1	HUMAN ADENOVIRUS IN MUNICIPAL SOLID WASTE LEACHATE AND
2	QUANTITATIVE RISK ASSESSMENT OF GASTROINTESTINAL ILLNESS TO
3	WASTE COLLECTORS
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Abstract

35 Leachate is a variable effluent from waste management systems generated during waste collection and on landfills. This study aimed to evaluate the gastrointestinal (GI) illness 36 risk of waste collectors exposed to human adenovirus (HAdV) contaminated leachate. In 37 this study, Ttwenty-two leachate samples (1L each) from a waste collection trucks and a 38 landfill were collected from March to December 2019 in the municipality of Rio de 39 Janeiro (Brazil) in a solid waste transfer station and a landfilland they were analyzed for 40 Human Adenovirus (HAdV), bacterial indicators, and various physico-chemical 41 parameters, For viral analysis, Samples samples were concentrated by ultracentrifugation 42 and processed for molecular analysis using QIAamp Fast DNA Stool mini kit® for DNA 43 extraction followed by nested-PCR and qPCR/PMA-qPCR TaqMan® system. HAdV was 44 detected by nested-PCR in 100% (9/9) and 83.33% (12/13) of the truck and landfill 45 leachate samples, respectively. Viral concentrations ranged from 8.31×10^1 to 6.68×10^7 46 genomic copies per 100 ml by qPCR and PMA-qPCR. HAdV species A, B, C, and F were 47 characterized using nucleotide sequencing. HAdV were isolated in A549 culture cells in 48

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49	100% (9/9) and 46.2% (6/13) from truck and landfill leachate samples, respectively.					
50	Regardless of the detection methods, HAdV concentration was predicted by the quantity					
51	of total suspended solids. Analysis of <u>A</u> quantitative microbial risk assessment was					
52	performed to measured the probability of gastrointestinal (GI) illness attributable to					
53	inadvertent oral ingestion of truck leachate (hand to mouth contact or direct splashing					
54	into the oral cavity), revealing the higher probability of disease for the direct splashing					
55	route exposure into the oral cavity (58%) than for the gloved hand-to-mouth (33%). In a					
56	scenario where waste collectors do not wear gloves as protective personal equipment, the					
57	risk increases to 67%. This is the first study on revealing infectious HAdV in solid waste					
58	leachate that and it indicates a potential health risk of exposure of the for waste collectors					
59	to GI illness.					
60						
61	Keywords: leachate; municipal solid waste; human adenovirus; PMA-qPCR; cell					
62	culture; QMRA.					
63						
64	1. Introduction					
65	Municipal solid waste (MSW) leachate is an effluent produced at different stages of waste					
66	management systems, such as landfills, collection trucks, and transfer stations. This					
67	effluent has variable physical chemical and biological characteristics that can be affected					
68	by MSW composition, landfill lifetime and structures, transfer stations operation mode					

as well as local weather conditions (Youcai 2019). MSW leachate has high values of

organic matter, ammonia nitrogen, chloride, and dissolved solids (Costa et al. 2019). The

physicalphysico-chemical composition of MSW leachate landfill has been widely studied

(Costa, Alfaia, and Campos 2019; Naveen et al. 2017), however, but truck leachate are

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still unknown (Benyoucef et al. 2015). Concerning microbiological contamination, some 3

74	studies have described the occurrence of bacteria in the leachate (Silva et al. 2011; Zhang					
75	et al. 2013), however, viruses were rarely considered few investigations report the virus					
76	detection, most of them related to methodologies of virus recovery from this matrix					
77	(Sobsey, Wallis, and Melnick 1974, 1975; Sobsey 1978; Costa, Alfaia, and Campos					
78	2019). The first work that quantifiesd infectious HAdV in a landfill was described by					
79	Carducci et al. (2013), who investigated the presence of HAdV and Torque Teno Virus					
80	(TTV) in the air and on the surfaces of a landfill in Italy (Carducci, Federigi, & Verani,					
81	2013). HAdV have been identified as potential viral indicator of human fecal					
82	contamination due to (1) stability, persistence and wide distribution in different					
83	environmental matrices; (2) detection in sewage without seasonality; (3) resistance to					
84	ultraviolet (UV) disinfection; (4) greater abundance in relation to other enteric viruses;					
85	(5) isolation in culture and molecular tools available; (6) viral specificity with the human					
86	host (Rames, Roiko, Stratton, & Macdonald, 2016). MSW handling implies exposure of					
87	waste collectors to chemical substances, dust, and microorganisms (Gorman Ng et al.					
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99	oral ingestion of microbes during waste collection is rarely addressed, except for
100	monitoring studies on fungal and bacteria contamination on workers' hands (Madsen et
101	al. 2016; Madsen et al. 2020). To fill this gap on occupational exposure risk, we
102	studied workersexposure scenarios from through inadvertent ingestion of collection
103	trucks leachate was investigated to estimate and estimated HAdV GI illness risk. In this
104	study, trucks and landfill leachate samples were characterized by analyzed for bacterial
105	and physicoal-chemical parameters and bacterial findings presence and investigated for
106	HAdV occurrence, concentration, genetic diversity, and infectivity.

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108 2. Materials and Methods

109 2.1 Study area and samples

Samples were collected in two stages from the waste management system in the 110 municipality of Rio de Janeiro, Brazil: fresh leachate from waste collection trucks at a 111 transfer station and raw leachate at an operational landfill. The transfer station serves 11 112 neighborhoods, comprehending 900 thousand inhabitants. The regional landfill serves 6.9 113 million inhabitants, receiving around ten thousand tons per day of MSW. 114 One-One-liter samples were collected from March to December 2019. Nine samples of 115 fresh leachate were collected from collection truck basins, each sample corresponding to 116 a pool of three to five trucks. Thirteen raw leachate samples were collected from a tank 117 at the entrance of the landfill leachate treatment plant. All samples were collected in 118 sterile polyethylene bottles, transported to the laboratory at 4 °C, and processed within 24 119 h. 120

