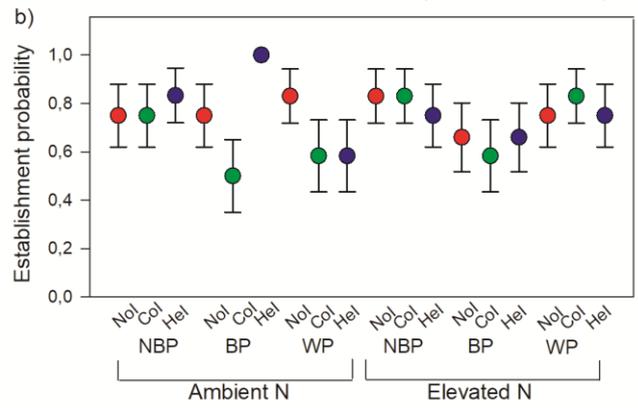
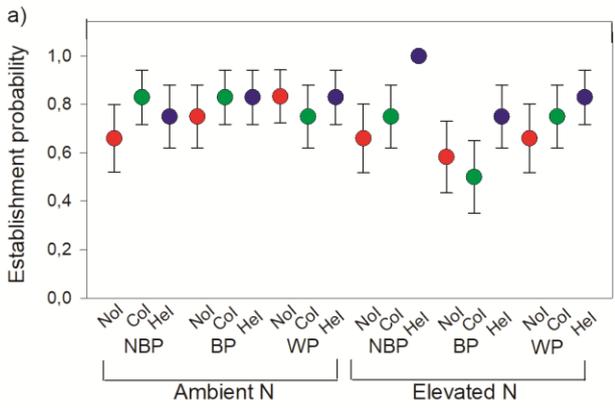
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- No Interaction (NoI)
- Conspecific Interaction (Col)
- Hetero-specific Interaction (Hel)



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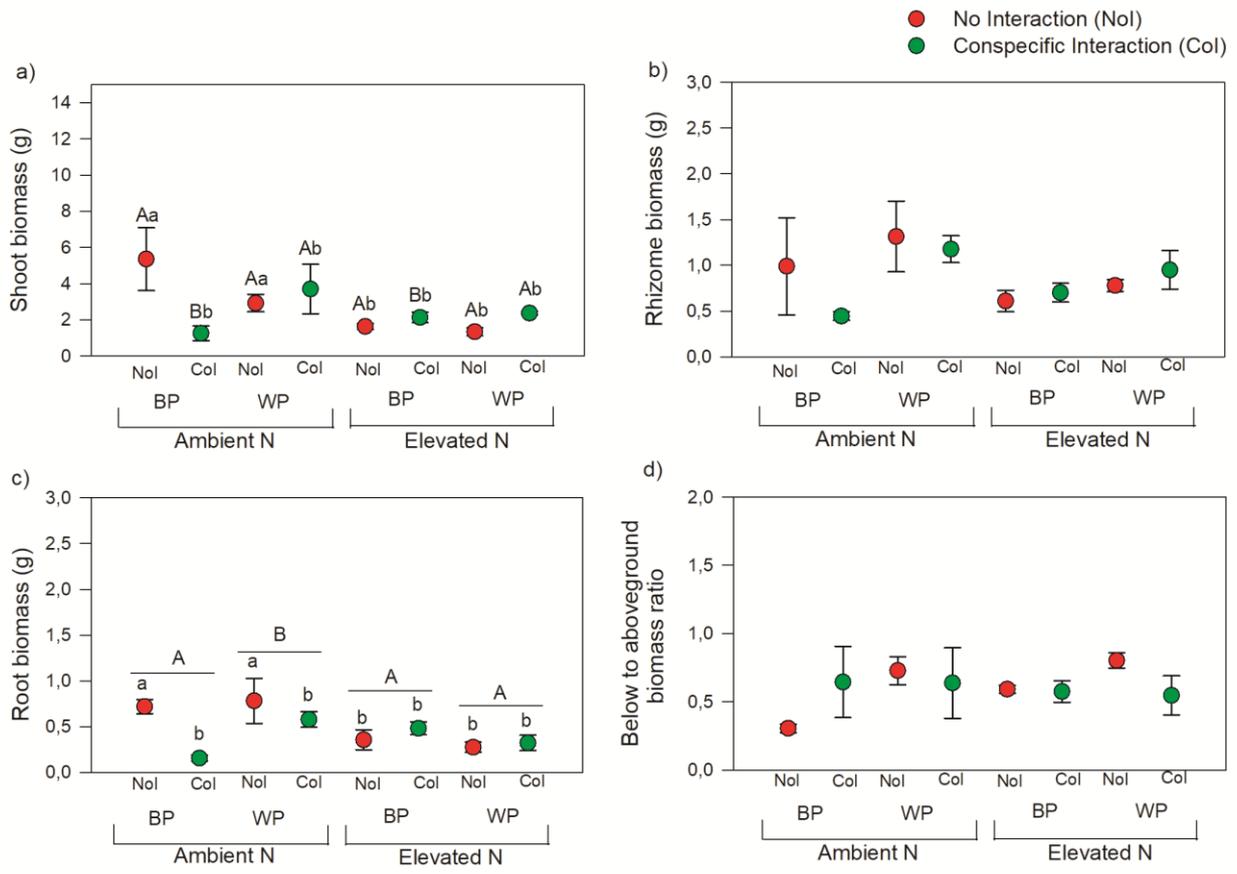