- 121
- 122 2.2 Virological Methods

123 2.2.1 Viral concentration

- 124 Viral concentration was performed by ultracentrifugation at 100.000 x g for 1 h at 4° C
- using a Sorvall[®] WX Ultra Centrifuge Series (Thermo Scientific, Waltham, MA, USA).
- 126 Pellet was resuspended on ice for 30 min with glycine buffer and then centrifuged 12.000
- 127 <u>x g for 15 min. Supernatant was centrifuged at 100.000 x g for 1 h at 4°C and pellet</u>
- 128 resuspended at 500 µL PBS according to a-previous studiesy (Pina et al. 1998, (Lanzarini
- et al., 2020)) and stored at -80 °C until nucleic acid extraction.
- 130 2.2.2 DNA extraction
- 131 Nucleic acids were extracted using the QIAamp Fast DNA Stool Mini Kit[®] (Qiagen,
- 132 Valencia, CA, USA) according to the manufacturer's instructions and a previous study
- 133 (Lanzarini et al., 2020). Positive A clinical fecal sample was used as a HAdV positive
- 134 <u>control.</u> and negative controls were processed. The UltraPureTM DNase/RNase-Free
- 135 (Invitrogen, Carlsbad, CA, USA) distilled water was used as the negative control.
- 136 2.2.3 Qualitative PCR
- 137 A semi-nested PCR was performed, using primers previously described to obtain 245
- base pairs (bp) fragments (Allard, Albinsson, and Wadell 2001), and the Platinum[®] Taq
- 139 DNA Polymerase enzyme was used according to the manufacturer's recommendations
- 140 (Invitrogen, Carlsbad, California, USA).
- 141 2.2.4 Real-Time PCR (qPCR) and PMA-qPCR

For PMA-qPCR, 200µl of each concentrated was inoculated with PMAxxx[™] (Biotium
Inc., Freemont, CA, USA) at a final concentration of 50 µM and incubated in a dark room
for 10 minutes under 200 rpm at 25 °C. The samples were incubated for 15 minutes in
the PMA-Lite[™] LED Photolysis Device (Biotium Inc., Hayward, CA, USA) for
photolysis (Fongaro et al. 2016). DNA extractions and qPCR reactions were processed
simultaneously for leachate samples treated or not with PMA.

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HAdV hexon gene was amplified using a set of primers and probe for TaqMan[®] system 148 149 protocol previously described (Hernroth et al. 2002). Each 15 µl qPCR contained 12.5 µl of 2 x TaqMan® Universal PCR Master Mix (Applied Biosystems, Foster City, CA, 150 USA), 1 μ l of each 22.5 μ M primer (0.9 μ M final concentration each), and 0.5 μ l of probe 151 (0.225 µM final concentration). Reactions were performed on ABI PRISM 7500 (Applied 152 Biosystems, Foster City, CA, USA) according to the following cycle: 2 min at 50 °C, 10 153 min at 95 °C, and 40 cycles of 15 s at 95 °C, 1 min at 60 °C. Synthetic DNA standard 154 curves were designed using specific viral gBlock Gene Fragments (Integrated DNA 155 TechnologiesTM, Coralville, Iowa, USA) with a ten-fold serial dilution $(10^7 - 10^0$ genome 156 copies (GC)/reaction) containing the same HAdV target sequence. 157

Truck and landfill leachate were tested in duplicate using undiluted and diluted samples (10^{-1} and 10^{-2}), totalizing six reactions per sample. All qPCR included non-template control (NTC), positive control, and DNAse/RNAse-free water as a negative control. The viral load obtained, correct<u>eding</u> for the dilution analyzed, was expressed as the number of GC per 100 ml. The limit of detection for this protocol was 2.0×10^3 GC 100 ml⁻¹ and samples with Ct \leq 38 were considered positive.

164 2.2.5 Nucleotide sequencing (Sanger method)

PCR product was purified from electrophoresis gel using Wizard® SV Gel and PCR 165 166 Clean-Up System (Promega, Madison, MI, USA). Sequences were obtained using the Big Dye Terminator v3.1 Cycle Sequencing Kit and the ABI Prism 3730 Genetic Analyzer 167 48 capillary sequencer (Applied Biosystems, Foster City, CA, USA) from the 168 "PDTIS/FIOCRUZ DNA Sequencing Platform", and analyzed using MEGA X[®] software 169 (Kumar et al. 2018) for editing, sequence analysis, and molecular characterization. The 170 sequences were compared with reference prototypes deposited at the NCBI GenBank 171 (https://www.ncbi.nlm.nih.gov/). Phylogenetic dendrograms were obtained using the 172

Maximum Likelihood method, employing genetic distance corrected by the Tamura 3
parameter model and Gamma Distributed (G) with 2000 bootstraps above 70% (Tamura

175 1992).

176 2.2.6 HAdV infectivity assay in A549 cells

The human lung epithelial carcinoma A549 cell line (ATCC CCL-185) was used for virus 177 isolation and titration onto 96-well polystyrene microtiter plates. Each leachate sample 178 was tenfold diluted $(10^{-1} \text{ to } 10^{-4})$ in Eagle's minimum essential medium (MEM) 179 supplemented with a 10% L-glutamine and 0.125% gentamycin. Each dilution was 180 seeded into five wells, containing 75 µl of the sample, 75 µl of MEM supplemented with 181 a 0.125% gentamycin, 0.1% HEPES buffer, and 50 µl of A549 cell (106 cells ml-1). Plates 182 were covered and incubated at 37 °C under 5% CO2 for 5 days. Examination for 183 cytopathic effects was performed with inverted light microscopy. The highest dilution 184 producing a cytopathic effect in 50% of the inoculated cells was determined using the 185 Spearman-Karber formula (Ramakrishnan 2016) and expressed as TCID₅₀ ml⁻¹ (50% 186 tissue culture infective dose per milliliter). The limit of detection was 10^{1.12} TCID₅₀ ml⁻¹. 187 188

189 2.3 Bacteriological and physico-chemical parametersphysical-chemical findings

Total coliforms and *Escherichia coli* (*E. coli*) from the fresh truck and landfill leachate
samples were quantified using Colilert[®] Quanti-Tray[®]/2000 (IDEXX Laboratories, Inc,
Westbrook, ME, USA), and results were expressed by Most Probable Number (MPN) per
100 ml. Samples were tested using ten-fold dilutions previously standardized, truck
leachate ranging from 10⁻⁹ to 10⁻¹³ and landfill from 10⁻¹ to 10⁻³.
The following physico-chemical parameters physical chemical-were analyzed according

to Standard Methods for the Examination of Water and Wastewater (APHA 2017): pH,

197 color, turbidity, total alkalinity, conductivity, total hardness, solids, carbon, chemical

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oxygen demand (COD), UV 254 nm, chloride, <u>phosphorusphosphor</u>, ammonia, and total
nitrogen.

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201 2.4 QMRA

QMRA inputs are the environmental concentration of pathogens and the amount of exposure to a matrix (dose) to estimate the associated probability (risk) of an adverse outcome (illness), using pathogen-specific functions describing the dose-response relationship. QMRA involves the description of the exposure scenario, the exposure assessment, the dose-response model, and the risk characterization (Haas, Rose, and Gerba 2014).

208 2.4.3 Selection of the exposure scenario

During a working day, waste collectors are were engaged in manually door-to-door waste 209 collection, filling of trucks, and driving of compactor systems, where they are were 210 exposed to leachate. We assume landfill Landfill workers as were assumed as not exposed 211 to leachate since this such effluent flows flowed in to a dedicated treatment facility without 212 human contact (Lanzarini et al. 2020). Accidental oral ingestion of truck leachate was 213 modelled for workers that did not adopt <u>We modeled the scenario of occupational</u> 214 exposure to truck leachate, which can be orally ingested purely by accident. We 215 considered the exposure scenario where the workers do not wear protective masks as 216 before the COVID-19 pandemic, but they do wear wearing a specific type of gloves for 217 prevention of accidents with sharp objects, manufactured by a mixture of synthetic fibers 218 219 and natural fibers (usually latex). Considering In a hypothetical scenario of difficulty in adhering to the use of gloves by waste collectors, we also the health risk without such 220

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221	<mark>protective personal equipment (PPE) was also calculated</mark> , included a scenario of risk	
222	analysis when gloves are not worn.	
223	2.4.4 Exposure assessment	
224	To estimate the dose, The dose was estimated using we consider the conceptual	ha formattato: Evidenziato
225	compartmental model developed by the Institute for Occupational Medicine (Cherrie et	
226	al. 2006; HSE 2007) for inadvertent ingestion of liquid chemical substances and dust in	
227	occupational settings, adapted to our specific exposure scenario. Oral exposure of waste	
228	collectors relied on two possible mechanisms for transferring HAdV to the oral cavity:	
229	contact of contaminated hands with the skin around the mouth (hand-to-mouth contact)	
230	and direct contamination of truck leachate into the mouth or the peri-oral area (splashing	
231	route). In both cases, we derived HAdV concentration (C _{HAdV}) was derived from the truck	ha formattato: Evidenziato
232	leachate monitoring data, considering the infective viral load results expressed in TCID ₅₀	
233	ml ⁻¹ . Since the number of samples (9) was insufficient to establish a theoretically fitted	
234	distribution, we created an empirical uniform distribution was drawn considering based	ha formattato: Evidenziato
235	on-the lower and upper bound of the experimental measurements, following an approach	ha formattato: Evidenziato
236	well established in QMRA literature -(Rasheduzzaman et al. 2019).	
237	Oral ingestion by hand-to-mouth contact, Waste collectors can contaminate gloved	Commentato [IF2]: I remove sentece in first person singular as suggested in one Reviewer comment
238	hands byWaste collectors can have their gloved hands contaminate by touching truck	
239	leachate accumulated on the surface of the vehicles, which have been were assumed to be	
240	heavily polluted with a leachate load of 1 drop per \mbox{cm}^2 (LL $_{\mbox{surface}})$ (Shatkin, Smith, and	
241	Moyer 2005). To estimate the exposure dose, we considered The input parameters used	ha formattato: Evidenziato
242	in the estimation of exposure dose were \pm the contact area between the hand and the oral	ha formattato: Evidenziato
243	compartment (SA _{hand-mouth})-was considered to estimate the exposure dose, the viral	
244	transfer efficiencies from surfaces-to-hands (TEsurface-hand), and from hand-to-mouth	
I		

245 (TE_{hand-mouth}), and hourly frequency of hand-to-mouth contact ($f_{contact}$), as reported in 246 Equation 1. To understand the role of gloves on the ingested dose, we used-different 247 values for TE_{surface-hand}, TE_{hand-mouth}, and $f_{contact}$ were used on the basis of the adoption of 248 such PPEbased on the use of protective personal equipment (PPE).

249 Workers without gloves. We chose a TE_{surface-hand} was derived from a study on viral transfer from hard nonporous surfaces to bare hand in a condition of high air relative 250 humidity (a Brazilian climate feature) since transfer efficiency of organisms is improved 251 252 under a high relative humidity of 40% to 65%, especially for viruses (Lopez et al. 2013). For TE_{hand-mouth}, we used data from a study investigating viral transfer from contaminated 253 bare hand-hand-to-to-mouth was used (34%, Rusin, Maxwell, and Gerba 2002), which 254 255 reported a value of 34%. For f_{contact}, we adopted the The fcontact frequency was 256 derived from of-workers employed in manufacturing and engineering sectors, who can perform manual work, but without wearing gloves (Zainudin 2004; Cherrie et al. 2006) 257 258 <u>Workers wearing gloves</u>. For The TE_{surface-hand}, we was selected from the viral transfer percentage to latex-gloved hands from a nonporous hard surface, contaminated with a 259 "wet" inoculum, to obtain a conservative estimation of health risk (Sharps, Kotwal, and 260 Cannon 2012). In the absence of information on TEhand-mouth for gloved hands, we adjusted 261 the data on bare hands to mouth (Rusin, Maxwell, and Gerba 2002) were adjusted with 262 results from studies on food production, which demonstrated that latex gloves are were 263 10% less efficient to transfer viruses than the bare hands, by a 10% (Tuladhar et al. 2013; 264 Rönnqvist et al. 2014). The f_{contact} was very low among workers wearing gloves, based on 265 laboratory and pesticide workers' behavior (Zainudin 2004; Cherrie et al. 2006). 266

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Apart from PPE adoption, exposure from each hand-to-mouth contact was considered independent of contact duration, as in other QMRA studies modeling non-dietary exposure to pathogens (Mattioli, Davis, and Boehm 2015).