Fig. 3

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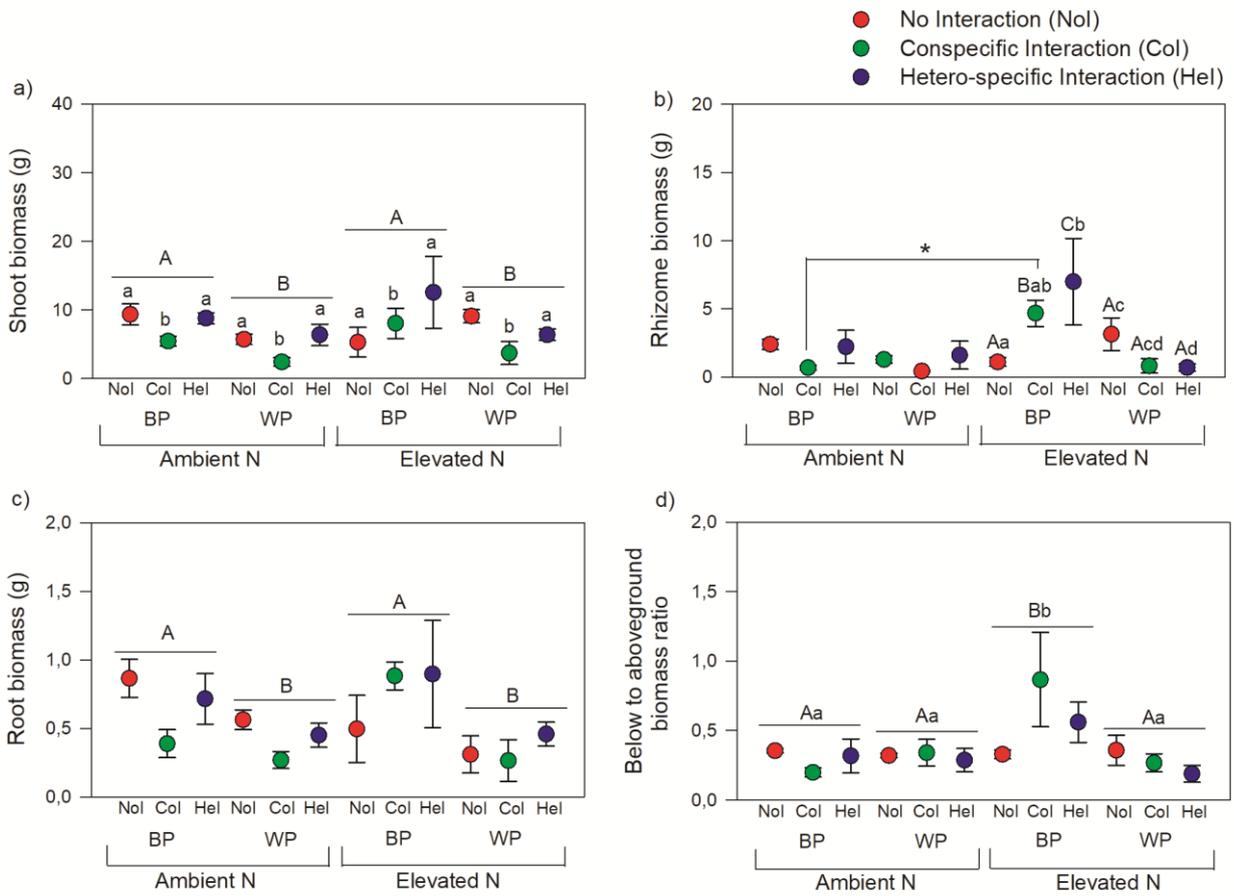