 $270 \qquad dose_{hand-to-mouth} = C_{HAdV} * LL_{surface} * TE_{surface/hand} * SA_{hand-mouth} * TE_{hand/mouth} * f_{contact} * n_{hours} (eq. 1)$

where dose_{hand-to-mouth} is the HAdV dose transferred to the oral cavity as a result of handto-mouth mechanisms during a working day [Number of HAdV in TCID₅₀], and each input variable is explained in Table S1.

Oral ingestion by splashing route. Workers can directly ingest truck leachate during the 274 operation of waste compaction by each vehicle. The number of compacting events during 275 a working day can vary since waste collection in densely populated city areas could imply 276 dozens of compacting activities. Although the button to activate compactors is on the side 277 of the truck to prevent accidents, workers are still exposed to dust and drops of truck 278 leachate, which can splash onto the oral and peri-oral compartments of waste collectors 279 who stand near the compactor. During a working day, We assume that during a working 280 281 day, the a worker can be was assumed to be exposed to 1 to 50 drops based on the 282 professional judgment of researchers of the field and after a discussion with the Municipal 283 Waste Company interview. We modeled therefore, the exposure volume of truck leachate accidentally ingested by splashing varied from 0.01 ml to 0.5 ml (V_{leachate}), since 284 10 µL droplets are considered because large aerosol liquid particles of 10 µL spread at a 285 286 short distance from a source (Sattar, Ijaz, and Gerba 1987). The HAdV dose accidentally ingested by direct splashing in the oral cavity was computed according to Equation 2. 287

 $dose_{splashing} = C_{HAdV} * V_{leachate}$ (eq. 2)

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289	where dose _{splashing}	g is the HAdV	dose transferred	in the oral	cavity as a re	sult of splashing
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- 290 mechanisms during a working day [Number of HAdV in TCID₅₀] and each input variable
- is explained in Table S1. 291

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292 2.4.5 Dose-response model

Dose-response models are mathematical functions obtained from clinical trials or

outbreak data to describe the relationship between a specific pathogens and hosts on the

- basis of the transmission routes (contact, ingestion, inhalation) (Federigi et al. 2019). 295
- 296 In this study, the illness was chosen as the endpoint of the dose-response model, which
- was We used a dose-response model that provided GI illness as the endpoint and chose 297
- the hypergeometric dose response equation developed by (Teunis, Schijven, and Rutjes 298
- 2016), with optimized model parameters values for oral ingestion, obtained from clinical 299
- trials on inoculation of adenovirus (AdV4, AdV7). The probability of infection (Pinf) was 300
- computed by the formula: 301

 P_{inf} (dose; α , β) = 1- $_1F_1(\alpha, \alpha + \beta, -dose)$

- 303 where α and β represent infection parameters specific for HAdV ($\alpha = 5.11$, $\beta = 2.80$) and ha formattato: Evidenziato $_{1}F_{1}$ represents is the confluent hypergeometric function of the first kind. 304 ha formattato: Evidenziato 305 The probability of illness (Pill) for each pathogen was estimated by multiplying the Pinf and the pathogenicity, which represents the probabilities of developing illness given the 306 infection, calculated as: 307 $P_{\text{ill}|\text{inf}}(\text{dose} \mid \eta, r) = 1 - (1 + \text{dose}/\eta)^{-r}$ 308 where r and η are the illness parameters for HAdV ($\eta = 6.53$, r = 0.41). 309 ha formattato: Evidenziato
- 2.4.6 Risk characterization 310

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A Monte Carlo analysis was run for 200 simulations, each one of 10.000 interactions

312 (Vensim package, Ventana Systems, Inc., Harvard, MA, USA). The result of the Monte

313 Carlo can be observed as<u>was</u> the daily probability of illness and based on 200 measures

to improve the accuracy of the health risk.

A sensitivity analysis has been was carried out to test the relative importance of the stochastic variables on models' results: C_{HAdV} , $f_{contact}$, $T_{Esurface/hand}$ for hand-to-mouth, and C_{HAdV} and $V_{leachate}$ for splashing mechanism. To determine the effect of these variables on the final risk estimate, their value was varied, one at a time, while keeping each of the other input parameters constant or fixed at their average value of its probability distribution function.

321

322 2.5 Statistical analysis

Microbiological data were log-transformed and analyzed using *GraphPad Prism* version 8.0.1. Samples below the limit of detection received half of this value for statistical analysis (Bucardo et al. 2011).

Fisher exact test was used to compare HAdV detection methods, and an unpaired t-test to compare HAdV, total coliforms, and *E. coli* concentrations between truck and landfill leachate. Spearman correlation and regression analysis were performed to investigate the correlation between HAdV infectivity, qPCR, and PMA-qPCR. Mann-Whitney test was used to compare the physico-chemical parameters physical chemical findings between truck and landfill leachate. The results were considered significant when P values were below 0.05.

A multiple regression model was performed using each microbial variable (total coliforms, *E. coli*, and HAdV) once at a time, as the dependent variable and the <u>physico-</u> ha formattato: Evidenziato

335	chemical parameters physical chemical findings as predictors (independent variables).	
336	The physico-chemical parameters physical-chemical findings were examined for	
337	collinearity based on Pearson's r correlation and those with low collinearity with each	
338	other (- $0.5 < r < 0.5$) were selected as predictors (Zuur, Ieno, and Elphick 2010). The	
339	best model was selected with Akaike Information Criterion (AIC) (Fox and Weisberg	
340	2011) and the analysis was carried out in R-Language using the car and effects packages	
341	(Team 2018).	
342	The exposure dose and the probability of GI illness were described in terms of the	
343	interquartile range (IQR).	
344		
345	3 Results	
345 346	3 Results 3.1 Viral and bacteriological findings parameters	
346	3.1 Viral and bacteriological findings parameters	
346 347	3.1 Viral and bacteriological findings parametersHAdV were detected in 100% (9) of truck leachate samples regardless of the detection	
346 347 348	3.1 Viral and bacteriological findings parameters HAdV were detected in 100% (9) of truck leachate samples regardless of the detection method used (semi-nested PCR, qPCR, PMA-qPCR, cell culture) and in 92% (12) of	
346 347 348 349	3.1 Viral and bacteriological findings parameters HAdV were detected in 100% (9) of truck leachate samples regardless of the detection method used (semi-nested PCR, qPCR, PMA-qPCR, cell culture) and in 92% (12) of landfill samples, with variable detection percentage according to the methodology (Table	
346 347 348 349 350	<i>3.1 Viral and bacteriological findings parameters</i> HAdV were detected in 100% (9) of truck leachate samples regardless of the detection method used (semi-nested PCR, qPCR, PMA-qPCR, cell culture) and in 92% (12) of landfill samples, with variable detection percentage according to the methodology (Table 1). No statistically significant difference was observed between cell culture detection with	
346 347 348 349 350 351	<i>3.1 Viral and bacteriological findings parameters</i> HAdV were detected in 100% (9) of truck leachate samples regardless of the detection method used (semi-nested PCR, qPCR, PMA-qPCR, cell culture) and in 92% (12) of landfill samples, with variable detection percentage according to the methodology (Table 1). No statistically significant difference was observed between cell culture detection with qPCR and PMA-qPCR in landfill leachate. Semi-nested PCR showed a significant	

Table 1. Human adenovirus (HAdV), total coliforms, and Escherichia coli (E. coli)

findings from truck and landfill leachate samples according to the methodology used.

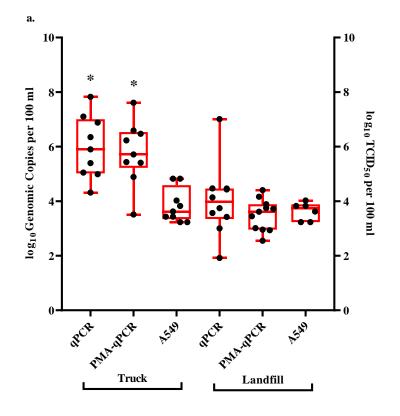
Mianoanaiana	Method	Leachate origin N. of positive (%)		Minimum- Maximum	
Microorganisms	Method	Truck (n=9)	Landfill (n=13)	Truck	Landfill
	Semi-nested-PCR	9 (100)	12 (92)	-	-
HAdV	qPCR*	9 (100)	11 (85)	$2.07 {\times} 10^4 {-}~6.68 {\times} 10^7$	$8.31 \times 10^{1} - 1.02 \times 10^{7}$

Commentato [IF4]: To be in line with reviewer #2 suggestion on chemical paramters

	PMA-qPCR*	9 (100)	10 (77)	$3.25{\times}10^3{-}4.06{\times}10^7$	$4.00{\times}10^2{-}2.53{\times}10^4$
	Infectivity assay** (A549 cell)	9 (100)	6 (46)	$1.70 \times 10^3 - 6.67 \times 10^4$	$1.70 \times 10^3 - 1.06 \times 10^4$
Total coliforms	Colilert [®] Quanti-Tray [®] ***	9 (100)	12 (92)	$1.00{\times}10^{10}{-}6.05{\times}10^{14}$	$2.00 \times 10^3 - 9.00 \times 10^5$
E. coli	Colilert® Quanti-Tray®***	7 (78)	8 (62)	$1.18{\times}10^{11}{-}1.30{\times}10^{13}$	$1.00 \times 10^{1} - 5.00 \times 10^{4}$

*GC 100 ml⁻¹ **TCID₅₀ 100 ml⁻¹ ***MPN 100 ml⁻¹ 357 358

361	Regarding the viral and bacteriological concentration, a higher result in samples of fresh	
362	leachate (truck) was observed when compared to landfill samples with statistically	
363	significant differences observed for HAdV qPCR (p=0.0031), PMA-qPCR (p<0.0001),	
364	total coliforms (p<0.0001), and E. coli (p<0.0001). No significant differences were found	
365	between A549 cell leachate origins based on HAdV infectivity (p=0.4854) (Figure 1).	
366		



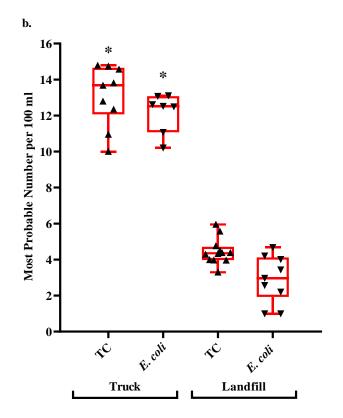
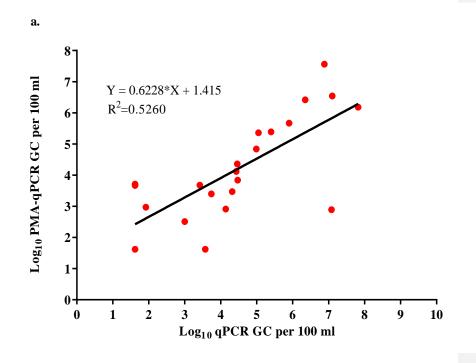


Figure 1. Quantification of human adenovirus (a), total coliforms (TC) and *Escherichia coli* (*E. coli*) (b) from truck and landfill leachate samples.

373

Considering HAdV quantification, a correlation between results obtained from qPCR and PMA-qPCR in truck leachate (p < 0.05) was observed and between infectivity titer and PMA-qPCR in landfill (p < 0.05). When the results obtained from the truck and the landfill samples were analyzed together, the correlation between qPCR and PMA-qPCR methods was maintained (p < 0.001) with a good linear correlation ($R^2 = 0.526$). (Figure 2a). The concentration values obtained by these molecular methods were also correlated with HAdV infectivity and PMA-qPCR ($R^2 = 0.543$) (Figure 2b).



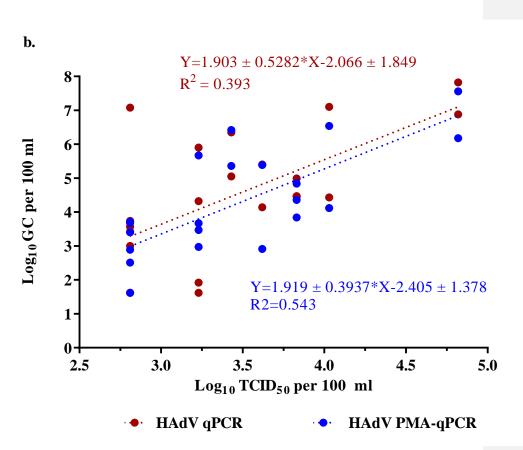




Figure 2. Regression analysis between HAdV detection methods. a) Correlation between
qPCR and PMA-qPCR. b.) Correlation between TCID₅₀ and qPCR/PMA-qPCR
quantification.