Fig. 4

940 **Table 1.** Results of ANOVA analyses testing the effects of plastic litter (without plastic,
 941 biodegradable/compostable bag macro-fragment), nitrogen level (ambient, elevated) and biotic condition
 942 (no interaction, interaction with a conspecific or a hetero-specific neighbour) on plant variables for
 943 *Thinopyrum junceum* (a) and *Sporobolus pumilus* (b). For *T. junceum* the level interaction with a hetero-
 944 specific was excluded from the analysis due to the low number of survived plants. Bold values indicate
 945 significance at $p < 0.05$. Data transformations and results of SNK test are reported.

946 a) <i>T. junceum</i>		Shoot biomass		Rhizome biomass		Root biomass		Below- to above ground biomass		
947										
948	Source	df	F	p	F	p	F	p	F	p
949	Plastic (P)	1	0.39	0.540	2.34	0.145	0.70	0.414	2.01	0.174
950	Nitrogen (N)	1	3.73	0.071	2.89	0.108	5.22	0.036	0.22	0.641
951	Biotic cond (B)	1	0.46	0.507	0.85	0.369	3.56	0.077	0.00	0.957
952	P x N	1	0.74	0.400	0.41	0.528	7.01	0.017	0.30	0.587
953	P x B	1	6.33	0.022	0.40	0.533	1.69	0.210	2.52	0.131
954	N x B	1	8.53	0.010	2.38	0.141	10.38	0.005	1.52	0.234
955	P x N x B	1	2.67	0.121	0.09	0.765	3.63	0.074	0.20	0.655
956	Residual	16								
957										
958	Transformation		Log(x)		x^2		Sqrt(x)			
959	SNK test		CoI: BP < WP				AN: BP < WP			
960			BP: CoI < NoI				WP: EN < AN			
961			NoI: EN < AN				NoI: EN < AN			
962			AN: CoI < NoI				AN: CoI < NoI			

963 b) <i>S. pumilus</i>		Shoot biomass		Rhizome biomass		Root biomass		Below- to above ground biomass		
964										
965	Source	df	F	p	F	p	F	p	F	p
966	Plastic (P)	1	4.57	0.042	7.91	0.009	9.27	0.005	3.72	0.065
967	Nitrogen (N)	1	0.18	0.674	2.35	0.138	0.11	0.740	1.60	0.217
968	Biotic cond (B)	2	4.44	0.022	1.91	0.168	1.26	0.299	0.33	0.720
969	P x N	1	1.21	0.281	0.93	0.342	0.74	0.395	6.85	0.015
970	P x B	2	2.78	0.081	2.89	0.074	0.33	0.715	0.85	0.438
971	N x B	2	0.57	0.567	1.70	0.202	2.96	0.070	1.20	0.316
972	P x N x B	2	1.43	0.258	3.88	0.034	0.88	0.426	2.25	0.126
973	Residual	24								
974										
975	Transformation		Log(x)		Log(x)		Sqrt(x)		Log(x)	
976	SNK test		WP < BP		BP-CoI: AN < EN		WP < BP		EN: WP < BP	
977			CoI < NoI = HeI		CoI,HeI-EN:				BP: AN < EN	
978					WP < BP					
979					BP-EN: NoI < HeI					
980					WP-EN: HeI < NoI					