398

399 3.2 Molecular characterization

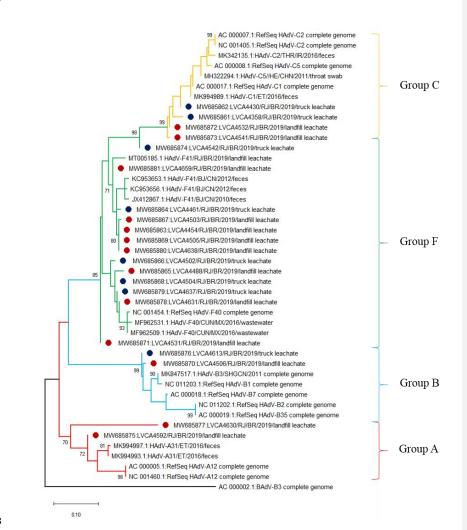
400 Nucleotide sequencing was successfully obtained from all 21 <u>samples in which HA</u>	from all 21 <u>samples in which</u>	HAdV
--	-------------------------------------	------

- 401 <u>was</u> detected and elassified the strains were classified as group F (61.9%), C (19.5%),
- and, A and B (9.5% each). From sequenced HAdV-F, 46.15% (6/13) were of type 40 and
- 40353.84% (7/13) of type 41, associated with GI symptoms. All species were detected in both20

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leachates, except HAdV group A, only detected in landfill samples (Figure 3). The
nucleotide sequence data reported here received Gen Bank accession numbers
(MW685861 to MW685881).

407



408

409 Figure 3. Dendrogram based on HAdV hexon partial nucleotide sequence (245 pb) of 21

410 Brazilian environmental samples from landfill (red circle) and truck leachate (blue circle).

411 Reference strains were obtained from GenBank. Maximum Likelihood phylogenetic tree 21 412 was constructed using Mega X Software, Tamura 3 parameter model, and Gamma

413 Distributed (G) with 2000 bootstraps above 70%.

- 414 *Physico-chemical parameters Physical-chemical findings*
- 415 <u>3.3</u>

416 The apparent color, turbidity, solids, total carbon (with a large percentage of organic 417 carbon), chemical oxygen demand (COD), and the phosphorusphosphor presented were higher values in the truck leachate when compared tothan in landfill leachate one with 418 statistical<u>ly</u> significance <u>differences</u> (p<0.0001). The values of pH, total alkalinity, total 419 inorganic carbon, ammonia, and total nitrogen showed significantly higher values in the 420 landfill when compared to the truck leachate (p<0.0001). Chloride and UV 254 nm did 421 not show a significant difference in both leachates (p>0,05). PhysicalPhysico-chemical 422 characterization from truck and landfill leachate samples is represented in Table S2. 423 The role of such variables on microbial concentrations was evaluated considering all 424 425 leachate samples together in multiple regression models, to increase the sample size for statistical analysis. Based on the analysis of the correlation coefficient, the following six 426 parameters were considered as predictors in the regression model: as conductivity, 427 chloride, total hardness, UV 254 nm, total suspended solids, and turbidity-were 428 considered. Conductivity, total suspended solids, and turbidity parameters were 429 430 correlated able to explain to more than 50% of the bacterial concentrations' variation, being 86% of total coliforms ($R^2adj = 0.8706$, p < 0.0001) and 63% of E. coli ($R^2adj =$ 431 0.6279, p < 0.001) variations - (\mathbb{R}^2 adj = 0.6279, p < 0.001). Total suspended solids and 432 conductivity were able to explain 36% and 58% of the variability of correlated to HAdV 433 $qPCR - (R^2adj = 0.3631, p < 0.01)$ and PMA- $qPCR \frac{variability}{variability} (R^2adj = 0.5806, p < 0.01)$ 434 0.0001), respectively, while infectious HAdV variability was explained only by total 435

Commentato [IF5]: I restored some words to explain the regression model, since we saved lenght in text after the revision process

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436 suspended solids $(R^2adj = 0.1819, p < 0.05)$ was correlated only with total suspended 437 solids. Regardless of the analytical method for HAdV detection, total suspended solids 438 represented the main predictor for viral concentration, with a positive correlation.

439

440 **3.3 3.4** *QMRA* **results**

than the C_{HAdV}.

459

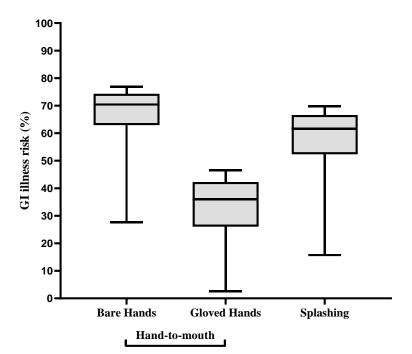
QMRA results showed that the mechanisms of oral ingestion determined different 441 exposure doses, with different GI illness risks during a working day (Figure 4). The 442 443 gloved hand-to-mouth mechanism was responsible for an average dose of 17.85 TCID₅₀ (IQR = 9.37 - 26.34) which corresponded to an average GI illness risk of 33% (IQR = 444 26% - 42%). The splashing mechanism determined a higher average exposure dose of 445 85.24 TCID₅₀ (IQR = 44.73 - 125.76) associated with a probability of illness of 58% 446 (IQR = 52% - 67%). Considering the total daily total exposure dose through oral ingestion 447 as the sum of the average dose of both mechanisms (103.10 TCID₅₀), the splashing route 448 449 is was responsible for the greater contribution to the total daily dose (~83% of the total daily dose) compared to gloved hand-to-mouth (~17%). In the case of no gloves, exposure 450 dose greatly increased to 209.32 TCID₅₀ (average value), responsible for an average 451 health risk of 67% (IQR = 63% - 74%). 452 Sensitivity analysis determined the impact of input variables on model output: CHAdV, 453 fcontact, and TEsurface/hand for the inadvertent ingestion via hand-to-mouth and CHAdV and 454 Vleachate for splashing (Figure S1). In the case of the hand-to-mouth mechanism, the most 455 impacting parameter was the f_{contact} and the C_{HAdV} in the truck leachate, followed by 456 TE_{surface/hand} (the same results have been obtained for bare hands). In the case of the 457 splashing mechanism, the V_{leachate} accidentally ingested is was slightly more important 458

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465 Figure 4. Gastrointestinal (GI) illness risk (%) for waste collectors, attributable to

466 inadvertent oral ingestion of truck leachate by different mechanisms: hand to mouth

- 467 <u>contact (bare and gloved hands) and splashing route.</u>
- 468

469 **4 Discussion**

470 4.1 Microbiological and physico-chemical parameters

Commentato [IF7]: I think we forgot that subtitle for the discussion

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471	Microbiological and physico-chemical parameters physical chemical findings
472	demonstrated a significant differences based on leachate source in the values and
473	concentrations of these parameters in leachate from different sources (truck and landfill).
474	HAdV was detected in both samples, with infectious virus concentrations greater than 2.0
475	$\times 10^3$ GC ml ⁻¹ in truck leachate.
476	The detected HAdV landfill concentrations detected were similar to those reported by a
477	monitoring study described on the air and surface of an <mark>in an</mark> Italian landfill . Concerning
478	HAdV concentrations in the landfill, our results were similar to a study carried out in Italy
479	on the air and surface of a landfill-(Carducci, Federigi, and Verani 2013). To our best
480	knowledge, there is no study <u>currently reported reporting</u> virus occurrence in truck
481	leachate.
482	The PMA, an intercalant dye that binds covalently and irreversibly to viral DNA, blocks
483	the amplification of nucleic acids due to the release of the polymerase enzyme, and only
484	the target genes of intact viruses were amplified by PMA-qPCR (Leifels et al., 2021). The
485	correlation observed by different viral quantification methodologies (qPCR, PMA-
486	qPCR, and cell culture assay) demonstrated that qPCR can be predictive of infectious
487	HAdV in fresh leachate (truck), while PMA-qPCR in landfill samples. Differences on
488	HAdV levels according to detection methods were similar to those obtained by (Leifels
489	et al. 2016) on In a similar study previously conducted in sewage samples, who found
490	demonstrated one log reduction in HAdV PMA-qPCR concentration and 2 logs in TCID ₅₀
491	when compared to results obtained by qPCR, and, The correlation observed by different
492	viral quantification methodologies (qPCR, PMA-qPCR, and cell culture assay)
493	demonstrated that qPCR can be predictive of infectious HAdV in fresh leachate (truck),
494	while PMA-qPCR in landfill samples. Previously, a study HAdV conducted in sewage
495	samples demonstrated one log reduction in PMA-qPCR concentration and 2 logs in
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497	rate of HAdV from 96% to 77%, similar to results obtained in cell culture assays (Leifels	Commentato [IF8]: Could this meaning is included in the above
498	et al. 2016).	
499	Fresh truck leachate is a liquid generated directly in the waste collection truck basin	
500	during the collection day and formed from by the initial decomposition of waste,	ha formattato: Evidenziato
501	rainwater, and the compaction process carried out in the trucks. This leachate presentered	
502	high concentrations of physical-chemical compounds as suspended solids and organic	
503	compounds (Benyoucef et al. 2015). Although it wasis not possible to calculate the	
504	collection time, it <u>wasis</u> estimated that <u>sampling they</u> do not exceed 24 hours from waste	
505	production, which would justify the high microbiological values detected in the fresh	ha formattato: Evidenziato
506	truck leachate, as well as the HAdV infectivity. Moreover, Tthe lower concentration of	ha formattato: Evidenziato
507	bacteria and infectious HAdV in the landfill leachate samples compared to that obtained	
508	from trucks can be influenced by physico-chemical parameters. In the landfill, such	ha formattato: Evidenziato
509	parameters can These vary depending on the landfill lifetime and its operating methods,	
510	the interference of events before sample collection, climatic conditions, waste	
511	composition, and solids waste degradation rate. The reduction of infective HAdV, as well	
512	as the lower concentration of bacteria in landfill samples, can be attributable to physical-	
513	chemical factors that depend on the landfill lifetime and operating methods, events	
514	occurrence before sample collection, climatic conditions, waste composition, and solid	
515	waste degradation rate (Bulc 2006; Tchobanoglous and Kreith 2002). The physico-	
516	chemical parameters in landfill leachate samples physical chemical findings (i.e., as pH,	
517	total alkalinity, total inorganic carbon, ammonia, and total nitrogen) in landfill leachate	
518	samples are were similar to those already described in other studies (Costa, Alfaia, and	ha formattato: Evidenziato
519	Campos 2019; Costa et al. 2019).	

TCID₅₀ when compared to results obtained by qPCR. PMA-qPCR decreased the detection

496

nmentato [IF8]: Could this part be removed, since the ning is included in the above sentence?

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520	Multivariate regression analysis, analyzing considering leachate from both sources, and	<
521	correlating physical chemical factors with detection of HAdV revealed that the	
522	concentration of suspended solids is was a predictor of the HAdV occurrence of these	_
523	viruses detected, regardless of the method used. Such result can be attributable to the	
524	Adsorption_adsorption_of viral particles to suspended solids. thatprolongs viral	
525	infectivity in the environmentleachate, protecting from enzyme degradation and UV	
526	inactivation (Fong and Lipp 2005). This Such result finding suggests can be used for	
527	further investigations on the adoption of physico-chemical parameters physical chemical	
528	findings as a proxy for viral contamination, since viral measurements are time-consuming	
529	and expensive, and cannot be used for real-time monitoring in the perspective of timely	
530	risk management (Hess, Niessner, and Seidel 2021).	
531	Regarding HAdV diversity, our results corroborated the role of disposable diapers, toilet	
532	paper, septic tank sludge, and secretions in MSW microbial contamination (Gerba et al.	
533	2011), since the excretion of all types of HAdV in feces is well known (Bonot et al. 2014).	
534	HAdV characteristics as DNA capsid virus, of high stability, resistance to adverse	
535	environmental conditions (Rames et al. 2016) contribute to the occurrence in the leachate.	
536	The high concentration of total coliforms and E. coli from waste collection truck leachate	
537	found in this study was similar to an <mark>other Brazilian</mark> investigation conducted in Brazil that	
538	demonstrated no differences between bacteriological contamination from household and	
539	health services trucks (Silva et al. 2011).	
540		
541	4.2 Risk analysis for truck worker	