981 WP: without plastic; BP: biodegradable plastic; NBP: non-biodegradable plastic; NoI: no neighbour or no interaction;
 982 CoI: plant grown with a conspecific; HeI: plant grown with a hetero-specific; AN: ambient level of nitrogen; EN:
 983 elevated level of nitrogen

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986 **Supporting Information for**

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988 **Combined effect of plastic litter and increased atmospheric nitrogen deposition on vegetative**
989 **propagules of dune plants: a further threat to coastal ecosystems**

990

991 by Virginia Menicagli, Elena Balestri*, Flavia Vallerini, Alberto Castelli, Claudio Lardicci

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995 **Supporting Information included in this file:**

- 996 • **Supplementary methods:** Sand sample collection and determination of total N and P
997 content and pH.
- 998 • **Table S1:** Summary of the generalized linear models testing the significance of the effects
999 of plastic, nitrogen deposition and biotic condition and their interactions on the probability
1000 of propagule establishment and plant survival of *Thinopyrum junceum* and *Sporobolus*
1001 *pumilus* as target.
- 1002 • **Table S2:** Summary of the generalized linear models testing the effects of plastic, nitrogen
1003 deposition and biotic condition and selected interactions on the probability of plant survival
1004 of *Thinopyrum junceum* and *Sporobolus pumilus* as target.
- 1005 • **Table S3:** Percentage of alive plants of *Thinopyrum junceum* and *Sporobolus pumilus* as
1006 target under different treatment combinations observed from May 2018 to the experiment
1007 end (April 2019).
- 1008 • **Table S4.** Summary of the generalized linear models testing the significance of the effects
1009 of plastic (biodegradable plastic vs no plastic), nitrogen deposition and biotic condition and
1010 their interactions on the probability of plant survival of *Sporobolus pumilus* as target.

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- **Fig. S1.** One-year old plants of *Thinopyrum junceum* and *Sporobolus pumilus* established from vegetative propagules in the presence of a non-biodegradable plastic bag macro-fragment or with biodegradable/compostable plastic bag macro-fragment or in bare sand without plastic (control).

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1037 **Supplementary methods: Sand sample collection and determination of total N and P content**
1038 **and pH.**

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1040 Before the experiment, sediment samples were collected using a plastic corer (3 cm diameter x10
1041 cm depth) in three randomly chosen areas at the study site (n = 4 samples per area) and transported
1042 in a refrigerator (at 5 °C) to the laboratory for analyses of N and phosphorus (P) content and pH.
1043 Samples from each area were pooled, air dried and sieved (2 mm mesh size) before chemical
1044 analyses. Total N was determined by Kjeldahl method (Bremmer et al., 1982) through a Kjeldahl
1045 distillation system (Vapodest, Gerhardt) and total P through mineralization followed by a
1046 colorimetric determination of phosphormolybdate complex (Violante and Adamo, 2000) through a
1047 spectrophotometer (Lambda 25, Perkin Elmer). Sediment pH was measured electrometrically using
1048 a pH meter in a solution of soil:demineralized water (1:2.5 w/v)(Mc Lean, 1982).

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1062 **Table S1.** Summary of the generalized linear models testing the significance of the effects of
1063 plastic, nitrogen (N) deposition and biotic condition and their interactions on the probability of
1064 propagule establishment and plant survival of *Thinopyrum junceum* and *Sporobolus pumilus*.
1065 Results of generalized linear fixed effects models testing the significance of the effects of plastic
1066 (without plastic, with non-biodegradable or biodegradable/compostable bag), nitrogen level
1067 (ambient, elevated) and biotic condition (no interaction, interaction with a conspecific or a hetero-
1068 specific neighbour) and their interactions on the probability of propagule establishment and plant
1069 survival of *Thinopyrum junceum* and *Sporobolus pumilus* as target. Significant results ($p < 0.05$)
1070 are in bold.