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4.2 Risk analysis for truck worker 541

- Two-The modelled exposures scenarios were modeled, assessing represented the worst-542
- case scenarios of workers involved in daily waste-related activities exposed toin the 543
- context of heavily polluted truck surfaces environment. Although such assumptions lead 544

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to an overestimation of the risk, an exposure assessment study carried out with Danish
waste collectors showed high bacteria and fungi levels on truck surfaces during the
collection of household waste by compactor trucks (Madsen et al. 2020). Moreover, We
we made the conservative assumption that all infective particles were enteropathogenic,
since a-the higher prevalence of HAdV-F, a virus associated with GI illness (Reis et al.
2016; Arcangeletti et al. 2014) was detected, although Group B and C serotypes
responsible for respiratory symptoms were also detected.

The lack of surveillance data on GI illness among Brazilian waste collectors <u>does_did_not</u> allow any inference between model outputs and epidemiological data. Nevertheless, previous studies revealed an increased incidence of GI symptoms among waste collectors in other countries (Poulsen et al. 1995; Thorn, Beijer, and Rylander 1998; Ivens et al. 1999).

The QMRA results showed the protective role of gloves for GI illness attributable to the 557 hand-to-mouth contact. In a real-world scenario, the hands could be also contaminated 558 from accidental self-inoculation during glove removal or penetration of the microbes 559 through the glove material, therefore hand hygiene procedures when removing gloves 560 should be performed. Considering QMRA results from both mechanisms, the most 561 562 relevant oral exposure mechanism was represented by the splashing route, so wearing 563 PPE could reduce these GI illness risk. Moreover, the exposure mechanism through gloved hand was less relevant than splashing and such result highlighted the importance 564 of wearing PPE (i.e., face shields) for reducing GI illness risk. In the present study, we 565 566 focused on the exposure through accidental leachate ingestion, because oral pathway is the main route for enteric disease. However, the inhalation of HAdV-laden aerosol could 567 be an additional exposure pathway, since aerosol particles with an aerodynamic diameter 568 larger than 10 µm remain entrapped on the surface of the respiratory system and can be 569

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570	swallowed. Some occupational QMRA studies quantified such ingestion rate between	
571	10% and 50% of the inhaled microorganisms (Brooks et al., 2005; Medema et al., 2004;	
572	Akpeimeh et al. 2020). The additional GI risk attributable to inhalation needs direct	
573	measurements of HAdV in air sample, but such investigations are beyond the scope of	
574	<u>this study.</u>	
575	Monitoring of HAdV load on workers' gloves is important to avoid assumptions on truck	
576	surface contamination and pathogens transfer efficiency from surface to hand, but it needs	
577	viral elution process from gloves which can reduce HAdV detection sensibility.	
578		
579	4.2.1 Limitation of the risk assessment analysis	
580	To investigate the inadvertent ingestion exposure by hands, we started from HAdV load	
581	measured in truck leachate and used input parameters derived from literature, such as the	
582	transfer efficiency from surface to hand and hand to mouth. The estimated health risk	
583	could be affected by uncertainties in the input parameters, especially on the HAdV	
584	concentration, that has been modelled as uniform distribution owing to the little sample	
585	size of monitoring data. Regarding inadvertent ingestion exposure by hands, transfer	
586	efficiency from surface-to-hand and hand-to-mouth were derived from literature, Transfer	
587	efficiency of biological agents is currently little investigated, and studies are focused on	
588	bare hands as reviewed by (Gorman Ng et al. 2012). The role of gloves is considered only	
589	in laboratory studies on food production, owing to their importance in the transmission	
590	of viral food-borne gastroenteritis (Rönnqvist et al. 2014; Sharps, Kotwal, and Cannon	
591	2012; Stals et al. 2013). Therefore, we used transfer efficiency from surface to gloved	
592	hand derived from such experimental studies, but the latex gloves of food-handling	
593	personnel could differ from waste collectors' gloves, thus determining a difference in	
594	viral acquisition from surfaces and/or in releasing to the oral compartment.	

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596 **5** Conclusions and future perspective

MSW leachate is a source of HAdV, with a higher prevalence of group F HAdV, and 597 potential risk of GI illnesses for waste collector's workers. HAdV detection and 598 quantification were higher in fresh truck leachate than in landfill samples, with 100% of 599 infectivity. In the exposure scenario of inadvertent ingestion of truck leachate, bare hand-600 to-mouth exposure was is-responsible for the highest GI risk, followed by splashing 601 exposure and gloved hand-to-mouth. This reinforces the need for risk analysis studies and 602 603 the adoption of PPE by waste collectors, especially on the evaluation of the protective role of wearing masks, whose adoption is progressively increasing as a result of changes 604 605 in infective risk perception during COVID-19 pandemic. Since some HAdV serotypes can be responsible for conjunctivitis (Gopalkrishna, Ganorkar, and Patil 2016), the role 606 of truck leachate in increasing the conjunctivitis risk should be further explored, as data 607 on ocular exposure are missing. 608

609

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Urban Cleaning Company of the city of Rio de Janeiro (Comlurb) as well as the Head of
the Waste Treatment Center of the city of Rio de Janeiro (CTR-Rio).

616

617 Authorship contribution statement

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618	Conceptualization: JCM, CFM, MPM; Formal analysis: NML, IF, MV, AC, MPM;
619	Funding acquisition: EMS, CFM, MPM; Field work: NML, RFM, CFM; Methodology:
620	NML, RFM, MDNB, IF, MV, AC, MPM; Software: LC, IF, NML; Supervision: AC,
621	MPM, CFM, JCM; Roles/Writing - original draft: NML, IF, MV, AC, MPM; Writing -

622 review & editing: NML, IF, LC, MV, CFM, AC, MPM.

623

624 Funding

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- 636 by any of the authors.

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