		Establishment				Survival					
		<i>T. junceum</i>		<i>S. pumilus</i>		<i>T. junceum</i>		<i>S. pumilus</i>			
	Order ^a	df	LRT ^b	p	LRT ^b	p	LRT ^b	p	LRT ^b	p	
1075	Plastic (P)	3	1	1.22	0.541	1.33	0.512	15.25	<0.001	22.61	<0.001
1076	Nitrogen (N)	3	1	1.23	0.266	0.00	1.000	0.82	0.365	6.86	0.008
1077	Biotic cond (B)	3	1	4.14	0.126	2.27	0.320	15.59	<0.001	0.09	0.951
1078	P x N	2	1	2.79	0.246	2.25	0.323	2.85	0.239	4.74	0.093
1079	P x B	2	1	0.88	0.926	3.79	0.434	2.88	0.577	0.60	0.962
1080	N x B	2	1	2.28	0.318	2.33	0.310	0.04	0.977	1.37	0.503
1081	P x N x B	1	1	3.50	0.476	5.50	0.239	3.06	0.546	2.89	0.574

1082 ^aOrder indicates the sequence in which fixed terms were removed from the model for the LRTs. ^b
1083 Significance of the fixed terms were tested using likelihood ratio tests (LRTs) comparing the
1084 models with and without the term of interest

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1091 **Table S2.** Summary of the generalized linear models testing the effects of plastic, nitrogen (N)
1092 deposition and biotic condition and selected interactions on the probability of plant survival of
1093 *Thinopyrum junceum* and *Sporobolus pumilus*.
1094 Results of generalized linear model testing the effects of plastic (without plastic, with non-
1095 biodegradable or biodegradable/compostable bag), nitrogen (ambient, elevated), and biotic
1096 condition (no interaction, interaction with a conspecific or a hetero-specific neighbour) and selected
1097 interactions on the probability of plant survival of *Thinopyrum junceum* and *Sporobolus pumilus* as
1098 target. Results are reported for models retaining only those terms identified as significant to the
1099 model according to Likelihood Ratio Test (LRTs). Because no terms were significant in the LRTs
1100 for propagule establishment, no model for establishment is included here. Significant results ($p <$
1101 0.05) are in bold.

	<i>T. junceum</i>				<i>S. pumilus</i>			
	Estimate	Std. Error	z	P	Estimate	Std. Error	z	P
1104 Intercept	1.469	0.537	2.736	0.006	1.336	0.460	2.904	0.003
1105 Plastic (Biodegradable)	-0.928	0.570	-1.629	0.103	0.281	0.531	0.529	0.596
1106 Plastic (Non-biodegradable)	-2.293	0.601	-3.811	<0.001	-2.099	0.564	-3.716	<0.001
1107 Nitrogen (Addition)	-	-	-	-	-1.170	0.462	-2.530	0.011
1108 Biotic cond (Conspecific)	1.096	0.578	1.893	0.058	-	-	-	-
1109 Biotic cond (Hetero-specific)	-1.098	0.540	-2.034	0.041	-	-	-	-

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1124 **Table S3.** Percentage of alive plants of (a) *Thinopyrum junceum* and (b) *Sporobolus pumilus* as
 1125 target under different treatment combinations observed from May 2018 to April 2019.

1126 a) *Thinopyrum junceum*

1127	Month	M	J	J	A	S	O	N	D	J	F	M	A
1128	Treatment												
1129	WP-AN-NoI	100	100	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3
1130	WP-AN-CoI	100	100	100	100	100	100	100	100	100	100	100	100
1131	WP-AN-HeI	100	100	83.3	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1132	WP-EN-NoI	100	100	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3
1133	WP-EN-CoI	100	100	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3
1134	WP-EN-HeI	100	100	66.6	50	50	50	50	50	50	50	50	50
1135	NBP-AN-NoI	100	100	50	50	50	50	50	50	50	50	50	50
1136	NBP-AN-CoI	100	100	83.3	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1137	NBP-AN-HeI	100	100	50	16.6	16.6	16.6	16.6	16.6	16.6	16.6	16.6	16.6
1138	NBP-EN-NoI	100	100	66.6	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3
1139	NBP-EN-CoI	100	100	50	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3
1140	NBP-EN-HeI	100	100	16.6	0	0	0	0	0	0	0	0	0
1141	BP-AN-NoI	100	100	83.3	50	50	50	50	50	50	50	50	50
1142	BP-AN-CoI	100	100	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3
1143	BP-AN-HeI	100	100	100	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3
1144	BP-EN-NoI	100	83.3	50	50	50	50	50	50	50	50	50	50
1145	BP-EN-CoI	100	100	100	100	100	100	100	100	100	100	100	100
1146	BP-EN-HeI	100	100	83.3	50	50	50	50	50	50	50	50	50

1147 b) *Sporobolus pumilus*

1148	Month	M	J	J	A	S	O	N	D	J	F	M	A
1149	Treatment												
1150	WP-AN-NoI	100	83.3	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1151	WP-AN-CoI	100	100	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1152	WP-AN-HeI	100	83.3	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1153	WP-EN-NoI	100	100	83.3	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1154	WP-EN-CoI	100	83.3	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1155	WP-EN-HeI	100	100	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1156	NBP-AN-NoI	100	100	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3
1157	NBP-AN-CoI	100	100	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3	33.3
1158	NBP-AN-HeI	100	100	50	50	50	50	50	50	50	50	50	50
1159	NBP-EN-NoI	100	100	0	0	0	0	0	0	0	0	0	0
1160	NBP-EN-CoI	100	100	16.6	16.6	16.6	16.6	16.6	16.6	16.6	16.6	16.6	16.6
1161	NBP-EN-HeI	100	100	0	0	0	0	0	0	0	0	0	0
1162	BP-AN-NoI	100	100	100	100	100	100	100	100	100	100	100	100
1163	BP-AN-CoI	100	100	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3
1164	BP-AN-HeI	100	100	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3	83.3
1165	BP-EN-NoI	100	100	50	50	50	50	50	50	50	50	50	50
1166	BP-EN-CoI	100	100	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6	66.6
1167	BP-EN-HeI	100	83.3	66.6	50	50	50	50	50	50	50	50	50

1168 WP: without plastic; BP: biodegradable plastic; NBP: non-biodegradable plastic; NoI: no neighbour or no interaction;
 1169 CoI: plant grown with a conspecific; HeI: plant grown with a hetero-specific; AN: ambient level of nitrogen; EN:
 1170 elevated level of nitrogen

1171 **Table S4.** Summary of the generalized linear models testing the significance of the effects of
 1172 plastic, nitrogen (N) deposition and biotic condition and their interactions on the probability of plant
 1173 survival of *Sporobolus pumilus* as target.
 1174 Results of generalized linear model testing the effects of plastic (without plastic, with
 1175 biodegradable/compostable bag), nitrogen (ambient, elevated), and biotic condition (no
 1176 interaction, interaction with a conspecific or a hetero-specific neighbour) and selected interactions
 1177 on the probability of plant survival of *Sporobolus pumilus* as target. Results are reported for
 1178 models retaining only those terms identified as significant to the model according to Likelihood
 1179 Ratio Test (LRTs). Significant results ($p < 0.05$) are in bold.

1180 *S. pumilus*

1181		Order ^a	df	LRT ^b	p
1182	Plastic (P)	3	1	0.26	0.608
1183	Nitrogen (N)	3	1	2.38	0.122
1184	Biotic cond (B)	3	1	0.13	0.934
1185	P x N	2	1	2.86	0.090
1186	P x B	2	1	0.17	0.914
1187	N x B	2	1	0.51	0.774
1188	P x N x B	1	1	1.37	0.503

1189 ^aOrder indicates the sequence in which fixed terms were removed from the model for the LRTs. ^b
 1190 Significance of the fixed terms were tested using likelihood ratio tests (LRTs) comparing the
 1191 models with and without the term of interest
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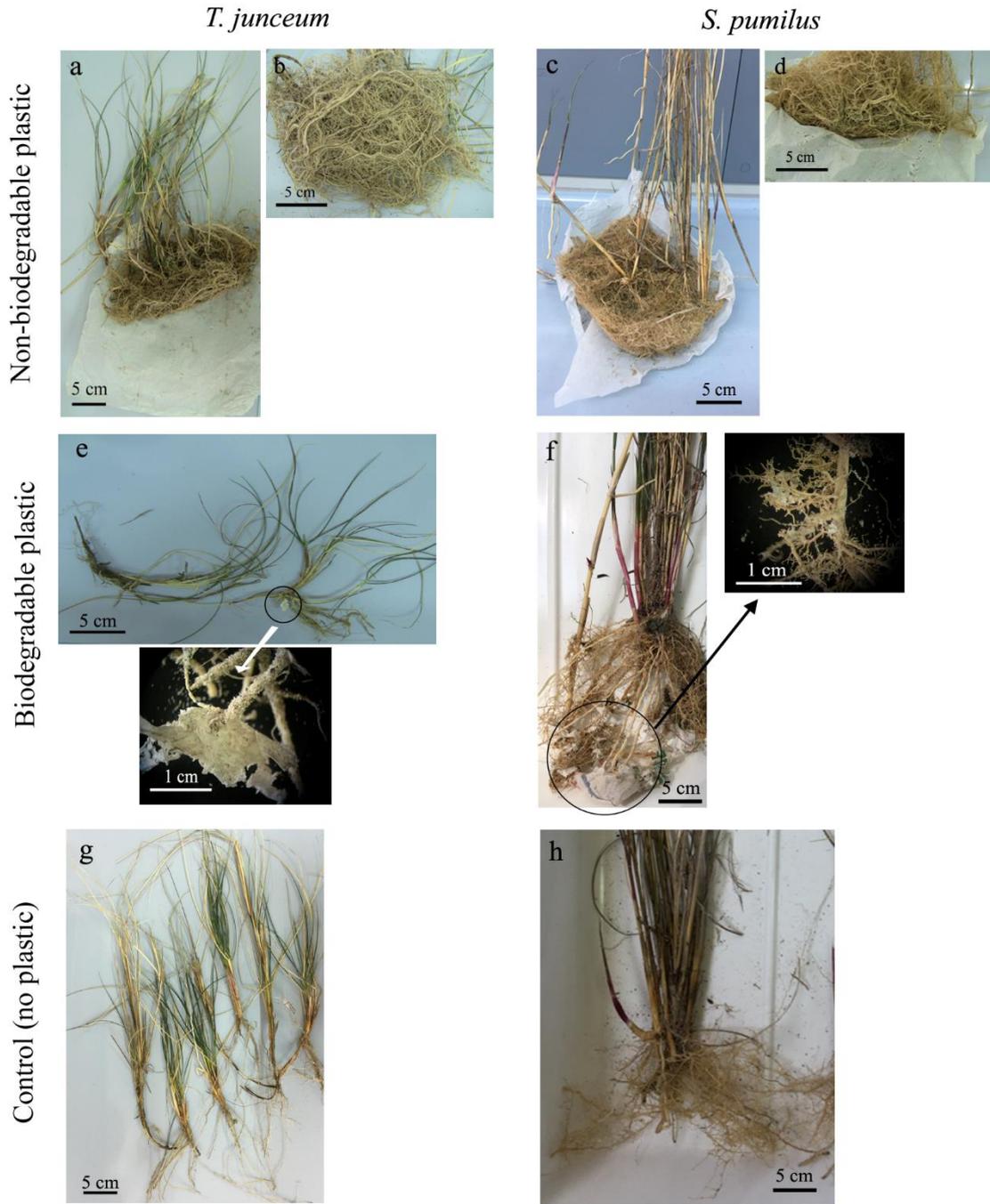


Fig. S1. One-year old plants of *Thinopyrum junceum* (left panel) and *Sporobolus pumilus* (right panel) established from vegetative propagules in the presence of a non-biodegradable plastic bag macro-fragment (a, b, c, d) or with biodegradable/compostable plastic bag macro-fragment (e, f) or in bare sand without plastic (g, h).

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Declarations of interest: none

CRedit authorship contribution statement

Virginia Menicagli: Conceptualization, Formal analysis, Investigation, Visualisation, Writing- Original draft. **Elena Balestri:** Conceptualization, Formal analysis, Investigation, Visualisation, Writing- Original draft. **Flavia Vallerini:** Formal analysis, Investigation, Visualisation. **Alberto Castelli:** Supervision, Writing- Review & Editing. **Claudio Lardicci:** Conceptualization, Investigation, Supervision, Writing- Review & Editing.